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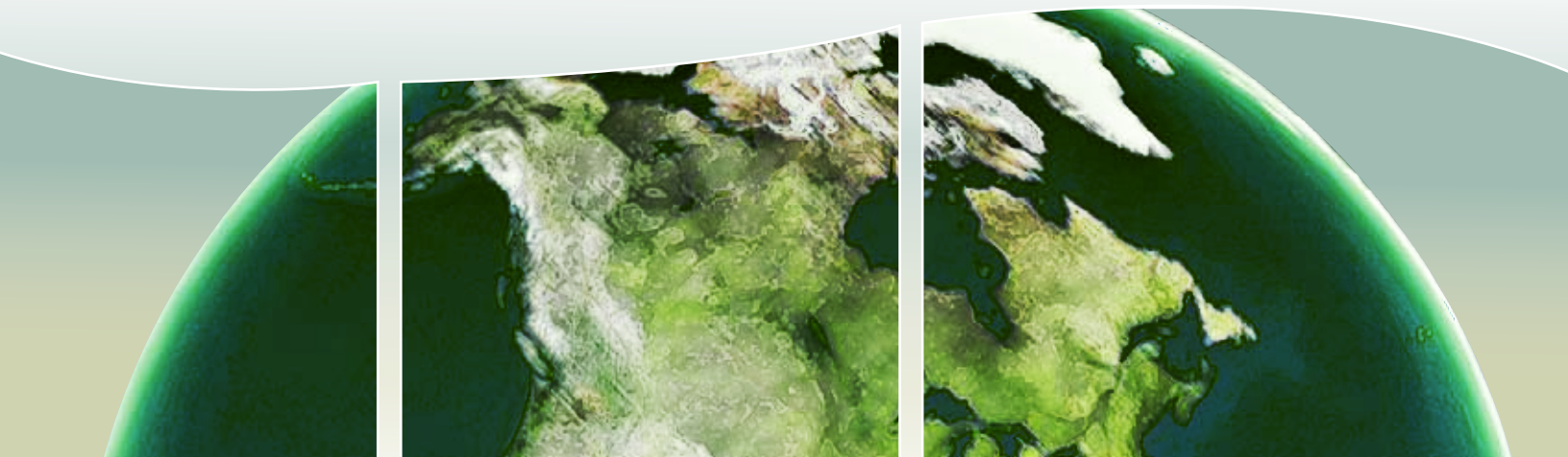
# National Inventory Report

1990-2009

GREENHOUSE GAS SOURCES  
AND SINKS IN CANADA

The Canadian Government's Submission  
to the UN Framework Convention on Climate Change

Part 1



Canada 

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# Foreword

Canada ratified the United Nations Framework Convention on Climate Change (UNFCCC) on December 4, 1992, and the Kyoto Protocol to the UNFCCC on December 17, 2002. Under Decisions 3/CP.1, 9/CP.2 and 3/CP.5 of the UNFCCC, national inventories for UNFCCC Annex I Parties should be submitted to the UNFCCC Secretariat each year, by April 15. In addition, in accordance with Decision 15/CMP.1, Annex I Parties that wish to participate fully in the Kyoto Protocol mechanisms under Articles 6, 12 and 17 should begin reporting their annual inventory and assigned amount information, when applicable in the context of their annual submissions under the Kyoto Protocol, from the year following the submission of the initial report required by decision 13/CMP.1. As such, this report represents Canada's annual inventory submission under the Framework Convention and the Kyoto Protocol.

Under the Copenhagen Accord, Canada committed to reducing its greenhouse gas emissions to 17 per cent below 2005 levels by the year 2020.

The UNFCCC and Kyoto Protocol monitoring, reporting and review guidelines for national inventories incorporate the methodological Good Practice Guidance that has been developed by the Intergovernmental Panel on Climate Change. These guidelines stipulate how emission estimates are to be prepared and what is to be included in the annual inventory report. By including additional information, the national inventory report (NIR) serves as a much better tool from which to generate indicators to compare Parties' performance under the UNFCCC and the Kyoto

Protocol. These guidelines also commit Parties to improve the quality of national and regional emissions and removals estimates on an ongoing basis. Priority areas for improvement include both the quality of input data and the methodologies utilized to develop emission and removal estimates. A number of areas have undergone improvements over the last few years as we improve the quality of the inventory. These improvements are described within the report.

Environment Canada, in consultation with a range of stakeholders, is responsible for preparing Canada's official national inventory. This National Inventory Report, prepared by staff of the Pollutant Inventories and Reporting Division of Environment Canada, complies with the UNFCCC reporting guidelines on annual inventories. It represents the efforts of many years of work and builds upon the results of previous reports, published in 1992, 1994, and yearly from 1996 to 2008, and incorporates a number of changes, including those most recently recommended by a UN Expert Review Team that reviewed Canada's initial report under the Kyoto Protocol in the fall of 2007. In addition to the inventory data, the inventory report contains analysis of recent trends in emissions and removals. In this submission, relevant information on Canada's National System, and National Registry, along with information required by articles 3.3 and 3.4 of the Kyoto Protocol, is now contained in a newly created Annex 11: Supplementary Information Required under Article 7.1 of the Kyoto Protocol.

Since the publication of the 1990 emissions inventory, an ever-increasing number of people have become interested in climate change and, more specifically, greenhouse gas emissions. While this interest has sparked a number of research activities, only a limited number have focused on measuring emissions and developing better emission estimates. There will always be uncertainties associated with emission inventories; however, ongoing work, both in Canada and elsewhere, will continue to improve the estimates and reduce uncertainties associated with them.

# List of Acronyms, Abbreviations and Units

AAC	Aluminum Association of Canada
AAFC	Agriculture and Agri-Food Canada
AC	air conditioning
AEUB	Alberta Energy and Utilities Board
AGEM	Aviation Greenhouse Gas Emission Model
Al	aluminium
Al <sub>2</sub> O <sub>3</sub>	alumina
API	American Petroleum Institute
ASH	manure ash content
ATV	all-terrain vehicle
AWMS	animal waste management system
BADA	Base of Aircraft Data
B <sub>0</sub>	maximum methane production potential
B <sub>100</sub>	100% biodiesel
BOD	biochemical oxygen demand
BOF	basic oxygen furnace
BOD <sub>5</sub>	five-day biochemical oxygen demand
C	carbon
CAC	Criteria Air Contaminant
CaC <sub>2</sub>	calcium carbide
CaCO <sub>3</sub>	calcium carbonate; limestone
CaMg(CO <sub>3</sub> ) <sub>2</sub>	dolomite (also CaCO <sub>3</sub> ·MgCO <sub>3</sub> )
CanFI	Canada's National Forest Inventory
CANSIM	Statistics Canada's key socioeconomic database
CanSIS	Canadian Soil Information System
CanWEA	Canadian Wind Energy Association
CaO	lime; quicklime; calcined limestone
CAPP	Canadian Association of Petroleum Producers
CBM	Carbon Budget Model
CBM-CFS3	Carbon Budget Model for the Canadian Forest Sector, version 3
CCFM	Canadian Council of Forest Ministers
CEA	Canadian Electricity Association
CEPA 1999	<i>Canadian Environmental Protection Act, 1999</i>
CF <sub>4</sub>	carbon tetrafluoride
C <sub>2</sub> F <sub>6</sub>	carbon hexafluoride
CFC	chlorofluorocarbon
CFS	Canadian Forest Service
CGA	Canadian Gas Association
CH <sub>3</sub> OH	methanol
CH <sub>4</sub>	methane
C <sub>2</sub> H <sub>6</sub>	ethane
C <sub>3</sub> H <sub>8</sub>	propane
C <sub>4</sub> H <sub>10</sub>	butane
C <sub>2</sub> H <sub>4</sub>	Ethylene
C <sub>6</sub> H <sub>6</sub>	Benzene
CIEEDAC	Canadian Industrial Energy End-Use Data Analysis Centre
CKD	cement kiln dust
CO	carbon monoxide

CO <sub>2</sub>	carbon dioxide
CO <sub>2</sub> eq	carbon dioxide equivalent
COD	chemical oxygen demand
CORINAIR	The Core Inventory of Air Emissions in Europe
CPPI	Canadian Petroleum Products Institute
CRF	Common Reporting Format
CSPA	Canadian Steel Producers Association
CT	conventional tillage
CTS	crop and tillage system
CVS	Canadian Vehicle Survey
DE	digestible energy
DM	dry matter
DMI	dry matter intake
DOC	degradable organic carbon
DOCF	degradable organic carbon dissimilated
DOM	dead organic matter
EAF	electric arc furnace
EC	Environment Canada
EDC	ethylene dichloride
EF	emission factor
EF <sub>BASE</sub>	basic emission factor
EMEP	European Monitoring and Evaluation Programme
EPA	Environmental Protection Agency (United States)
EPGTD	Electric Power Generation, Transmission and Distribution
eq	equivalent
ERCB	Energy Resources Conservation Board
ERT	Expert Review Team
EU	European Union
FAA	Federal Aviation Administration (United States)
FAACS	Feasibility Assessment of Afforestation for Carbon Sequestration
FCR	fuel consumption ratio
FGD	flue gas desulphurization
FLCL	forest land converted to cropland
FLSL	forest land converted to settlement
FLWL	forest land converted to wetland
FOI	Swedish Defence Research Agency
FTA	fraction of BOD in sludge that degrades anaerobically
FTILL	tillage ratio factor
g	gram
GCD	great-circle distance
GCV	gross calorific value
GDP	gross domestic product
GE	gross energy
Gg	gigagram
GHG	greenhouse gas
GHGRP	Greenhouse gas reporting program
GHV	gross heating value
GIS	geographic information system
GL	gigalitre
Gt	gigatonne
GTIS	Global Trade Information Services



GVWR	gross vehicle weight rating
GWP	global warming potential
H <sub>2</sub>	hydrogen
H <sub>2</sub> O	water
ha	hectare
HCFC	hydrochlorofluorocarbon
HCl	hydrochloric acid
HDD	heating degree-day
HDDT	heavy-duty diesel truck
HDDV	heavy-duty diesel vehicle
HDGV	heavy-duty gasoline vehicle
HE	harvest emissions
HFC	hydrofluorocarbon
HHV	higher heating value
HM	heavy metal
HNO <sub>3</sub>	nitric acid
HRAI	Heating, Refrigeration and Air Conditioning Institute of Canada
HSS	horizontal stud Söderberg
HWP	harvested wood product
HWP-C	carbon stored in harvested wood products
IAI	International Aluminium Institute
ICAO	International Civil Aviation Organization
IE	included elsewhere
IEA	International Energy Agency
I/M	inspection and maintenance
IPCC	Intergovernmental Panel on Climate Change
IT	intensive tillage
ITL	International Transaction Log
KAR	kilometre accumulation rate
k	methane generation rate constant
K <sub>2</sub> CO <sub>3</sub>	potassium carbonate
kg	kilogram
kha	kilohectare
kPa	kilopascal
kt	kilotonne
kWh	kilowatt-hour
L	litre
L <sub>0</sub>	methane generation potential
lb.	pound
LDDT	light-duty diesel truck
LDDV	light-duty diesel vehicle
LDGT	light-duty gasoline truck
LDGV	light-duty gasoline vehicle
LFG	landfill gas
LHV	lower heating value
LMC	land management change
LPG	liquefied petroleum gas
LTO	landing and takeoff
LULUCF	Land Use, Land-use Change and Forestry
m	metre
m <sup>3</sup>	cubic metre



MAI	mean annual increment
MARS	Monitoring, Accounting and Reporting System
MC	motorcycle
MCED	Manufacturing, Construction and Energy Division of Statistics Canada
MCF	methane conversion factor (Agriculture)
MCF	methane correction factor (Waste)
Mg	magnesium; also megagram
MgCO <sub>3</sub>	magnesite; magnesium carbonate
MGEM	Mobile Greenhouse Gas Emission Model
MGEM07	Mobile Greenhouse Gas Emission Model 2007
MgO	magnesia; dolomitic lime
Mha	megahectare, equivalent to a million hectares
ML	megalitre
mol	mole
MMIC	Motorcycle & Moped Industry Council
MODTF	Modeling and Database Task Force
mol	mole
MS	manure system distribution factor
MSW	municipal solid waste
Mt	megatonne
MTOW	maximum takeoff weight
mV	millivolt
MW	megawatt
N	nitrogen
N <sub>2</sub>	nitrogen gas
Na <sub>2</sub> CO <sub>3</sub>	sodium carbonate; soda ash
Na <sub>3</sub> AlF <sub>6</sub>	cryolite
NA	not applicable
N/A	not available
NAICS	North American Industry Classification System
NCV	net calorific value
NE	not estimated
NEB	National Energy Board
NGL	natural gas liquid
NH <sub>3</sub>	ammonia
NH <sub>4</sub> <sup>+</sup>	ammonium
NH <sub>4</sub> NO <sub>3</sub>	ammonium nitrate
NHTSA	National Highway Traffic Safety Administration (United States)
NIR	National Inventory Report
NMVOC	non-methane volatile organic compound
N <sub>2</sub> O	nitrous oxide
NO	nitric oxide; also used for not occurring
NO <sub>2</sub>	nitrogen dioxide
NO <sub>3</sub>	nitrate
NO <sub>x</sub>	nitrogen oxides
NOC	Nitrous Oxide of Canada
NPRI	National Pollutant Release Inventory
NRCan	Natural Resources Canada
NSCR	non-selective catalytic reduction
NT	no tillage
O <sub>2</sub>	oxygen

ODS	ozone-depleting substance
OECD	Organisation for Economic Co-operation and Development
OEM	original equipment manufacturer
OS/HOU	oil sands and heavy oil upgrading
PFC	perfluorocarbon
PJ	petajoule
PKT	passenger kilometres travelled
POP	persistent organic pollutant
ppb	part per billion
ppbv	part per billion by volume
P/PE	precipitation/potential evapotranspiration
ppm	part per million
QA	quality assurance
QC	quality control
RA	reference approach
RES-D	Report on Energy Supply and Demand in Canada
RPP	refined petroleum product
RT	reduced tillage
SA	sectoral approach
SAGE	System for assessing Aviation's Global Emissions
SAN	styrene-acrylonitrile resin
SBR	styrene-butadiene
SCR	selective catalytic reduction
SF <sub>6</sub>	sulphur hexafluoride
SIC	Standard Industrial Classification
SiC	silicon carbide
SLC	Soil Landscapes of Canada
SMR	steam methane reforming
SO <sub>2</sub>	sulphur dioxide
SO <sub>x</sub>	sulphur oxides
SOC	soil organic carbon
STC	Statistics Canada
SUV	sport utility vehicle
t	tonne
TC	Transport Canada
t-km	tonne-kilometre
TKT	tonne-kilometres travelled
TJ	terajoule
TWh	terrawatt-hour
UN	United Nations
UNFCCC	United Nations Framework Convention on Climate Change
UPCIS	Use Patterns and Controls Implementation Section
UOG	upstream oil and gas
VCM	vinyl chloride monomer
VKT	vehicle kilometres travelled
VSS	vertical stud Söderberg
VOC	volatile organic compound
VS	volatile solids
WMO	World Meteorological Organization
wt	weight

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# Executive Summary

## ES.1 Greenhouse Gas Inventories and Climate Change

Warming of the climate system is unequivocal. Most of the observed increase in global average temperatures since the mid-20th century is very likely due to the observed increase in anthropogenic greenhouse gas (GHG) concentrations (IPCC 2007). The contribution of human activities to enhancing the greenhouse effect has been recognized worldwide by both the scientific and policy communities.

The ultimate objective of the United Nations Framework Convention on Climate Change (UNFCCC) is to achieve stabilization of GHG concentrations in the atmosphere at a level that would prevent dangerous anthropogenic interference with the climate system. In support of this goal, articles 4 and 12 and Decision 3/CP.5 of the Convention commit all Parties to develop, periodically update, publish and make available to the Conference of the Parties national inventories of anthropogenic emissions by sources and removals by sinks of all GHGs not controlled by the Montreal Protocol. Development and maintenance of a National Inventory Submission is a key obligation of UNFCCC and Kyoto Protocol signatories.

Canada's National Inventory Submission is the annual communication through which Canada meets its annual reporting obligations under the UNFCCC, and demonstrates compliance with monitoring and reporting requirements under the Kyoto Protocol. The National Inventory Submission further constitutes the authoritative indicator and basis of comparison of national performance; is a source of high-quality, detailed information for Canadians on key emission trends for specific sources, sectors and regions; and provides a core data source for setting base-line emissions and scenarios development.

Canada's 2011 National Inventory Submission to the UNFCCC has been prepared in accordance with the UNFCCC Guidelines on annual inventories,

Decision 18/CP.8, and other relevant decisions. As Canada is a Party to the Kyoto Protocol, details on the national registry, changes to the national system, and information on Kyoto units are also provided in accordance with Decision 15 of the Conference of the Parties serving as the meeting of the Parties to the Kyoto Protocol (CMP1).

Under the Copenhagen Accord, Canada has committed to reducing its GHG emissions to a level 17% below the 2005 level by the year 2020. Canada is committed to tackling climate change through sustained action to build a low-carbon economy that includes reaching a global agreement, working with our North American partners and taking action domestically.

## ES.2 Summary of 2009 GHG Emissions and Removals

In 2009, Canada's total greenhouse gas emissions were estimated to be 690 megatonnes of CO<sub>2</sub> equivalent (Mt CO<sub>2</sub> eq)<sup>1</sup> a decrease of approximately 6% (42 Mt) from the updated 2008 level of 732 Mt. This was the second year in a row that emissions decreased, caused in part by the global recession and reduced use of coal for electricity generation.

Since 2005, total Canadian GHG emissions have decreased by 41 Mt. This drop is primarily the result of declines in Electricity and Heat Generation (due to decreasing use of fossil fuel) and manufacturing (due to declining output, which affected both Industrial Process and Energy emissions). There were significant fluctuations between 2005 and 2009 - although in 2008, GHG emissions stood at levels almost identical to 2005, the sharp decline in 2009 led to significant reductions in the period.

The primary GHG emitted from anthropogenic activities in 2009 was CO<sub>2</sub>, which contributed 79% of Canada's total emissions (Figure S-1). The majority of these emissions result from the combustion of fossil fuels. Methane (CH<sub>4</sub>) accounted for 13% of Canada's total emissions, resulting from activities in the Agriculture and Waste sectors, as well as fugitive emissions from oil and natural gas systems. Nitrous oxide (N<sub>2</sub>O) emissions from activities such as agriculture soil management and transportation accounted for 7% of the emissions. Perfluorocarbons (PFCs), sulphur hexafluoride (SF<sub>6</sub>) and hydrofluorocarbons

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<sup>1</sup> Unless explicitly stated otherwise, all emission estimates given in Mt represent emissions of GHGs in Mt CO<sub>2</sub> eq.

Figure S-1 Canada's Total Emissions Breakdown by Gas

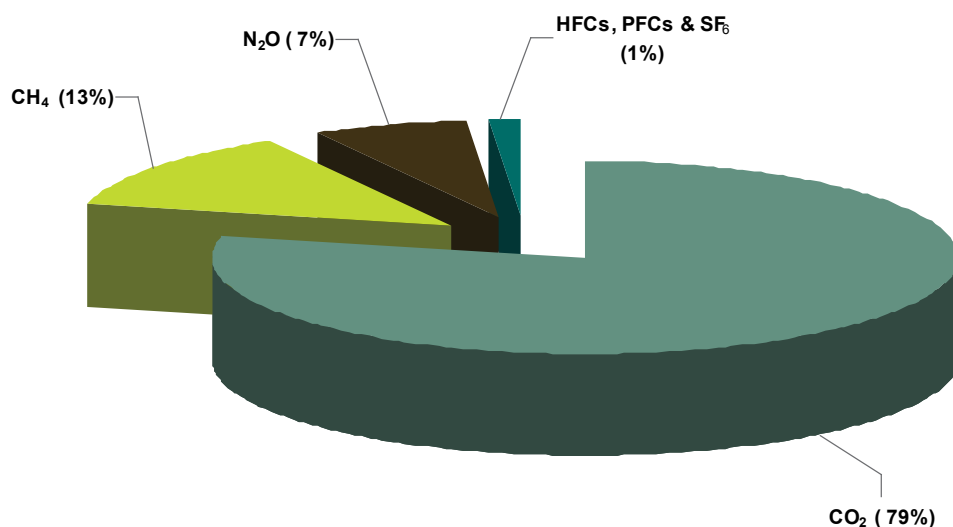
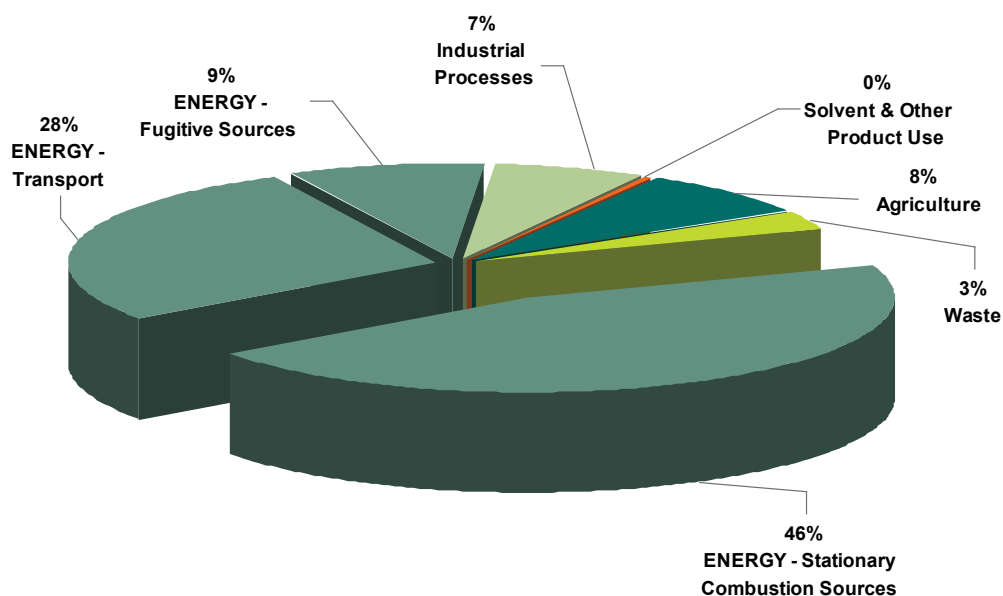


Figure S-2 Canada's Emissions Breakdown by Sector



(HFCs) constituted the remainder of the emissions (slightly more than 1%).

On a sector basis, the Energy Sector produced the majority of Canada's GHG emissions in 2009, at 82% or 566 Mt, with 73% of these emissions resulting from the combustion of fossil fuels, and the remaining 9% emitted from fugitive sources. The remaining 18% was largely generated by sources within the Agriculture (8% of total emissions) and Industrial Processes sectors (7%), with minor contributions from the Waste (3%) and Solvent and Other Product Use

sectors (Figure S-2). The Land Use, Land-use Change and Forestry (LULUCF) Sector was a net sink in 2009, removing 12 Mt from the atmosphere; however, these emissions are excluded from national inventory totals. A subset of LULUCF estimates will be accounted for at the end of the five-year first commitment period of the Kyoto Protocol (2012 inventory year). Table S-1 provides additional details about Canada's emissions and removals by sector for 1990–2009. Further sector breakdowns and a complete time series can be found in Annex 12.

Table S-1 Canada's GHG Emissions 1990–2009

Greenhouse Gas Categories	1990	1995	2000	2005	2008	2009
	<i>kt CO<sub>2</sub> equivalent</i>					
<b>TOTAL<sup>1</sup></b>	<b>590 000</b>	<b>640 000</b>	<b>716 000</b>	<b>731 000</b>	<b>732 000</b>	<b>690 000</b>
<b>ENERGY</b>	<b>468 000</b>	<b>508 000</b>	<b>586 000</b>	<b>595 000</b>	<b>597 000</b>	<b>566 000</b>
<b>a. Stationary Combustion Sources</b>	<b>279 000</b>	<b>292 000</b>	<b>343 000</b>	<b>339 000</b>	<b>339 000</b>	<b>315 000</b>
Electricity and Heat Generation	91 600	96 100	127 000	123 000	116 000	97 900
Fossil Fuel Production and Refining	51 000	54 000	67 000	66 000	69 000	64 000
Petroleum Refining and Upgrading	18 000	17 000	16 000	19 000	20 000	20 000
Fossil Fuel Production	34 000	37 000	50 000	48 000	49 000	44 000
Mining & Oil and Gas Extraction	6 650	8 520	12 400	18 600	27 600	31 300
Manufacturing Industries	56 000	55 800	56 000	48 800	43 700	42 600
Iron and Steel	5 270	6 050	6 330	5 770	4 750	4 030
Non-ferrous Metals	3 260	3 150	3 220	3 290	3 680	3 120
Chemical	8 220	10 200	10 000	7 040	7 270	7 570
Pulp and Paper	14 400	12 800	12 000	7 920	5 160	4 510
Cement	3 820	4 030	4 240	5 020	4 630	3 610
Other Manufacturing	21 000	19 500	20 200	19 800	18 200	19 700
Construction	1 870	1 170	1 070	1 360	1 260	1 080
Commercial & Institutional	25 700	28 900	33 100	36 700	35 200	36 000
Residential	43 000	45 000	45 000	42 000	43 000	41 000
Agriculture & Forestry	2 390	2 750	2 540	1 970	2 260	2 050
<b>b. Transport<sup>2</sup></b>	<b>146 000</b>	<b>160 000</b>	<b>180 000</b>	<b>193 000</b>	<b>196 000</b>	<b>190 000</b>
Civil Aviation (Domestic Aviation)	7 200	6 600	7 500	7 700	7 800	7 200
Road Transportation	96 700	107 000	118 000	130 000	132 000	131 000
Light-duty Gasoline Vehicles	45 500	43 800	41 900	40 000	39 700	41 400
Light-duty Gasoline Trucks	20 300	27 300	36 300	42 500	42 600	41 300
Heavy-duty Gasoline Vehicles	7 440	6 230	5 460	6 540	6 840	6 990
Motorcycles	152	125	161	254	264	245
Light-duty Diesel Vehicles	469	429	466	574	652	663
Light-duty Diesel Trucks	702	1 310	1 660	1 930	2 020	1 940
Heavy-duty Diesel Vehicles	20 000	26 100	30 900	37 600	39 200	38 200
Propane & Natural Gas Vehicles	2 200	2 100	1 100	720	880	780
Railways	7 000	6 000	7 000	6 000	7 000	7 000
Navigation (Domestic Marine)	5 000	4 400	5 100	6 400	5 900	5 100
Other Transportation	30 000	36 000	43 000	43 000	43 000	40 000
Off-road Gasoline	7 800	7 700	8 800	8 300	7 400	7 600
Off-road Diesel	16 000	16 000	23 000	24 000	28 000	26 000
Pipelines	6 850	11 900	11 200	10 100	7 460	6 320
<b>c. Fugitive Sources</b>	<b>42 100</b>	<b>55 600</b>	<b>63 000</b>	<b>63 100</b>	<b>62 300</b>	<b>60 700</b>
Coal Mining	2 000	2 000	900	700	800	700
Oil and Natural Gas	40 200	53 900	62 100	62 400	61 500	60 000
Oil	4 190	5 150	5 440	5 650	5 550	5 530
Natural Gas	11 400	14 900	17 700	19 200	19 700	19 400
Venting	20 200	28 800	33 500	32 100	30 700	28 700
Flaring	4 400	5 100	5 400	5 500	5 500	6 400
<b>INDUSTRIAL PROCESSES</b>	<b>56 800</b>	<b>58 900</b>	<b>53 500</b>	<b>57 200</b>	<b>54 500</b>	<b>46 300</b>
<b>a. Mineral Products</b>	<b>8 300</b>	<b>8 800</b>	<b>9 600</b>	<b>9 500</b>	<b>8 600</b>	<b>6 800</b>
Cement Production	5 400	6 100	6 700	7 200	6 600	5 100
Lime Production	1 800	1 900	1 900	1 700	1 500	1 200
Mineral Product Use <sup>3</sup>	1 090	877	1 020	589	489	449
<b>b. Chemical Industry</b>	<b>17 000</b>	<b>18 000</b>	<b>9 000</b>	<b>10 000</b>	<b>10 000</b>	<b>8 100</b>
Ammonia Production	5 000	6 500	6 800	6 300	6 700	6 200
Nitric Acid Production	1 010	1 000	1 230	1 250	1 230	1 150
Adipic Acid Production	11 000	11 000	900	2 600	2 400	660
Petrochemical Production <sup>4</sup>	110	90	97	79	73	63
<b>c. Metal Production</b>	<b>22 600</b>	<b>22 600</b>	<b>22 500</b>	<b>19 600</b>	<b>18 500</b>	<b>15 000</b>
Iron and Steel Production	10 200	11 300	11 500	10 100	10 600	7 650
Aluminum Production	9 300	9 200	8 200	8 200	7 400	7 200
SF <sub>6</sub> Used in Magnesium Smelters and Casters	3 110	2 110	2 780	1 290	462	193
<b>d. Production and Consumption of Halocarbons and SF<sub>6</sub><sup>5</sup></b>	<b>990</b>	<b>730</b>	<b>3 200</b>	<b>5 400</b>	<b>5 700</b>	<b>7 000</b>
<b>e. Other &amp; Undifferentiated Production</b>	<b>8 000</b>	<b>8 400</b>	<b>9 200</b>	<b>12 000</b>	<b>11 000</b>	<b>9 400</b>
<b>SOLVENT &amp; OTHER PRODUCT USE</b>	<b>180</b>	<b>210</b>	<b>250</b>	<b>180</b>	<b>340</b>	<b>260</b>
<b>AGRICULTURE</b>	<b>47 000</b>	<b>53 000</b>	<b>55 000</b>	<b>58 000</b>	<b>58 000</b>	<b>56 000</b>
<b>a. Enteric Fermentation</b>	<b>16 000</b>	<b>19 000</b>	<b>20 000</b>	<b>22 000</b>	<b>20 000</b>	<b>19 000</b>
<b>b. Manure Management</b>	<b>5 700</b>	<b>6 500</b>	<b>6 900</b>	<b>7 500</b>	<b>6 800</b>	<b>6 600</b>
<b>c. Agriculture Soils</b>	<b>25 000</b>	<b>27 000</b>	<b>29 000</b>	<b>28 000</b>	<b>31 000</b>	<b>30 000</b>
Direct Sources	14 000	14 000	15 000	15 000	17 000	16 000
Pasture, Range and Paddock Manure	2 200	2 800	3 100	3 400	3 200	3 000
Indirect Sources	9 000	10 000	10 000	10 000	10 000	10 000
<b>d. Field Burning of Agricultural Residues</b>	<b>210</b>	<b>170</b>	<b>120</b>	<b>41</b>	<b>45</b>	<b>45</b>
<b>WASTE</b>	<b>19 000</b>	<b>20 000</b>	<b>20 000</b>	<b>21 000</b>	<b>21 000</b>	<b>22 000</b>
<b>a. Solid Waste Disposal on Land</b>	<b>18 000</b>	<b>19 000</b>	<b>19 000</b>	<b>20 000</b>	<b>20 000</b>	<b>20 000</b>
<b>b. Wastewater Handling</b>	<b>780</b>	<b>860</b>	<b>930</b>	<b>980</b>	<b>1 000</b>	<b>1 000</b>
<b>c. Waste Incineration</b>	<b>400</b>	<b>350</b>	<b>250</b>	<b>240</b>	<b>250</b>	<b>260</b>
<b>Land Use, Land-use Change and Forestry</b>	<b>-67 000</b>	<b>190 000</b>	<b>-62 000</b>	<b>54 000</b>	<b>-17 000</b>	<b>-12 000</b>
<b>a. Forest Land</b>	<b>-93 000</b>	<b>170 000</b>	<b>-74 000</b>	<b>46 000</b>	<b>-22 000</b>	<b>-17 000</b>
<b>b. Cropland</b>	<b>11 000</b>	<b>4 700</b>	<b>-140</b>	<b>-4 300</b>	<b>-6 300</b>	<b>-6 900</b>
<b>c. Grassland</b>	<b>-</b>	<b>-</b>	<b>-</b>	<b>-</b>	<b>-</b>	<b>-</b>
<b>d. Wetlands</b>	<b>5 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>2 000</b>
<b>e. Settlements</b>	<b>9 000</b>	<b>8 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>
<b>LAND USE, LAND-USE CHANGE AND FORESTRY</b>						
<b>Activities under the Kyoto Protocol</b>						
<b>a. Article 3.3</b>						
Afforestation / reforestation	NA	NA	NA	NA	-738	-797
Deforestation	NA	NA	NA	NA	14 533	14 699
<b>b. Article 3.4</b>						
Cropland Management	3 732	NA	NA	NA	-11 711	-12 406

Notes:

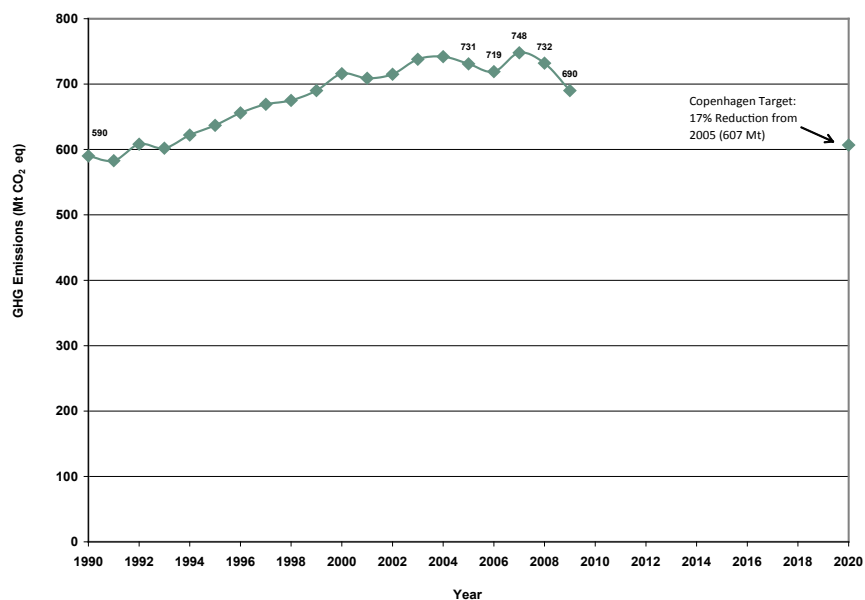
1. National totals exclude all GHGs from the Land Use, Land-use Change and Forestry Sector. The estimates for LULUCF activities under the Kyoto Protocol will be accounted for over the five years (2008–2012) of the first commitment period under the Protocol.

2. Emissions from Fuel Ethanol are reported within the gasoline transportation subcategories.

3. The category Mineral Product Use includes CO<sub>2</sub> emissions coming from the use of limestone & dolomite, soda ash and magnesite.4. The category Petrochemical Production includes emissions coming from production of silicon/calcium carbides, carbon black, ethylene, methanol, ethylene dichloride and styrene. CO<sub>2</sub> emissions are included in Other & Undifferentiated Production.

5. Production of HFCs (HCFC-22 exclusively) only occurred in Canada from 1990 to 1992. HFC consumption began in 1995.

Figure S–3 Canadian Emissions in 1990–2009\*



\* Under the Copenhagen Accord, Canada committed to reducing emissions to 17% below 2005 levels by 2020.

## ES.3 National Trends 1990–2009

Canada's emissions in 2009 were 17% (100 Mt) above the 1990 total of 590 Mt (Figure S–3). Steady increases in annual emissions characterized the first fifteen years of this period, followed by fluctuating emission levels between 2005 and 2008 and a steep decline thereafter.

Almost all of the emission changes since 1990 are attributable to six major areas: the fossil fuel (coal, oil and gas) industries,<sup>2</sup> transport,<sup>3</sup> electricity generation, manufacturing,<sup>4</sup> commercial/institutional and agriculture. The relative contribution of each of these has varied somewhat, depending on the time period (Figure S–4). The long-term (1990–2009) trend of emission growth has been driven primarily by the fossil fuel industries and transportation, whereas the short term (2005–2009) emission decline has been driven by electricity generation and manufacturing.

### Shorter-term Trends

Since 2005, total Canadian GHG emissions have decreased by 41 Mt (or 5.7%, identical to the 2008–2009 reduction). Although the gross domestic product (GDP) rose marginally between 2005 and 2009 (Table S–2), its annual rate of increase slowed in each successive year of the period. The recent economic recession began in the last quarter of 2008, causing Canada's GDP to shrink by 2.5% in 2009.

Fluctuations in emission levels since 2005 (seen in Figure S–3 and Figure S–5) are due primarily to changes in the mix of sources used for electricity production, changing emissions from fossil fuel production (due to the level of petroleum extraction activities), and varying demand for heating fuels for winters. As indicated earlier, emission reductions in this period have been driven primarily by changes occurring in the electric power generation and manufacturing areas, offset to a small extent by increases in the fossil fuel industries.

Large emission fluctuations from electricity generation have occurred recently. In some areas coal power usage has increased, while it has decreased in others. For example, in Ontario efforts have been made to reduce coal-fired generation of electricity (e.g. in 2005 Lambton generating station was shut down). At the same time, fossil fuel generation varied with the availability of electricity from hydro, nuclear and, to some extent, wind power sources.

2 "Fossil fuel industries" comprise the sum of the subsectors of Mining and Oil and Gas Extraction, Fossil Fuel Production and Refining, Pipelines (Transportation), and Fugitive Releases.

3 The "Transport" subsector refers to Transportation minus Pipelines.

4 "Manufacturing" includes the Manufacturing Industries subsector (Energy Sector) and the Industrial Processes Sector.

Figure S-4 Emission Trends for 1990–2009 and 2005–2009, Broken Down by Major Sector

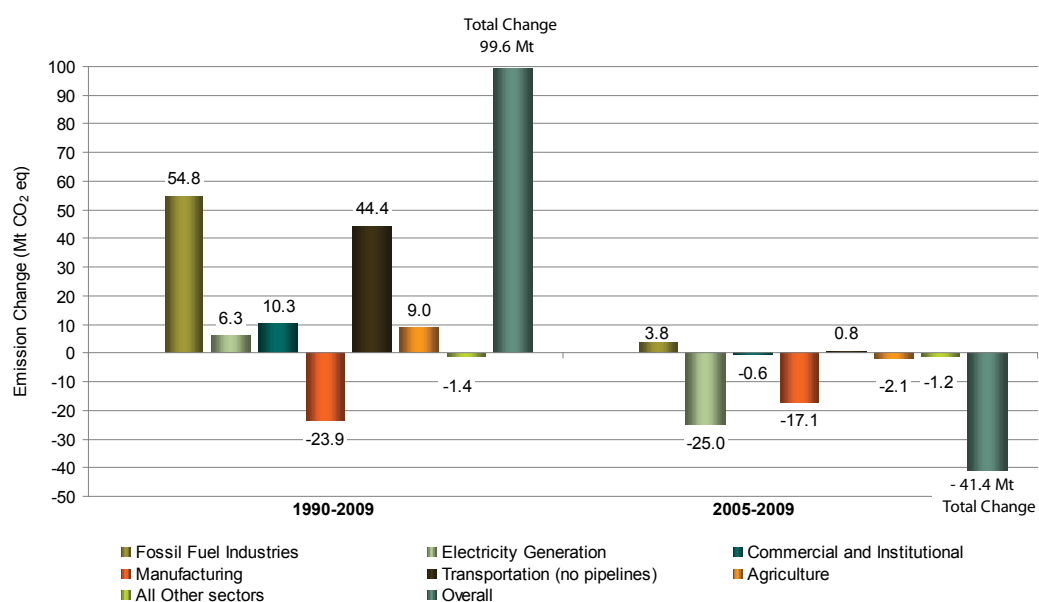
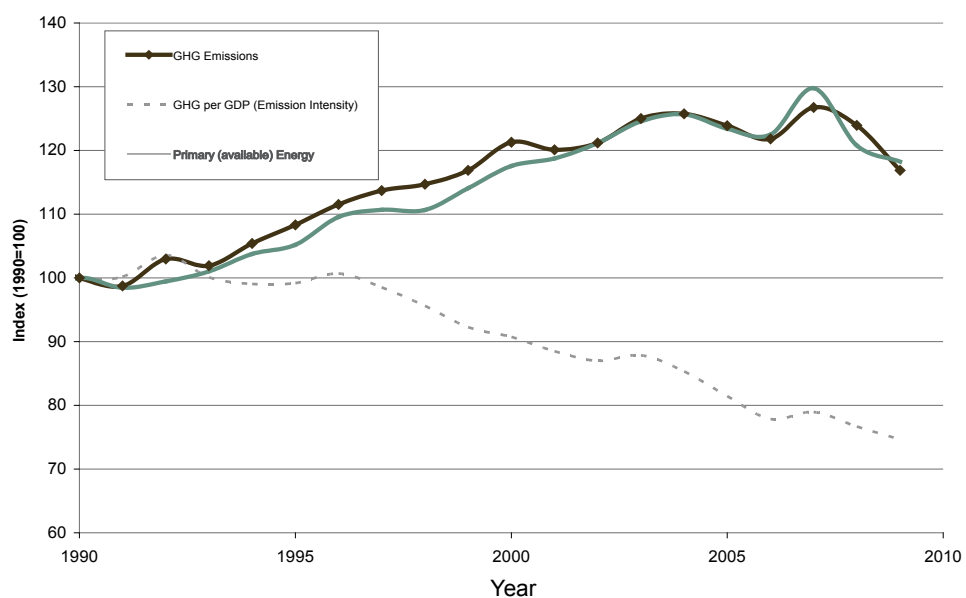


Figure S-5 Indexed Trend in GHG Emissions along with Key Metrics, 1990–2009



In fact, renewable generation has made some significant inroads in Canada. The federal government's Wind Power Production Incentive (WPPI), and ecoEnergy for Renewable Power, Ontario's new Renewable Energy Standard Offer Program (RESOP) are government incentives that have directly funded projects generating renewable power. In total, the projects funded under all of these incentives have a generation capacity of over 6000 MW, which could be offsetting as much as 3.6 Mt CO<sub>2</sub> eq of GHG emissions annually<sup>5</sup>.

Overall, emissions from electricity and heat generation dropped by 25 Mt since 2005. In 2009, Canadian electricity GHG emissions shrank by 18 Mt (16%) from 2008 levels as demand fell and coal-fired generation dropped to its lowest level since 1990.

Emissions from manufacturing decreased by 17 Mt (15%) between 2005 and 2009, due to significantly lowered production evidenced by falling manufacturing GDP, particularly in the last year of the period. In 2009 the value of exported Canadian industrial goods and machinery fell by about 30% compared to 2008 (Statistics Canada 2011).

The fossil fuel industries showed an increase of about 4 Mt (2%) in GHG emissions between 2005 and 2009. This rise was fuelled by a 40% growth in emissions from oil sands activities, offset by 12% and 1% reductions in conventional oil production and natural gas production and processing, respectively. During the period, crude oil exports increased, driven by increasing demand for bitumen and

synthetic crude in the U.S. Although net GHG growth over the 2005–2009 period was positive, emissions from this sector fell in 2009 as compared to 2008.

Between 2005 and 2009, transportation (not including pipelines) GHG emissions rose by less than 1 Mt. This is in contrast to the long-term, high-growth trend for this sector. In fact, emissions fell by about 5 Mt between 2008 and 2009. Most of this drop occurred in diesel transport. Emissions from both heavy-duty diesel on-road vehicles for shipping, and off-road vehicles (for industry) fell, primarily a result of reduced economic activity.

It is to be noted that over the past number of years Canada has developed more than 40 programs that have encouraged or targeted measures to achieve emission reductions. These include energy efficiency programs, renewable power incentives and funding programs for GHG reduction projects. These programs are likely to have already had an impact on reducing Canada's emissions<sup>6</sup>. In addition, Canada has initiated additional sector-specific reduction programs.

### Long-term Trends

Between 1990 and 2009 the fossil fuel industries were responsible for about 54% of the total 100 Mt growth, and transport was responsible for about 45%. Major increases in oil and gas production (much of it for export), as well as a large increase in the number of motor vehicles, especially

5 Environment Canada, *Analysis of Notable Climate Change Incentives in Canada*. Internal document, Pollutant Inventories and Reporting Division, Environment Canada. Mar, 2011

6 Altogether, the major federal GHG reduction initiatives already in place may, by 2009, have been contributing on the order of 25–40 Mt of annual greenhouse gas emission reductions (Ibid). There are also a number of additional provincial programs, including forms of carbon-based taxation in both British Columbia and Quebec, and Ontario has introduced the *Green Energy Act*.

**Table S–2 Trends in Emissions and Economic Indicators for Selected Years (1990–2009)**

	1990	1995	2000	2005	2006	2007	2008	2009
Total GHG (Mt)	590	637	716	731	719	748	732	690
Change Since 1990 (%)	NA	7.9	21.3	23.9	21.8	26.7	23.9	16.9
Annual Change (%)	NA	2.4	3.8	-1.5	-1.7	4.0	-2.2	-5.7
Average Annual Change (%)*	NA	1.6	2.1	1.6	1.4	1.6	1.3	0.9
GDP (Billions 2002\$)	825	899	1101	1248	1283	1311	1318	1286
Change Since 1990 (%)	NA	8.9	33.3	51.2	55.5	58.9	59.7	55.8
Annual Change (%)	NA	2.8	5.2	3.0	2.8	2.2	0.5	-2.5
GHG Intensity (Mt/\$B GDP)	0.72	0.71	0.65	0.59	0.56	0.57	0.56	0.54
Change Since 1990 (%)	NA	-0.9	-9.0	-18.1	-21.6	-20.2	-22.4	-25.0
Annual Change (%)	NA	-0.4	-1.4	-4.4	-4.4	1.8	-2.7	-3.3

\*Average annual change since 1990.

GDP: Statistics Canada - Table 384-0002 - Expenditure-based, annual, chained (billions)

Annual Change: Implies change over previous calendar year.



light-duty gasoline trucks (vans, SUVs and pick-ups) and heavy-duty diesel vehicles (commercial transport trucks) have contributed to the significant rise in GHG emissions. The growth in emissions since 1990 largely mirrors an increase in primary energy use (Figure S-5).

Emissions from the manufacturing area fell by about 24 Mt (21%), counteracting the dominant rising trend. Here fuel-switching, efficiency and technology improvements, as well as reductions in manufacturing output (especially in the Pulp and Paper and Iron and Steel subsectors) resulted in the emission reductions.

Agriculture was responsible for about a 9 Mt increase in emissions between 1990 and 2009, largely the result of increasing use of fertilizers and greater numbers of beef cattle and swine.

Also contributing to the 1990–2009 emissions growth was the Commercial/Institutional sector. The service industries are centered in the Commercial/Institutional category and have contributed a major portion of Canada's growth in economic output (Informetrica 2011). The much lower emission intensity (GHG per dollar output) of the service industries as compared to say, the fossil fuel industries, means that the Commercial/Institutional category has contributed only about 10% to the rise in domestic GHG emissions, much less than fossil fuel industries, in spite of contributing more to gross domestic product (GDP).

Though greenhouse gas emissions rose by 17% between 1990 and 2009, Canada's economy grew much more rapidly. Between 1990 and 2009, GDP rose by 56% (Table S-2). As a result, the emission intensity for the whole economy (GHG per GDP) has improved considerably, dropping by 25%. There have been some variations over time, however.

At the beginning of the period, energy prices were low (EIA 2004) and this significantly limited the economic incentives to improve energy efficiency (Figure S-5). Between 1990 and 1994, emission intensity remained stable, with emissions rising nearly in step with economic growth (which was strengthening after a recession in the early 1990s). In this time frame, emissions and GDP both rose by about 11%. Beginning in 1995, however, there was a decoupling of GDP and emissions.

The trend in the years since the late 1990s demonstrates a decline in the rate of increase of GHG emissions (even if the steep drop in 2009 is ignored). From 1990 to 2000 the average annual growth in emissions was 2.1%, while

in contrast, between 2000 and 2008, the average annual emission growth was 0.3%.

The change in the rate of growth in emissions since about 1997–2000 can be attributed to:

- Increases in efficiency, modernization of industrial processes, and structural changes in the composition of the economy, which are long-term trends that have had an increased impact on emissions since the late 1990s.
  - The structural changes have involved a move from an industrial-oriented economy to a more service-based economy. Between 2000 and 2008, the GDP of the service industries rose by 28%, while heavy industries and manufacturing together grew by only 3%. As mentioned earlier, service industries have a much lower economic GHG intensity than that of the goods-producing industries, so this ongoing change has lowered Canadian GHG emissions.
  - Together, efficiency increases and technological and structural changes have resulted in a continuing weakening of the link between GDP growth and emissions, so that the GHG intensity of the economy has decreased on average by 2.2% per year since 1996 (see Figure S-5). This has resulted in the economy growing much more rapidly than emissions.
- A leveling off of emissions from electric power generation, which had been rising rapidly until then. In 2000, coal generation was at or close to its highest level ever. Since then, the contribution of coal-fired generation to the electricity supply mix has been declining (Statistics Canada 2010a).
- The increased prevalence of energy efficiency and emission reduction programs, including federal programs such as the eco-Energy retrofit program and its predecessors and renewable energy incentives, such as the federal Wind Power Production Incentive (WPPI), which commenced in 2002 (IEA 2011).
- The peak in the production of conventional oil in 1998 in Canada and the leveling off of gas production in 2002 (Statistics Canada 2010a). In both cases, this was the result of limited conventional reserves. More recently, conventional oil and natural gas production has fallen—this has reduced fugitive emissions, to some extent offsetting the impact of oil sands growth.

## ES.4 Sector Trends for 1990–2009

### ES.4.1 Energy Sector—2009 GHG Emissions (566 Mt)

#### Short-term Changes

As with the overall trend, GHG emissions from the Energy Sector reversed and fell over the short term. In 2009, GHG emissions from the Energy Sector dropped by 29 Mt (about 5%). This decline was driven by Electricity and Heat Generation and the Manufacturing Industries.

Electricity and Heat Generation emissions shrank by 25 Mt (about 7%) from 2005 levels. Between 2005 and 2009, however, there were large emission variations that were the primary cause of the fluctuation in national emissions during this period (see Section ES.3, National Trends). Decreased electricity demand due to the economic recession contributed significantly to the decrease in emissions between 2008 and 2009.

GHG emissions from Manufacturing Industries dropped by 6 Mt (15%) between 2005 and 2009, due to significantly lowered production. In this period, the GDP for manufacturing dropped by 15%—the decline between 2008 and 2009 alone was 12.4%. Globalization, structural shifts in the economy away from manufacturing and towards the service industries (see Section ES.3) as well as the global recession in 2008–2009 all impacted on this trend.

As indicated earlier, the fossil fuel industries contradicted the overall trend, showing a 4.3 Mt (2.8%) increase in GHG emissions between 2005 and 2009.

#### Long-term Trends

By far the largest portion of Canada's total emission growth is observed in the Energy Sector. The long-term Sector emission trends (1990–2009) showed both declines and increases, but the increases were well ahead of the declines, for a net growth of 98 Mt, or 21%. As described above in Section ES.3, most of the growth in national emissions is observed in the fossil fuel industries, transportation, and to some extent electricity and commercial & institutional areas, all of which fall under the Energy Sector.

The activities of the fossil fuel industries include combustion sources (Fossil Fuel Production and Refining, Mining & Oil and Gas Extraction, and Pipelines) and

fugitive sources (Coal Mining, and Oil and Natural Gas).<sup>7</sup>

The fossil fuel industries registered a net increase of about 55 Mt of GHG emissions from 1990 to 2009 (51% growth). These emissions are related to coal mining and the production, transmission, processing, refining and distribution of all oil and gas products.

By 2009, total production of crude oil and natural gas had increased 57% over 1990 levels. Elevated demand, particularly in the United States, drove these trends, with the export market growing by far the most rapidly<sup>8</sup>. Although increasing demand provides a large portion of the explanation for the emission trend, it does not paint the complete picture.

Since well before 1990, easily removable reserves of conventional crude have been shrinking. Thus, energy consumption per unit of conventional oil produced has been increasing (Neitzert et al. 1999). At the same time, energy- and GHG-intensive production from oil sands activities has become increasingly competitive with conventional oil extraction. These trends contributed to the rapidly rising emission increases in the oil and gas industry over the 1990–2009 period.

Oil sands mining, extraction and upgrading activities were about 1.6 times more GHG-intensive than conventional oil production in 2009. However, the oil sands industry has been reducing its per-unit emissions, and in 2009 intensity was 29% lower than in 1990<sup>9</sup>. This reduction in intensity is positive, as larger and larger portions of production are derived from oil sands.

Most transportation emissions in Canada are related to Road Transport, which dominated the GHG growth trend in this area. Emissions from Road Transportation rose by 35 Mt (36%) between 1990 and 2009. Of particular interest in this subsector is a 21 Mt (104%) increase in emissions from light-duty gasoline trucks (LDGTs), which are less efficient from a fuel consumption perspective. This was partially offset by 4.1 and 1.4 Mt emission reductions from gasoline-fuelled cars (light-duty gasoline vehicles, or LDGVs) and

<sup>7</sup> There is also some overlap with Mining (which, as a result of categorizations by the Alberta Energy Utilities Board and Statistics Canada, includes a portion of oil sands production activities), but emissions from Mining are not included in this discussion of the fossil fuel industries.

<sup>8</sup> But note that a significant portion of the refined petroleum products consumed in Canada is derived from imported oil.

<sup>9</sup> This estimate is derived from an aggregate combination of sources which may change as more information and methodological refinements become available.

alternatively fuelled cars (propane and natural gas vehicles), respectively.

The primary source of this net trend of rising emissions is the increase in the number of passenger-kilometres travelled (more people drove further) (NRCan 2009). However, it was the passenger-kilometres driven by light trucks that increased, while those driven by cars decreased. Contributing to this trend was the fact that the number of light trucks on the road more than doubled between 1990 and 2007, while the number of automobiles remained virtually constant. Since light trucks have higher emissions per kilometre than automobiles, the rising popularity of sport utility vehicles and pickups worsened the emission impact of increasing passenger-kilometres. Contributing further to this is the overall trend towards increasing horsepower for all classes of passenger vehicles, which has negated the rather substantial efficiency improvements made in internal combustion engines (NRCan 2009).

Emissions from heavy-duty diesel vehicles (large freight trucks) rose by 18 Mt between 1990 and 2009, a 91% increase. Spurred on by free trade and the deregulation of the trucking industry, the amount of freight shipped grew rapidly over that period. In addition, the quantity shipped by truck (as opposed to other modes of transport, such as rail) increased as a result of customer requirements for just-in-time delivery and cross-border freight (NRCan 2009).

Electricity and Heat Generation also saw increases in emissions. Rising demand for electricity caused GHG emissions to grow by 6.3 Mt between 1990 and 2009, with significant fluctuations. In 2009, total electricity generation was approximately 115 TWh (terawatt-hours) or 23% above the 1990 level (Statistics Canada 2010b). Although there has been fluctuation in the trend over the last few years, overall intensity has improved due to lower RPP usage and increased generation from hydro, wind and natural gas sources. Summaries of provincial electricity intensities can be found in Annex 13.

Of primary importance in this trend is that, starting in the mid-1990s, the GHG emissions associated with coal-fired electricity generation progressively increased. They then decreased between 2002 and 2009. This was due to the return to service of a number of nuclear units and a commitment to reduce coal-fired electricity generation in Ontario, as well as fuel switching to natural gas in a number of regions of the country. At 1.2% of generation by 2009, non-hydro renewable energy sources have had some

impact on emission reductions. The impact is predicted to become more significant in future years because installed capacity of wind power in Canada has been rising rapidly. Regardless, fuel and generation costs are likely to continue to play a major role in determining whether coal-fired generation and the associated GHG emissions will be reduced further in the future.

The Commercial & Institutional subsector displayed a 10.3 Mt (40%) decrease in GHG emissions between 1990 and 2009. One key factor behind this trend is the increasing efficiency of this subsector, as floor space of commercial and institutional buildings (e.g. offices, schools, stores and government edifices) increased by 37% between 1990 and 2008 (the last year for which figures are available).

## ES.4.2 Industrial Processes— 2009 GHG Emissions (46 Mt)

The Industrial Processes Sector generally covers GHG emissions arising from non-energetic sources such as limestone calcination (CO<sub>2</sub>) in cement production, or the use of hydrofluorocarbons (HFCs) and perfluorocarbons (PFCs) as replacement refrigerants for ozone-depleting substances (ODSs). The Sector has shown a gradually declining profile of emissions since 1990. For 2009, a decrease of 15% from the 2008 level was observed. The decrease was primarily driven by the reductions in the manufacturing output associated with the iron and steel, aluminium, cement and chemical industries. The source categories comprising the Sector, however, have shown considerable individual variations across the time series. Figure S-6 shows the variations for the five main categories in the Sector from 1990 to 2009.

In the Metal Production category, CO<sub>2</sub> emissions from the production of iron and steel have been decreasing since the early 1990s, despite moderate increases in steel production in Canada (up to 2008; 2009 saw a large reduction in production). This reflects a reduction in emission intensities achieved by the steel industry's increased use of recycled steel (compared to pig iron production, which is a carbon-intensive process). The sharp decrease in emissions observed between 2008 and 2009 (around 30%) reflects the recessionary year's effect on output. The aluminium industry, while increasing its production by almost 100% since 1990, shows a reduction of its process emissions by 23%. This has been achieved by strict emission controls applied to the electrolytic production process and

Figure S-6 Contribution of Constituent Source Categories to Overall Industrial Processes Sector Emissions

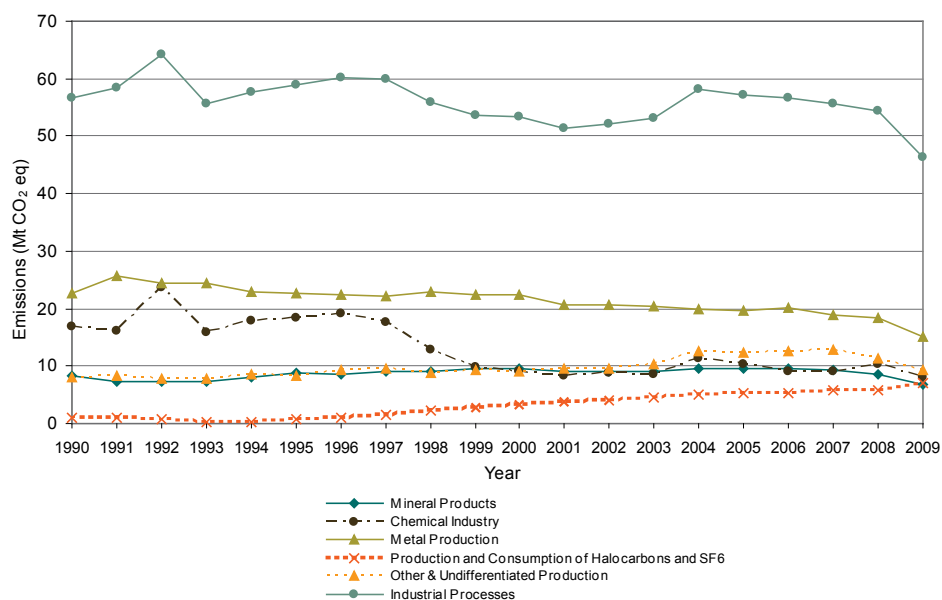
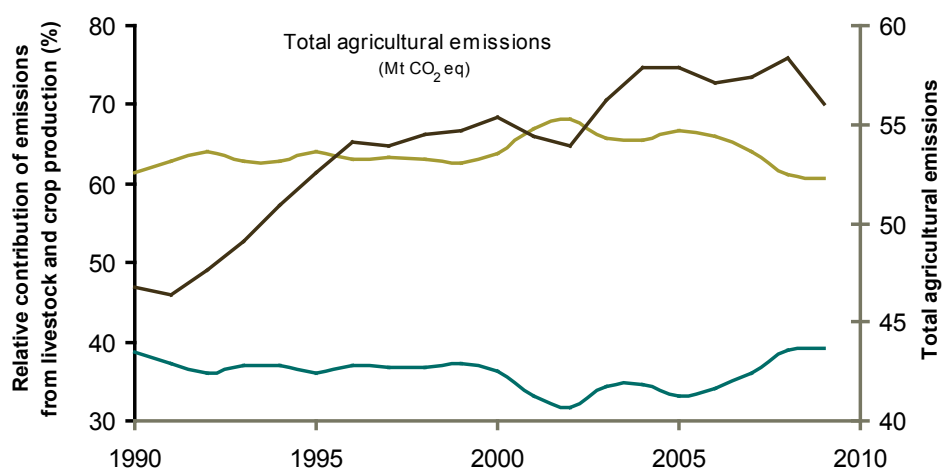


Figure S-7 Relative GHG Contribution from Livestock and Crop Production and Total Agricultural Emissions, 1990–2009



the prevention of anode effects. The reduction in overall chemical industries' GHG emissions, from process activities, of 50% between 1990 and 2009, is a combined result of the introduction of N<sub>2</sub>O controls in the now closed adipic acid plant (reductions in the 10 Mt CO<sub>2</sub> eq range), which is partly offset by increases in emissions from the Ammonia Production and Nitric Acid Production categories.

### ES.4.3 Agriculture—2009 Emissions (56 Mt)

Canadian agriculture can be differentiated into livestock and crop production components. The livestock component is dominated by the beef and swine industries, while crop production is mainly dedicated to the production of cereal and oil seeds. Dairy and poultry production are controlled to meet national demand. A wide variety of specialty crops and animals are produced, but represent a very small portion of the overall agricultural economy. The components are highly regionalized; approximately 70% of beef cattle and more than 90% of wheat, barley and canola are produced on the Prairies in a semi-arid to sub-humid ecozone. On the other hand, approximately 70% of dairy cattle, 60% of swine and poultry, 95% of corn and 90% of soybeans are produced in the humid mixed-wood plains ecozone of eastern Canada. Traditionally, Canada's Agriculture Sector has been composed of small family farms but, over the past 30 years, intensification has occurred in the Agriculture Sector and, as a consequence, the number of farms has decreased and farm size and productivity have increased.

Emissions directly related to animal and crop production accounted for 56 Mt CO<sub>2</sub> eq or 8.1% of total 2009 GHG emissions for Canada, an increase of 9 Mt CO<sub>2</sub> eq or 20% since 1990. Agriculture accounts for 24% and 72% of the national CH<sub>4</sub> and N<sub>2</sub>O emissions, respectively. All these emissions are from non-energy sources: N<sub>2</sub>O accounted for 61% of estimated sectoral emissions in 2009 and CH<sub>4</sub> for 39%.

The main drivers of the emission trend in the Agriculture Sector are the expansion of the beef cattle and swine populations, and increases in the application of synthetic nitrogen fertilizers in the Prairies. Beef, swine and poultry populations in Canada are 23%, 19% and 31% higher, respectively, than in 1990. The significant growth in animal populations largely accounts for the 19% increase in emissions, from 29 to 34 Mt CO<sub>2</sub> eq in emissions associated with

animal production over the 1990–2009 period.

Overall, through the 1990s, the combination of increased livestock populations and increasing emissions per animal in some animal categories resulted in a change in the relative proportion of GHGs coming from livestock production from 63 to 67%, increasing to a high of 69% during the drought years of 2001 and 2002 (Figure S–7). Since 2005, emissions from the Agriculture Sector have stabilized, with declines in emissions from livestock production being compensated for by increases in emissions from crop production. In 2009 a continued reduction in emissions from livestock production and a reduction in emissions from crop production, resulted in an apparent decrease in emissions, but this reduction may be insignificant in relation to inter-annual variability.

### ES.4.4 Land Use, Land-use Change and Forestry—2009 (Net Removals of 12 Mt)

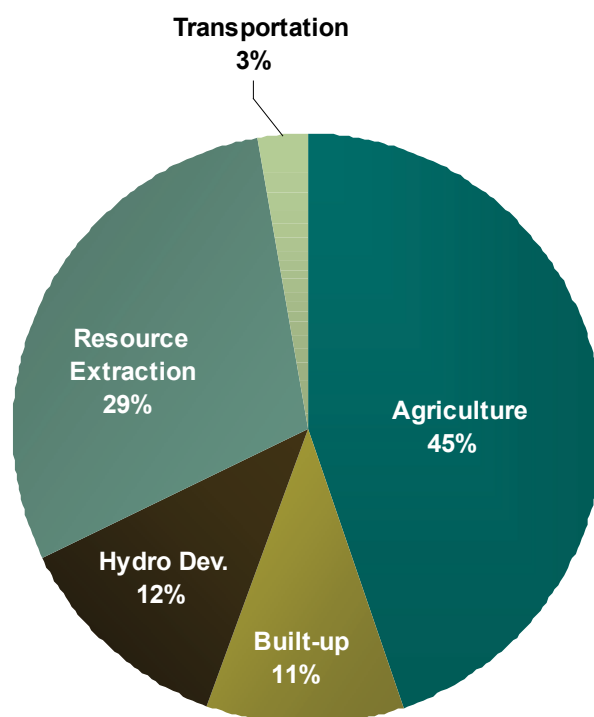
The Land Use, Land-use Change and Forestry (LULUCF) Sector reports GHG fluxes between the atmosphere and Canada's managed lands, as well as those associated with land-use change. All emissions and removals in the LULUCF Sector are excluded from the national totals.

In this sector, the net GHG flux is calculated as the sum of CO<sub>2</sub> emissions to, and removals from the atmosphere, plus non-CO<sub>2</sub> emissions. In 2009, this net flux amounted to removals of 12 Mt, which would have decreased the total Canadian GHG emissions by about 2%. Trends in the LULUCF Sector are primarily driven by those in forest land, cropland and forest conversion.

The net flux in forest land displays an important annual variability due to the erratic pattern of forest wildfires, which masks underlying patterns of interest in the Sector. Important subsectoral trends associated with human activities in managed forests include a 28% increase in the carbon removed in harvested wood biomass between 1990 and the peak harvest years of 2004–2005. Since then, significant reductions in forest management activities have occurred, with a 39% decline in harvest levels, which in 2009 reached their lowest point for the two decades covered by this report (32 Mt C). This trend reflects a deep restructuring of the Canadian forest economic sector, aggravated by the consequences of the economic recession in the United States, Canada's main



Figure S–8 Percent of Total Forest Area Conversion since 1990 by Economic Sector



Source: Hayne S and McGovern M, 2011

export market. Nonetheless, the immediate and long-term effect of major natural disturbances in managed forests, notably the Mountain Pine Beetle infestation in western Canada, will undoubtedly continue to dominate the apparent trend.

Another trend of interest is the steady decline in emissions from cropland, from 11 Mt CO<sub>2</sub> eq in 1990 to a net removal of 6.9 Mt CO<sub>2</sub> eq in 2009. This pattern largely results from changing agricultural land management practices in western Canada, such as the extensive adoption of conservative tillage practices, reduction in summerfallow and an increase in perennial forage crops. A decline in forest conversion to cropland has also contributed to this trend.

Since 1990, approximately one million hectares of forest have been lost in Canada. The conversion of forests to other land is driven by a great variety of circumstances across the country, including policy and regulatory frameworks, market forces and resource endowment. The economic drivers of forest conversion are diverse and result in heterogeneous spatial and temporal patterns of forest conversion. Primary drivers of forest conversion include agricultural expansion, resource extraction and hydroelectric development (Figure S–8). Geographically, the highest average rates of forest conversion occur in the Boreal Plain

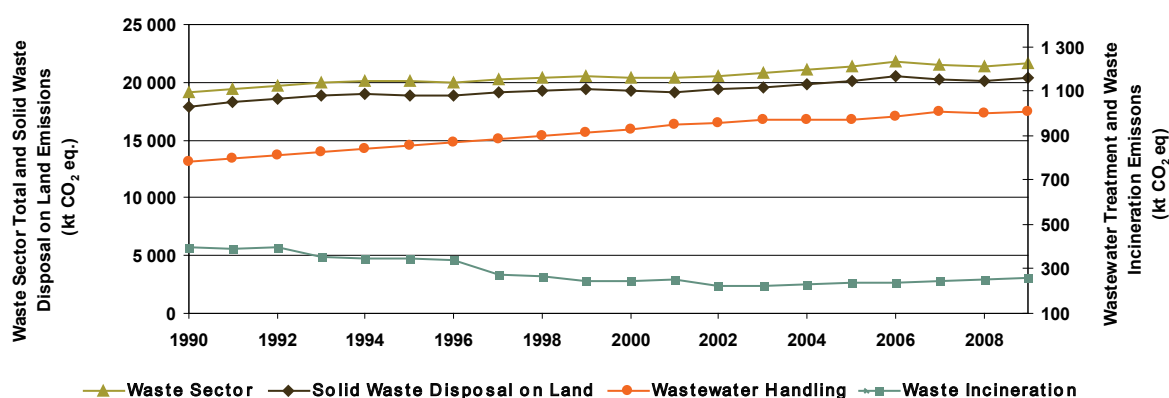
(24 kha per year) and the Boreal Shield East (9 kha per year), rates that account for 45% and 17%, respectively, of the total forest area lost in Canada since 1990. GHG emissions from forest conversion have declined from 26 Mt CO<sub>2</sub> eq in 1990 to 18 Mt CO<sub>2</sub> eq in 2009, reflecting the decline in annual rates of deforestation to agriculture from 42 kha in 1990 to 19 kha in 2009. In contrast, forest clearing for oil and gas extraction has steadily increased, from 5.3 kha per year in 1990 to 10.6 kha per year in 2009.

This year includes the second reporting of Land Use, Land-use Change and Forestry activities under articles 3.3 and 3.4 of the Kyoto Protocol, with emission and removal estimates for afforestation and deforestation (mandatory), and cropland management (elected by Canada) for the years 2008 and 2009. These estimates do not affect the national totals, and will only be accounted for at the end of the five-year commitment period (2012 inventory year).

### ES.4.5 Waste—2009 Emissions (22 Mt)

The primary source category in the Waste Sector is CH<sub>4</sub> Emissions from Solid Waste Disposal on Land, which contribute about 90% of the CO<sub>2</sub> eq GHG emissions in 2009 to

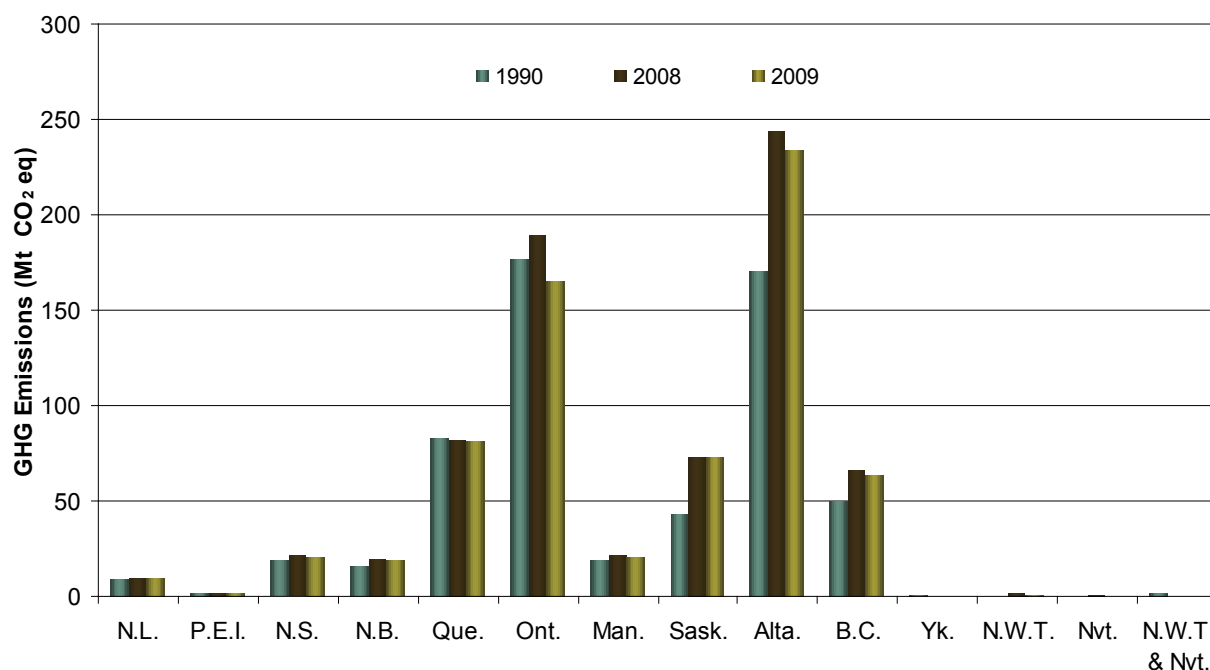
Figure S-9 GHG Emissions from Waste, 1990–2009



this sector. The CH<sub>4</sub> emissions from publicly and privately owned municipal solid waste landfills produce the bulk of emissions (about 82% of total), with a smaller part (about 8%) coming from pulp and paper and saw mill industries that landfill wood residues on-site—though this practice is declining as markets for wood residues build up. The remainder of this sector's emissions relate to CH<sub>4</sub> and N<sub>2</sub>O from wastewater treatment, and a smaller part coming from CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O from incineration facilities. Figure S-9 shows the trends in emissions from 1990 to 2009 for the Sector and the three constituent subcategories.

The overall emissions from the Sector grew by 13%, mostly from increases in emissions from landfill operations. The population growth for the same period, however, was 22%, indicating the positive effect of increasing landfill gas (LFG) capture activities at various sites across Canada. The amount of LFG captured increased by 81% to 349 kt of CH<sub>4</sub> in 2009; 51% of the LFG was utilized in energy applications and the rest was flared. The number of landfill sites with LFG capture systems is rapidly rising in Canada (about a 45% increase since 2005). Waste diversion programs, such as the composting of organics by municipal composting

Figure S-10 Emissions by Province in 1990, 2008, 2009





facilities, are also on the increase, as some provinces have put a total ban on organics in landfills. The effects of such programs are expected to be more pronounced in the near future, as seen from the drops in average per capita rates of emissions in the Solid Waste subsector, where a drop of 5% has already been observed since 2005.

Wastewater treatment and waste incineration facilities in Canada are minor sources of CH<sub>4</sub> and N<sub>2</sub>O emissions and have generally stayed stable.

## ES.5 Provincial and Territorial GHG Emissions

Figure S–10 shows where Canadian provinces stand with regard to GHG emissions for the years 1990, 2008 and 2009. While Ontario, with its large manufacturing industry base, started off as the largest-emitting province in 1990, it has been surpassed by Alberta in more recent years as Alberta has increased its production of petroleum resources for export markets. The recession year 2009 also affected the demand for Ontario's products and electricity generation, further reducing its emissions. The replacement of coal-powered generating units with natural gas units and increased electricity production from hydroelectric resources were other emission-reducing factors for Ontario. In 2009, the combined emissions from Alberta and Ontario contributed almost 58% (33.8% and 23.9%, respectively) to the national total of 690 Mt. The provinces of Quebec and British Columbia—relying on abundant hydroelectric resources for their electricity production—show more stable emission results across the time series. In the case of Quebec, some decreases have been observed since 1990. The latter profiles are more or less attributable to other Canadian provinces, except for Saskatchewan, where again the increased activities in the oil and gas industry, as well as potash and uranium mining, increased emissions by around 70% between 1990 and 2009.

## ES.6 National System and Quality Management

The *Canadian Environmental Protection Act, 1999* (Canada 1999) provides the legislative authority to designate Environment Canada's Pollutant Inventories and Reporting Division as the single national entity with responsibility for the preparation and submission of the National Inventory Submission to the UNFCCC and for the establishment of a

national system. Canada's national system covers the institutional arrangements for the preparation of the inventory, including

- the roles and responsibilities of the inventory agency and of the various players involved;
- the processes for inventory preparation, data collection and estimates development;
- quality management of the inventory; and
- the procedures for official approval of the inventory.

Submission of information to the national system, including details on institutional arrangements for inventory preparation, is also an annual requirement under the UNFCCC reporting guidelines on annual inventories (see Chapter 1, Section 1.2).

Quality assurance and quality control (QA/QC) is an integral part of the preparation of this inventory (see Annex 6). Canada's quality system includes a QA/QC plan, an archiving system, documented processes for data collection and estimate development, identification of key sources through analysis (Annex 1), quantitative uncertainty assessments (Annex 7), and a process of performing recalculations for improvement of the inventory (Chapter 9).

## ES.7 Structure of Submission

The UNFCCC requirements are the annual compilation and submission of the National Inventory Report (NIR) and Common Reporting Format (CRF) tables. The CRF tables are a series of standardized data tables, containing mainly numerical information, which are submitted electronically. The NIR contains the information to support the CRF tables, including a comprehensive description of the methodologies used in compiling the inventory, the data sources, the institutional structures and quality assurance and quality control procedures.

Part 1 of the NIR includes chapters 1 to 9. Chapter 1 (Introduction) provides an overview of Canada's legal, institutional and procedural arrangements for producing the inventory (i.e. the national inventory system) as well as a description of Canada's facility emission-reporting system. Chapter 2 provides an analysis of Canada's GHG emission trends in accordance with the UNFCCC reporting structure as well as an assessment of emission trends by Canadian economic sectors. Chapters 3 to 8 provide descriptions and additional analysis for each broad emission and removal category according to UNFCCC Common

Reporting Format requirements. Chapter 9 presents a summary of recalculations and planned improvements.

Part 2 of the NIR consists of annexes 1 to 11, which provide a key category analysis, detailed explanations of estimation methodologies, a comparison of the sectoral and reference approaches in the Energy Sector, quality assurance and quality control procedures, completeness assessments, inventory uncertainty, emission factors, rounding procedures, a summary of ozone and aerosol precursors, and supplementary information required under articles 7.1 and 3.14 of the Kyoto Protocol.

Part 3 comprises annexes 12 to 15, which present summary tables of GHG emissions for each provincial and territorial jurisdiction, sector and gas, as well as additional details on the GHG intensity of electricity generation and detailed trend tables by province and territory.

This NIR also includes reporting of LULUCF activities under articles 3.3 and 3.4 of the Kyoto Protocol, with emission and removal estimates for afforestation and deforestation (mandatory), and cropland management (elected by Canada) for the years 2008 and 2009. These Kyoto estimates do not affect Canada's national emissions total, and will only be accounted for at the end of the five-year commitment period (2012 inventory year). In addition, Chapter 1 and Annex 11 of this report provide supplementary information required under articles 7.1 and 3.14 of the Kyoto Protocol.

# Chapter 1

## Introduction

### 1.1. Greenhouse Gas Inventories and Climate Change

In order to understand climate change, it is important to differentiate between weather and climate. Weather is the state of the atmosphere at a given time and place and is usually reported as temperature, air pressure, humidity, wind, cloudiness and precipitation. The term “weather” is used mostly when reporting these conditions over short periods of time.

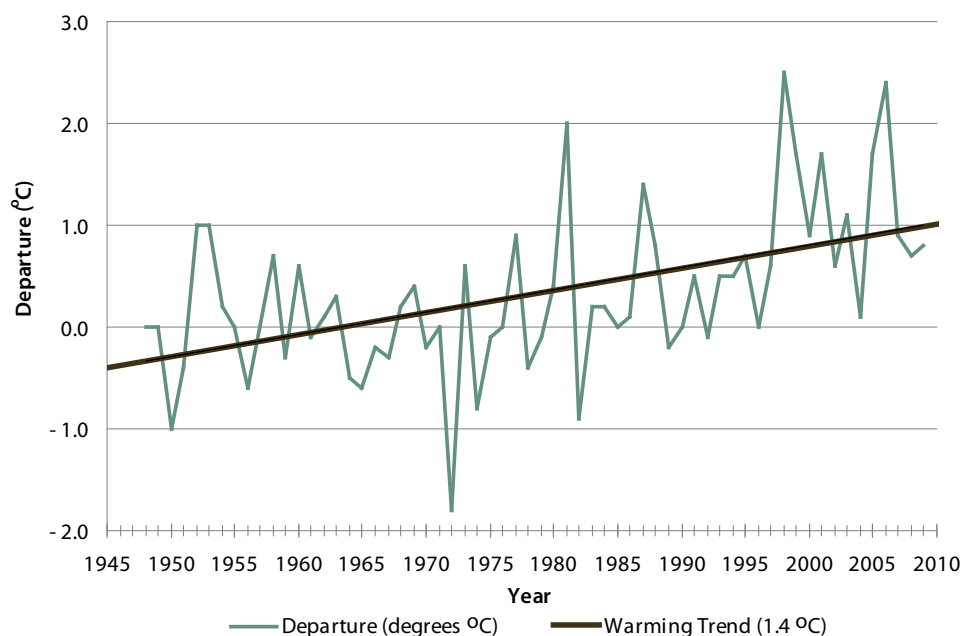
On the other hand, climate is the average pattern of weather (usually taken over a 30-year period) for a particu-

lar region. Climatic elements include precipitation, temperature, humidity, sunshine, wind velocity, phenomena such as fog, frost and hailstorms, and other measures of the weather. The term “climate change” refers to changes in long-term weather patterns caused by natural phenomena and human activities that alter the chemical composition of the atmosphere through the build-up of greenhouse gases (GHGs), which trap heat and reflect it back to the Earth’s surface.

It is now well known that atmospheric concentrations of GHGs have grown significantly since pre-industrial times. Since 1750, the concentration of atmospheric carbon dioxide (CO<sub>2</sub>) has increased by 38%; of methane (CH<sub>4</sub>) by 157%; and of nitrous oxide (N<sub>2</sub>O) by 19% (WMO 2009). Between 1970 and 2004, global GHG emissions due to human activities have increased by approximately 70% (IPCC 2007a). These trends can be largely attributed to fossil fuel use (including energy supply, transportation, residential and commercial buildings and industrial use) and land-use change, including the permanent loss of forest cover.

According to the Intergovernmental Panel on Climate Change’s (IPCC’s) *Fourth Assessment Report* (IPCC 2007b), the impacts of climate change will vary regionally. In general, temperatures and sea levels are expected to rise and the frequency of extreme weather events is expected to in-

Figure 1–1 Annual Canadian Temperature Departures and Long-term Trend, 1948–2009 (°C)



Source data: Environment Canada (2010)

crease. In some regions, the impacts could be devastating, while other regions could benefit from climate change. The impacts will depend on the form and magnitude of the change and, in the case of adverse effects, the ability of natural and human systems to adapt to the changes. In Canada, the impact of climate change may be felt in extreme weather events, the reduction of fresh water resources, increased risk and severity of forest fires and pest infestations, a reduction in arctic ice and an acceleration of glacial melting.

Canada's national average temperature for 2009 was 0.8°C above normal (Figure 1–1). Annual temperatures in Canada have remained above or close to normal since 1993, with a warming trend of 1.4°C over the last 62 years (Environment Canada 2010).

### 1.1.1. Reporting of Canada's National Greenhouse Gas Inventory

Canada ratified the United Nations Framework Convention on Climate Change (UNFCCC) in December 1992, and the Convention came into force in March 1994. The ultimate objective of the UNFCCC is to stabilize GHG concentrations in the atmosphere at a level that would prevent dangerous interference with the climate system. In its actions to achieve its objective and to implement its provisions, the UNFCCC lays out a number of guiding principles and commitments. It requires governments to gather and share information on greenhouse gas emissions, national policies and best practices; to launch national strategies for addressing greenhouse gas emissions and adapting to expected impacts; and to cooperate in preparing for adaptation to the impacts of climate change. Specifically, Articles 4 and 12 and Decision 3/CP.5 of the Convention commit all Parties to develop, periodically update,<sup>1</sup> publish and make available to the Conference of the Parties national inventories of anthropogenic<sup>2</sup> emissions by sources and removals by sinks of all GHGs not controlled by the Montreal Protocol using comparable methodologies.

This National Inventory Report (NIR) provides Canada's annual greenhouse gas emissions estimates for the period 1990–2009. The NIR, along with the Common Reporting

Format (CRF) tables, comprise Canada's submission to the UNFCCC and have been prepared in accordance with Decision 18/CP.8 of the Convention and other relevant decisions.

### 1.1.2. Greenhouse Gases and the Use of Global Warming Potentials (GWPs)

This report provides estimates of Canada's emissions and removals of the following GHGs: CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, sulphur hexafluoride (SF<sub>6</sub>), perfluorocarbons (PFCs) and hydrofluorocarbons (HFCs). In addition, and in keeping with the UNFCCC reporting guidelines for Annex I Parties, Annex 10 contains estimates of the following ozone and aerosol precursors: carbon monoxide (CO), nitrogen oxides (NO<sub>x</sub>), non-methane volatile organic compounds (NMVOC) and sulphur oxides (SO<sub>x</sub>).

#### 1.1.2.1. Carbon Dioxide (CO<sub>2</sub>)

CO<sub>2</sub> is a naturally occurring, colourless, odourless, incombustible gas formed during respiration, combustion, decomposition of organic substances, and the reaction of acids with carbonates. It is present in the Earth's atmosphere at low concentrations and acts as a greenhouse gas. The global carbon cycle is made up of large carbon flows and reservoirs. Through these, CO<sub>2</sub> is constantly being removed from the air by its direct absorption into water and by plants through photosynthesis and, in turn, is naturally released into the air by plant and animal respiration, decay of plant and soil organic matter, and outgassing from water surfaces. Small amounts of carbon dioxide are also injected directly into the atmosphere by volcanic emissions and through slow geological processes such as the weathering of rock (Hengeveld et al. 2005). Although human-caused releases of CO<sub>2</sub> are relatively small compared to the amounts that enter and leave the atmosphere due to the natural active flow of carbon, and represent approximately 1/20 (Hengeveld et al. 2005), human activities now appear to be significantly affecting this natural balance. This is evident in the measurement of the steady increase of atmospheric CO<sub>2</sub> concentrations since pre-industrial times across the globe. Anthropogenic sources of CO<sub>2</sub> emissions include the combustion of fossil fuels and biomass to produce energy, building heating and cooling, transportation, land-use changes including deforestation, the manufacture of cement and other industrial processes.

<sup>1</sup> Annex I Parties (or developed countries) are required to submit a national inventory annually by April 15.

<sup>2</sup> Anthropogenic refers to human-induced emissions and removals that occur on managed lands.

### 1.1.2.2. Methane (CH<sub>4</sub>)

CH<sub>4</sub> is a colourless, odourless, flammable gas that is the simplest hydrocarbon. CH<sub>4</sub> is present in the Earth's atmosphere at low concentrations and acts as a greenhouse gas. CH<sub>4</sub> usually in the form of natural gas, is used as feedstock in the chemical industry (e.g. hydrogen and methanol production), and as fuel for various purposes (e.g. heating homes and operating vehicles). CH<sub>4</sub> is produced naturally during the decomposition of plant or organic matter in the absence of oxygen, as well as released from wetlands (including rice paddies), and through the digestive processes of certain insects and animals such as termites, sheep and cattle. CH<sub>4</sub> is also released from industrial processes, fossil fuel extraction, coal mines, incomplete fossil fuel combustion and garbage decomposition in landfills.

### 1.1.2.3. Nitrous Oxide (N<sub>2</sub>O)

N<sub>2</sub>O is a colourless, non-flammable, sweet-smelling gas that is heavier than air. Used as an anaesthetic in dentistry and surgery, as well as a propellant in aerosol cans, N<sub>2</sub>O is most commonly produced via the heating of ammonium nitrate (NH<sub>4</sub>NO<sub>3</sub>). It is also released naturally from oceans, by bacteria in soils, and from animal wastes. Other sources of N<sub>2</sub>O emissions include the industrial production of nylon and nitric acid, combustion of fossil fuels and biomass, soil cultivation practices, and the use of commercial and organic fertilizers.

### 1.1.2.4. Perfluorocarbons (PFCs)

PFCs are a group of human-made chemicals composed of carbon and fluorine only. These powerful greenhouse gases were introduced as alternatives to ozone-depleting substances (ODSs) such as chlorofluorocarbons (CFCs) in manufacturing semiconductors. PFCs are also used as solvents in the electronics industry, and as refrigerants in some specialized refrigeration systems. In addition to being released during consumption, they are emitted as by-product during aluminium production.

### 1.1.2.5. Hydrofluorocarbons (HFCs)

HFCs are a class of human-made chemical compounds that contain only fluorine, carbon and hydrogen, and are powerful greenhouse gases. As HFCs do not deplete the ozone layer, they are commonly used as replacements for ozone-depleting substances (ODSs), such as chlorofluorocarbons (CFCs), hydrochlorofluorocarbons (HCFCs) and

halons in various applications including refrigeration, fire-extinguishing, semiconductor manufacturing and foam blowing.

### 1.1.2.6. Sulphur hexafluoride (SF<sub>6</sub>)

SF<sub>6</sub> is a synthetic gas that is colourless, odourless, and non-toxic (except when exposed to extreme temperatures), and acts as a greenhouse gas due to its very high heat-trapping capacity. SF<sub>6</sub> is primarily used in the electricity industry as insulating gas for high-voltage equipment. It is also used as a cover gas in the magnesium industry to prevent oxidation (combustion) of molten magnesium. In lesser amounts, SF<sub>6</sub> is used in the electronics industry in the manufacturing of semiconductors, and also as a tracer gas for gas dispersion studies in industrial and laboratory settings.

### 1.1.2.7. Global Warming Potentials

It should be noted that greenhouse gases are not equal. In fact, each greenhouse gas has a unique atmospheric lifetime and heat-trapping potential. Therefore, to interpret the emission data presented in this report, it is important to understand that the radiative forcing<sup>3</sup> effect of a gas within the atmosphere is a reflection of its ability to cause atmospheric warming. Direct effects occur when the gas itself is a GHG, whereas indirect radiative forcing occurs when chemical transformation of the original gas produces a gas or gases that are GHGs or when a gas influences the atmospheric lifetimes of other gases. The global warming potential (GWP) of a GHG takes into account both the instantaneous radiative forcing due to an incremental concentration increase and the lifetime of the gas and is a relative measure of the warming effect that the emission of a radiative gas (i.e. a GHG) might have on the surface troposphere.

By definition, a global warming potential is the time-integrated change in radiative forcing due to the instantaneous release of 1 kg of the gas expressed relative to the radiative forcing from the release of 1 kg of CO<sub>2</sub>. The concept of a global warming potential has been developed to allow scientists and policy-makers to compare the ability of each greenhouse gas to trap heat in the atmosphere relative to CO<sub>2</sub>. Often greenhouse gas emissions are calculated in terms of how much CO<sub>2</sub> would be required

<sup>3</sup> The term "radiative forcing" refers to the amount of heat-trapping potential for any given GHG. It is measured in units of power (watts) per unit of area (metres squared).

Table 1–1 1995 IPCC GWPs and Atmospheric Lifetimes

GHG	Formula	100-Year GWP	Atmospheric Lifetime (years)
Carbon Dioxide	CO <sub>2</sub>	1	Variable
Methane	CH <sub>4</sub>	21	12 ± 3
Nitrous Oxide	N <sub>2</sub> O	310	120
Sulphur Hexafluoride	SF <sub>6</sub>	23 900	3 200
Hydrofluorocarbons (HFCs)			
HFC-23	CHF <sub>3</sub>	11 700	264
HFC-32	CH <sub>2</sub> F <sub>2</sub>	650	5.6
HFC-41	CH <sub>3</sub> F	150	3.7
HFC-43-10mee	C <sub>5</sub> H <sub>2</sub> F <sub>10</sub>	1 300	17.1
HFC-125	C <sub>2</sub> HF <sub>5</sub>	2 800	32.6
HFC-134	C <sub>2</sub> H <sub>2</sub> F <sub>4</sub> (CHF <sub>2</sub> CHF <sub>2</sub> )	1 000	10.6
HFC-134a	C <sub>2</sub> H <sub>2</sub> F <sub>4</sub> (CH <sub>2</sub> FCF <sub>3</sub> )	1 300	14.6
HFC-143	C <sub>2</sub> H <sub>3</sub> F <sub>3</sub> (CHF <sub>2</sub> CH <sub>2</sub> F)	300	3.8
HFC-143a	C <sub>2</sub> H <sub>3</sub> F <sub>3</sub> (CF <sub>3</sub> CH <sub>3</sub> )	3 800	48.3
HFC-152a	C <sub>2</sub> H <sub>4</sub> F <sub>2</sub> (CH <sub>3</sub> CHF <sub>2</sub> )	140	1.5
HFC-227ea	C <sub>3</sub> HF <sub>7</sub>	2 900	36.5
HFC-236fa	C <sub>3</sub> H <sub>2</sub> F <sub>6</sub>	6 300	209
HFC-245ca	C <sub>3</sub> H <sub>3</sub> F <sub>5</sub>	560	6.6
Perfluorocarbons (PFCs)			
Perfluoromethane	CF <sub>4</sub>	6 500	50 000
Perfluoroethane	C <sub>2</sub> F <sub>6</sub>	9 200	10 000
Perfluoropropane	C <sub>3</sub> F <sub>8</sub>	7 000	2 600
Perfluorobutane	C <sub>4</sub> F <sub>10</sub>	7 000	2 600
Perfluorocyclobutane	c-C <sub>4</sub> F <sub>8</sub>	8 700	3 200
Perfluoropentane	C <sub>5</sub> F <sub>12</sub>	7 500	4 100
Perfluorohexane	C <sub>6</sub> F <sub>14</sub>	7 400	3 200

Sources:

GWP: IPCC, 1995. Available online at [http://unfccc.int/ghg\\_data/items/3825.php](http://unfccc.int/ghg_data/items/3825.php)

Atmospheric Lifetime: IPCC, 1995, Table 2.9.

Note:

The CH<sub>4</sub> GWP includes the direct effect and those indirect effects due to the production of tropospheric ozone and stratospheric water vapour. Not included is the indirect effect due to the production of CO<sub>2</sub>.

Figure 1–2 Per Capita GHG Emission Trend for Canada, 1990–2009

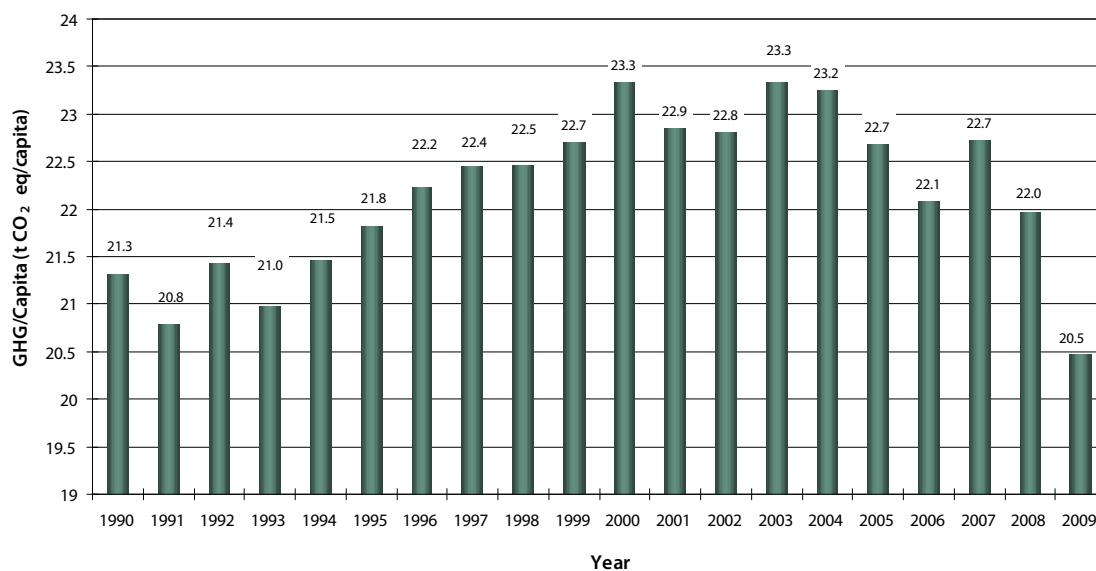
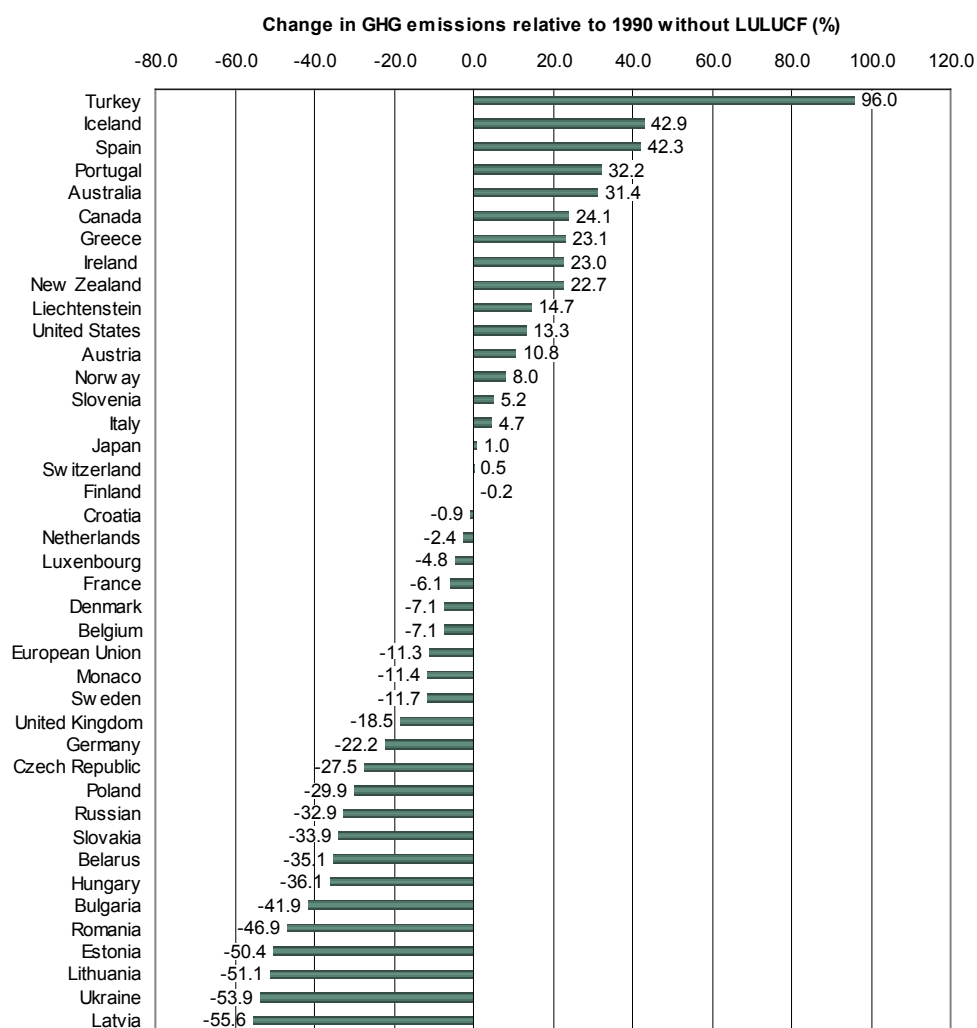


Figure 1–3 Change in Aggregate GHG Emissions for Annex I Parties without LULUCF, 1990–2008



Source: UNFCCC (2010).

to produce a similar warming effect. This is called the carbon dioxide equivalent (CO<sub>2</sub> eq) value and is calculated by multiplying the amount of the gas by its associated GWP. For example, the 100-year global warming potential for methane (CH<sub>4</sub>) is 21. As such, an emission of one hundred kilotonnes (100 kt) of methane is equivalent to 21 x 100 kt = 2100 kt CO<sub>2</sub> eq.

Consistent with Decision 2/CP.3, the 100-year GWPs, provided by the IPCC in its *Second Assessment Report* (Table 1–1) and required for inventory reporting under the UNFCCC, are used in this report.

### 1.1.3. Canada's Contribution

While Canada represented only about 2% of total global GHG emissions in 2005 (Environment Canada 2011), it is one of the highest per capita emitters, largely as a result of its size, climate (i.e. energy demands due to climate), and resource-based economy. In 1990, Canadians released 21.3 tonnes (t) of GHGs per capita. In 2009, this had decreased to 20.5 t of GHGs per capita (Statistics Canada 2010) (Figure 1–2).

In terms of growth in total anthropogenic GHG emissions without Land Use, Land-use Change and Forestry (LU-LUCF), Canada ranks sixth among the Annex I Parties, with an increase in emissions of 24.1% over the 1990–2008 period (Figure 1–3), and ranks first among the G8 countries.



## 1.2. Institutional Arrangements for Inventory Preparation

The following section describes the national system, national registry and the roles and responsibilities of the various agencies and players in the implementation of the national system in Canada. The process for the preparation of the inventory is outlined in Section 1.3. Annex 11 contains additional details on the specific requirements of Article 7 of the Kyoto Protocol, including additional details on the national registry and supplementary information required under Articles 3.3 and 3.4.

The national entity responsible for Canada's national inventory system is the Pollutant Inventories and Reporting Division of Environment Canada. The National Inventory Focal Point is:

Director  
Pollutant Inventories and Reporting Division  
Science and Risk Assessment Directorate  
Science and Technology Branch  
Environment Canada  
10th Floor, 200 Sacré-Coeur Boulevard  
Gatineau QC K1A 0H3  
Telephone: 819-953-3588  
Fax: 819-956-6047  
Email: Aida.Warah@ec.gc.ca

A detailed description of the functions of the Pollutant Inventories and Reporting Division is provided in the Institutional Arrangements Section (Section 1.2.3).

### 1.2.1. The National System

Under Article 5.1 of the Kyoto Protocol, each Party to the Protocol included in Annex I shall have in place, no later than January 1, 2007, a national system for the estimation of anthropogenic emissions from sources and removals by sinks of all GHGs not controlled by the Montreal Protocol. The national system encompasses the institutional, legal and procedural arrangements necessary to ensure that Parties meet their reporting obligations, that quality inventories are prepared and that proper documentation and archiving occur in order to facilitate third-party review and to assess compliance with targets under the Kyoto Protocol.

Canada's national system was examined in November 2007 during the in-country review of Canada's initial report. The review team concluded that Canada's national system

contained all the necessary elements: institutional arrangements for the preparation of the inventory, including procedures for official approval; a quality assurance/quality control (QA/QC) plan; a working archives system; an adequate description of the process for collecting data and developing estimates; ability to identify key categories and generate quantitative uncertainty analysis; and a process for performing recalculation for improvement of the inventory (UNFCCC 2008).

### 1.2.2. Canada's National Registry

The assessment of compliance with the Kyoto Protocol target is based on a comparison of a country's inventory of total GHG emissions for the 2008–2012 period with its total holdings of Kyoto accounting units for that same period. In accordance with Article 7.4 of the Kyoto Protocol, Canada has put a national registry in place, which went live on February 12, 2010.

### 1.2.3. Institutional Arrangements

The *Canadian Environmental Protection Act*, 1999 (CEPA 1999) provides the legislative authority for Environment Canada to implement a UNFCCC and Kyoto Compliant national inventory system and responsibility for the preparation and submission of the national inventory to the UNFCCC (Canada 1999).

Recognizing the need to draw on the best available technical and scientific expertise and information in accordance with good practice and international quality standards, Environment Canada has defined roles and responsibilities for the preparation of the inventory, both internally and externally.

Sources and sinks of GHGs originate from a tremendous range of economic sectors and activities. As such, Environment Canada is involved in many partnerships with data providers and expert contributors in a variety of ways, ranging from informal to formal arrangements. These partnerships comprise other government departments including Statistics Canada, Natural Resources Canada and Agriculture & Agri-Food Canada which are described in greater detail in the following sections. Environment Canada also has arrangements with industry associations, consultants and universities, as described in 1.2.3.3, and collaborates with provincial and territorial governments on a bilateral basis.

Figure 1–4 Partners of the National System

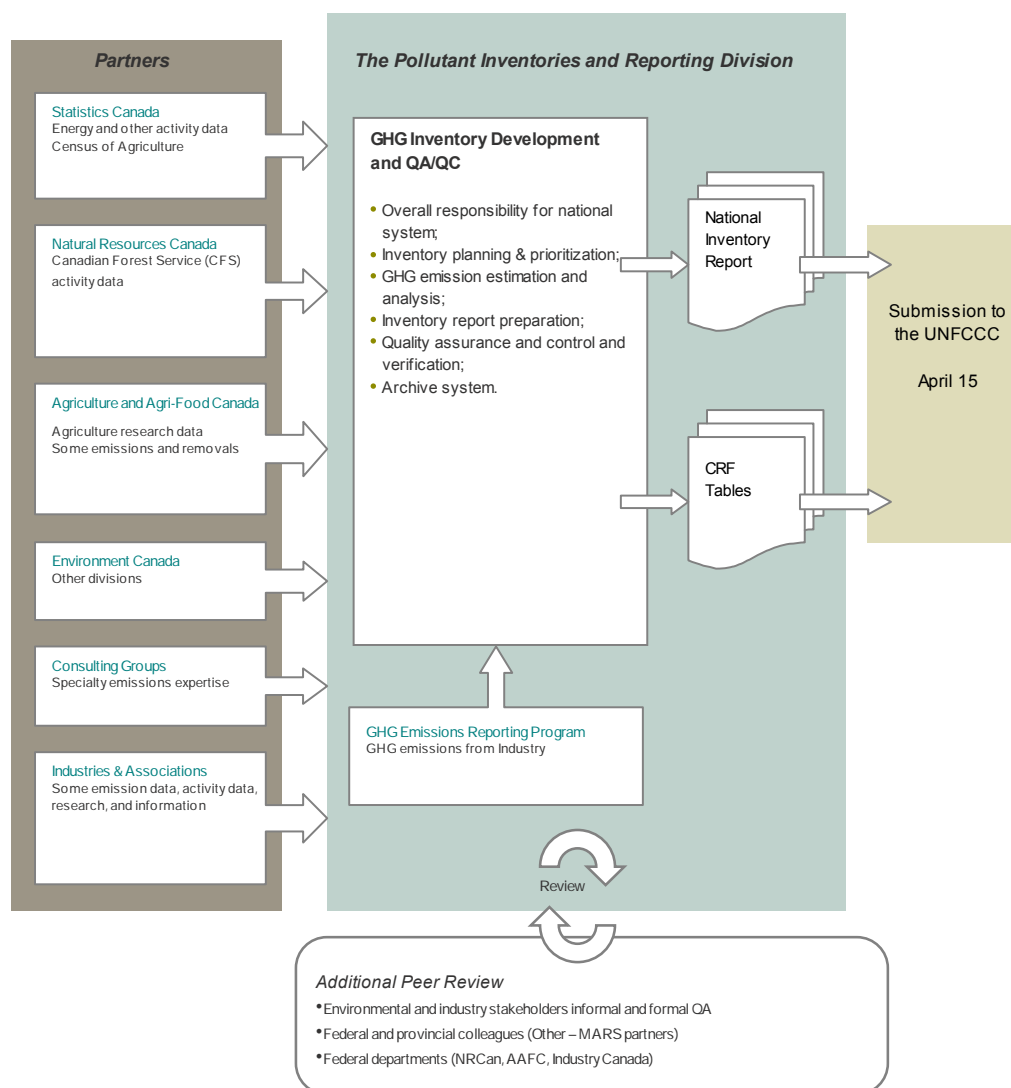


Figure 1–4 identifies the different partners of the inventory agency and their contribution.

### 1.2.3.1. Statistics Canada

Canada's national statistical agency, Statistics Canada, provides Environment Canada with a large portion of the underlying activity data to estimate GHG emissions for the Energy and the Industrial Processes Sectors. Statistics Canada is responsible for the collection, compilation and dissemination of Canada's energy balance in its Annual Report on Energy Supply–Demand in Canada (RESD). The energy balance is transmitted annually to Environment Canada according to the terms of a Memorandum of Understanding established between the two departments.

Statistics Canada also conducts an annual Industrial Consumption of Energy (ICE) survey, which is a comprehensive survey of industries that feeds into the development of the energy balance.

Statistics Canada's quality management system for the energy balance includes an internal and external review process. Owing to the complexity of energy data, a Working Group on Energy Statistics consisting of members from Statistics Canada, Environment Canada and Natural Resources Canada (NRCan) was established to provide advice, direction and recommendations on improvements to the energy balance. In addition, a high-level Energy Steering Committee was formed in 2008 to review timing, quality and technical issues related to the RESD and ICE data. Refer to Annex 2 of this report for additional informa-

tion on the use of the energy balance in the development of energy estimates.

Statistics Canada is also responsible for gathering other energy data such as mining and electricity information, and other non-energy-related industrial information, including urea and ammonia production. In addition, the statistics agency collects agricultural activity data (related to crops, crop production and management practices) through the Census of Agriculture and provides data on other activities such as animal population data.

### 1.2.3.2. Natural Resources Canada and Agriculture and AgriFood Canada: Canada's Monitoring and Accounting System for Land Use, Land-use Change and Forestry

Since 2005, Environment Canada has officially designated responsibilities to Agriculture and AgriFood Canada (AAFC) and the Canadian Forest Service of Natural Resources Canada (NRCan/CFS) for the development of key components of the Land Use, Land-use Change and Forestry (LULUCF) Sector and has established formal and explicit governance mechanisms to that effect through Memoranda of Understanding. Canada's Monitoring, Accounting and Reporting System (MARS) for LULUCF is overseen by an interdepartmental steering committee chaired by Environment Canada, with representatives from AAFC and NRCan-CFS. Technical working groups address the subsectors of Forestry, Agriculture and Land-use Change, to ensure the best available information and data from scientific research are integrated into the LULUCF Sector of the inventory. Under this framework, Environment Canada's partners provide estimates, complete and transparent documentation, uncertainty analyses and quality control reports.

NRCan/CFS has developed the National Forest Carbon MARS which has contributed major improvements to the LULUCF Sector. This program annually develops and delivers estimates for forest land, land conversion to forest land (afforestation) and forest land converted to other land (deforestation).

AAFC has developed the Canadian Agricultural Greenhouse Gas MARS, which also significantly enhanced the quality of the LULUCF Sector. In concert with NRCan/CFS, AAFC delivers cropland estimates for the LULUCF sector that include the effect of management practices on agri-

cultural soils and the residual impact of land conversion to cropland. In addition, AAFC provides scientific support to the Agriculture Sector of the inventory.

Environment Canada manages and coordinates the annual inventory development process, develops other LULUCF estimates, undertakes cross-cutting quality control and quality assurance, and generally ensures the consistency of land-based estimates through an integrated land representation system.

In addition, the Earth Science Sector of NRCan contributes earth observation expertise, while the Canadian Space Agency has supported the development of Earth Observation products to improve land information within LULUCF MARS.

### 1.2.3.3. Other Partnerships

In addition to its support to Canada's MARS for LULUCF (section 1.2.3.2 above), Natural Resources Canada (NRCan) provides energy expertise and analysis, serves as expert reviewer for the Energy data and collects and provides activity data on mineral production, ethanol consumption and wood residues. Road vehicle fuel efficiency data are provided by both Transport Canada and Natural Resources Canada.

When required, and resources permitting, contracts are established with consulting firms and universities to conduct in-depth studies—for example, on updating emission factors. A bilateral agreement with the Aluminum Association of Canada (AAC) has been signed, under which process-related emission estimates for CO<sub>2</sub>, PFCs and SF<sub>6</sub> are to be provided annually to Environment Canada. A similar agreement has been negotiated with the Canadian Electricity Association (CEA) for provision of SF<sub>6</sub> emissions and supplementary data relating to power transmission systems. Environment Canada has also been collaborating with magnesium casting companies, member companies of the Canadian Fertilizer Institute and companies that import or distribute HFCs, with regard to their annual data on GHG emissions and/or supporting activity data.

## 1.3. Process for Inventory Preparation

This section describes in general terms the annual inventory development cycle from the planning phase to the submission to the UNFCCC. Continuous data collection and improvements are integral parts of the national inventory planning and quality management cycles (see Section 1.6).

The inventory schedule is built around a continuous process of methodological improvements, refinements and review, according to the quality management and improvement plans. The Inventory Coordinator is responsible for preparing the inventory schedule based on the results of the lessons-learned review of the previous inventory cycle, QA/QC follow-up, the UNFCCC review report, collaboration with provincial and territorial governments and the improvement plan to identify priorities and areas for improvement. Based on these outcomes, methodologies and emission factors are reviewed, developed and/or refined if necessary during the period from May to September. QA reviews of methodologies and emission factors are undertaken for categories for which a change in methodology or emission factor is proposed and for categories that are scheduled for a QA review of methodology or emission factor.

By the end of October, methodologies are finalized and the data collection process is almost complete. The data used to compile the national inventory are generally from published sources. Data are collected either electronically or manually (hard copies) from the source agencies and are entered into spreadsheet-based emission accounting systems, databases and/or models and controlled for quality. Between November and January, draft estimates and a national report are prepared by industry experts. Emissions are calculated by designated inventory experts, reviewed internally and then reported according to UNFCCC guidelines in the CRF and the NIR. QC checks and estimates are signed off by sectoral managers before the report and national totals are prepared. The inventory process also involves key category assessment, recalculations, uncertainty work and documentation preparation.

Over the months of February and March, the compiled inventory is first reviewed internally and components of it are externally reviewed by experts, government agencies and provincial and territorial governments, after which the

NIR is fully edited. Comments from the review are documented and, where appropriate, incorporated in the NIR and CRF, which are normally submitted to the UNFCCC electronically prior to April 15 of each year. Initial checks on the April submission are performed by the UNFCCC in May and June. A final inventory report is prepared and submitted, if necessary. Once finalized, the NIR is then further edited, translated and readied for publication.

### 1.3.1. Procedures for the Official Consideration and Approval of the Inventory

Typically, the NIR is compiled annually by February. Once completed, the draft NIR and a summary of the data and trends analysis is prepared for approval for submission to the UNFCCC Secretariat. In the process of considering the national inventory and the results, several briefings of senior officials typically take place prior to the report being sent to the Minister. Once approved, the National Inventory Focal Point prepares a letter of submission to accompany the NIR and CRF tables, which are then sent electronically.

## 1.4. Methodologies and Data Sources

The inventory is structured to match the reporting requirements of the UNFCCC and is divided into the following six main Sectors: Energy, Industrial Processes, Solvent and Other Product Use, Agriculture, LULUCF, and Waste. Each of these Sectors is further subdivided within the inventory. The methods described have been grouped, as closely as possible, by UNFCCC Sector and subsector.

The methodologies contained in the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997), the *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000), and the *Good Practice Guidance for Land Use, Land-Use Change and Forestry* (IPCC 2003) are followed to estimate emissions and removals of each of the following direct GHGs: CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, HFCs, PFCs and SF<sub>6</sub>. The 2006 IPCC Guidelines for National Greenhouse Gas Inventories (IPCC 2006) contains updated methodologies; however, these guidelines have not yet been adopted by the UNFCCC.

While not mandatory, the new UNFCCC reporting guidelines encourage Annex I Parties to provide information on the following indirect GHGs: SO<sub>x</sub>, NO<sub>x</sub>, CO and NMVOCs. For all categories except LULUCF, these gases (referred to as criteria air contaminants, or CACs) are inventoried and reported separately. CAC emissions in Canada are reported to the United Nations Economic Commission for the Environment.<sup>4</sup> As noted, a summary of these emissions is also included in the NIR (see Annex 10: Ozone and Aerosol Precursors).

In general, an emissions and removals inventory can be defined as a comprehensive account of anthropogenic sources of emissions and removals by sinks and associated data from source categories within the inventory area over a specified time frame. It can be prepared “top-down,” “bottom-up,” or using a combination approach. Canada’s national inventory is prepared using a “top-down” approach, providing estimates at a sectoral and provincial/territorial level without attribution to individual emitters.

Emissions or removals are usually calculated or estimated using mass balance, stoichiometry or emission factor relationships under average conditions. In many cases, activity data are combined with average emission factors to produce a “top-down” national inventory. Large-scale regional estimates, based on average conditions, have been compiled for diffuse sources, such as transportation. Emissions from landfills are determined using a simulation model to account for the long-term slow generation and release of these emissions.

Manipulated biological systems, such as agricultural lands, forestry and land converted to other uses, are typical sources or sinks diffused over very large areas. Processes that cause emissions and removals display considerable spatial and interannual variability, and they also span several years or decades. The most practical approach to estimating emissions and removals may require a combination of repeated measurements and modelling. The need, unique to these systems, to separate anthropogenic impacts from large natural fluxes creates an additional challenge.

The methodologies (Annexes 2 and 3) and emission factors (Annex 8) described in this document are considered to be the best available to date, given the available activity data. That being said, in some cases, a more accurate method or emission factor may be available, but the necessary activ-

ity data are lacking at the national level, so the more accurate method cannot be used. Some methods have undergone revision and improvement over time, and some new sources have been added to the inventory over time.

Methodology and data improvement activities, which take into account results of QA/QC procedures, reviews and verification, are planned and implemented on a continuous basis by the staff of Environment Canada’s Pollutant Inventories and Reporting Division. These methodology and data improvement activities are carried out with a view to further refining and increasing the transparency, completeness, accuracy, consistency and comparability of the national inventory. As a result, changes in data or methods often lead to the recalculation of GHG estimates for the entire time series, from the 1990 base year to the most recent year available. Further discussion of recalculations and improvements can be found in Chapter 9.

### 1.4.1. Mandatory GHG Reporting

In March 2004, the Government of Canada established the Greenhouse Gas Emissions Reporting Program (GHGRP) under section 46(1) of CEPA 1999 to collect GHG emissions information annually from Canadian facilities on a mandatory basis.

Greenhouse gas information reported under the GHGRP is collected through Environment Canada’s Single Window Reporting system. Environment Canada launched this system to support integrated data collection to allow industry to submit information which is common to multiple programs and jurisdictions only once.

Under the auspices of the Canadian Council of Ministers of the Environment (CCME), the Canadian federal, provincial and territorial governments agreed to work together to minimize duplication and reduce the reporting burden for industry and governments through the use of a one window system to collect GHG information. Environment Canada’s Single Window Reporting system was expanded to support an inclusive Canadian approach for GHG reporting. Provincial partners using this system to collect GHG information to meet their GHG reporting regulations include Alberta, British Columbia and Ontario.

The GHGRP applies to industrial and other facilities that are the largest emitters of GHGs and sets out basic reporting requirements. The program’s four main objectives are to:

<sup>4</sup> Available online at <http://www.ceip.at>



- provide Canadians with timely information on GHG emissions;
- enhance the level of detail in the National Greenhouse Gas Inventory;
- support provincial and territorial requirements for GHG emissions information; and
- support the development of regulations.

The types of large industrial facilities reporting GHG emissions include:

- power generation plants that use fossil fuels to produce electricity, heat or steam;
- integrated steel mills;
- oil and gas extraction operations;
- facilities involved in mining, smelting and refining of metals;
- pulp, paper and sawmills;
- petroleum refineries; and
- chemical producers.

Information gathered through the GHGRP from these large industrial facilities supports policy decisions and the potential development of future GHG regulations.

As published in the Canada Gazette, facilities that have emissions of 100 kt CO<sub>2</sub> eq or more annually for the years 2004–2008 had to submit a GHG emission report by June 1 of the following year. As of 2009, the reporting threshold changed to 50 kt. Voluntary submissions from facilities with GHG emissions below the reporting threshold are encouraged.

Specific estimation methods are not prescribed, and reporters can choose the quantification methodologies most appropriate for their own particular industry or application. However, reporting facilities must use methods for estimating emissions that are consistent with the guidelines adopted by the UNFCCC and developed by the IPCC and used in the preparation of the national GHG inventory.

Environment Canada's GHGRP website ([www.ec.gc.ca/ges-ghg/default.asp?lang=En&n=040E378D-1](http://www.ec.gc.ca/ges-ghg/default.asp?lang=En&n=040E378D-1)) provides public access to the reported GHG emission information (GHG totals by gas by facility)

#### 1.4.1.1. Use of Reported GHG Data in the National Inventory Report

Facility-level GHG emission data are used, where appropriate, to confirm emission estimates in the NIR developed

Table 1–2 Facility-Reported 2009 GHG Emissions by Province

Province	Number of Facilities	Total Emissions (kt CO <sub>2</sub> eq)	Percentage of Total Emissions (%)
Newfoundland and Labrador	8	4 378	2%
Prince Edward Island	1	74	0%
Nova Scotia	12	10 773	4%
New Brunswick	14	10 118	4%
Quebec	69	20 341	8%
Ontario	125	49 179	20%
Manitoba	12	2 132	1%
Saskatchewan	31	22 428	9%
Alberta	160	117 358	47%
British Columbia	62	13 154	5%
Northwest Territories	28	519	0%
Total	522	250 454	100%

Note: Totals may not add up due to rounding.

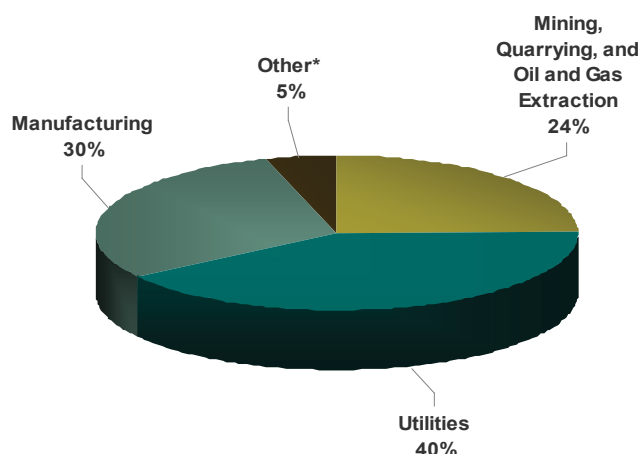
from national and provincial statistics. The extent to which the reported GHG emission information can be fully integrated is dependent upon the level of detail and type of data available. Environment Canada will continue to use these data as an important component of the overall inventory development process in comparing and verifying the inventory estimates.

#### 1.4.1.2. Effect of Threshold Change on Reported Emissions, 2008–2009

For the last reporting period (i.e. 2009), the mandatory reporting threshold was lowered from 100 kt CO<sub>2</sub> eq to 50 kt CO<sub>2</sub> eq. This threshold change resulted in a 49% increase in the number of facilities reporting to the program between 2008 and 2009 (522 for 2009, up from 350 for 2008), and the emissions from the new facilities equalled 11 megatonnes (Mt) (4%) of the 2009 total reported emissions.

As expected, the lowering of the mandatory reporting threshold had a significant effect on the number of new facilities reporting for 2009. Of the 159 facilities reporting for the first time, 116 had emissions between 50 and 100 kt. Facilities with emissions in this range were only required to report in 2009. The number of voluntary reporters (i.e.

Figure 1–5 Facility-Reported 2008 GHG Emissions by Industrial Sector



\*“Other” includes Transportation and Warehousing, as well as Administrative and Support, Waste Management and Remediation Services industrial sectors.

facilities with emissions below 50 kt CO<sub>2</sub> eq) also increased to 72 facilities in 2009.

#### 1.4.1.3. Reported 2009 Facility GHG Emissions

In its sixth year of reporting, the collected GHG data cover the period from 2004 to 2009. Under the new reporting threshold of 50 kt CO<sub>2</sub> eq, a total of 522 facilities reported GHG emissions for the 2009 calendar year, collectively emitting a total of 250 Mt of GHGs.<sup>5</sup> Facilities can voluntarily report their GHG emissions if their emissions are below the reporting threshold, and 72 facilities did so in 2009. Facilities in Alberta accounted for the largest share of reported emissions, with approximately 47% of the total, followed by those in Ontario with 20%. Next were Saskatchewan and Quebec, which accounted for 9% and 8% of reported emissions, respectively (Table 1–2).

When completing the GHG report, a reporter is required to identify the main activities occurring at its facility by selecting the North American Industry Classification System (NAICS) code that corresponds to these activities. In 2009, three industrial sectors accounted for the majority of GHG emissions: Utilities, primarily those generating electricity, representing 40%; Manufacturing, accounting for 30%; and Mining, Quarrying, and Oil and Gas Extraction, accounting for 24% (Figure 1–5).

For more information on the facility data reported under Environment Canada's GHGRP, including short-term and long-term changes, please see the Environment Canada publication *Overview of the Reported 2009 Greenhouse Gas Emissions*, available online at [www.ec.gc.ca/ges-ghg/default.asp?lang=En&n=8044859A-1](http://www.ec.gc.ca/ges-ghg/default.asp?lang=En&n=8044859A-1).

## 1.5. Key Categories

The IPCC Good Practice Guidance (IPCC 2000, 2003) defines procedures (in the form of decision trees) for the choice of estimation methods recommended in the IPCC Guidelines. The decision trees formalize the choice of estimation method most suited to national circumstances, considering at the same time the need for accuracy and the available resources (both financial and human). Generally, the precision and accuracy of inventory estimates can be improved by using the most rigorous (highest-tier) methods; however, owing to practical limitations, the exhaustive development of all emissions categories is not possible. Therefore, it is good practice to identify and prioritize key categories in order to make the most efficient use of available resources.

In this context, a *key category* is one that is prioritized within the national inventory system because its estimate has a significant influence on a country's total inventory of direct GHG emissions in terms of the absolute level of emissions (level assessment), the trend in emissions from the base year to the current year (trend assessment), or both. As

<sup>5</sup> Data presented are current as of September 23, 2010.



much as possible, two important inventory aspects of key categories should receive special consideration:

- preferential use of detailed, higher-tier methods; and
- additional attention with respect to QA/QC.

In the absence of quantitative data on uncertainties, a simplified Tier 1 method of identifying key categories provides a good approximation of those areas to which priority should be given to improve inventory estimates.

For the 1990–2009 GHG inventory, level and trend key category assessments were performed according to the Tier 1 approach, as presented in the IPCC Good Practice Guidance (IPCC 2000, 2003). The emission and removal categories used for the key category assessment generally follow those in the CRF and the LULUCF CRF; however, they have been aggregated in some cases and are specific to the Canadian inventory.

Major key categories based on the level and trend assessments (including LULUCF) are the fuel combustion categories (Stationary Combustion – Gaseous, Liquid and Solid Fuels, Road Transportation, and Off-road Transport), and the LULUCF category Forest Land Remaining Forest Land. Details and results of the assessments are presented in Annex 1.

## 1.6. Quality Assurance/Quality Control

The national inventory and NIR must be prepared in accordance with international reporting guidelines and methods agreed to by the UNFCCC, including methodological procedures and guidelines prescribed by the IPCC. QA/QC and verification procedures are an integral part of the preparation of the inventory. The Pollutant Inventories and Reporting Division annually conducts QA/QC activities and is committed to improving data and methods in collaboration with industry, the provinces and territories, academia and the international community to ensure that a credible and defensible inventory is developed. Improvement activities, which take into account results of QA/QC procedures, reviews and verification, are planned and implemented on a continuous basis to further refine and increase the transparency, completeness, accuracy, consistency and comparability of the national inventory. As a result, changes in data or methods often lead to the recalculation of GHG estimates for the entire time series, from the 1990 base year to the most recent year available.

The reader is referred to Annex 6 of this report for more information on quality assurance/quality control.

## 1.7. Inventory Uncertainty

While national GHG inventories should be accurate, complete, comparable, transparent and verifiable, estimates will always inherently carry some uncertainty. Uncertainties<sup>6</sup> in the inventory estimates may be caused by systematic model uncertainty or (more likely) due to random uncertainties present within the input parameters. While reducing model uncertainty requires in-depth reviews of the estimation models, random uncertainties may be reduced by improvements to the activity data regimes and evaluation of emission factors and other model parameters. IPCC guidelines specify that the primary purpose of quantitative uncertainty information is to assist in setting priorities to improve future inventories and to guide decisions about which methods to use. Typically, the uncertainties associated with the trends and the national totals are much lower than those associated with individual gases and sectors.

Annex 7 presents the uncertainty assessment for Canadian GHG emissions. While more complex (Tier 2) methods are in some cases applied to develop uncertainty estimates at the sectoral or category level, for the inventory as a whole these uncertainties were combined with the simple (Tier 1) error propagation method, using Table 6.1 in IPCC (2000). Separate analyses were conducted for the inventory as a whole with and without LULUCF. The calculation of trend uncertainties was only performed without the LULUCF Sector. For further details on uncertainty related to specific sectors, see the uncertainty sections throughout chapters 3 to 8.

The uncertainty for the national inventory, not including the LULUCF Sector, is  $\pm 3.9\%$ , consistent with the previously reported ranges of  $-3\%$  to  $+6\%$ . The Energy Sector had the lowest uncertainty at  $\pm 2.4\%$ , while the Agriculture Sector had the highest uncertainty at  $\pm 39\%$ . The Industrial Processes, Solvent & Other Product Use, and Waste Sectors had uncertainties of  $\pm 7.6\%$ ,  $\pm 19.3\%$  and  $\pm 34.4\%$ , respectively.

The categories that make the largest contribution to uncertainty at the national level are:

<sup>6</sup> Inventory definition of “uncertainty”: a general and imprecise term that refers to the lack of certainty (in inventory components) resulting from any causal factor, such as unidentified sources and sinks, lack of transparency, etc. (IPCC 2000).

- Agriculture – Indirect Agricultural Soils  $\text{N}_2\text{O}$ , Fuel Combustion;
- Other Transportation (Off-road)  $\text{N}_2\text{O}$ ;
- Agriculture – Direct Agricultural Soils  $\text{N}_2\text{O}$ ;
- Fuel Combustion – Public Electricity and Heat Combustion  $\text{CO}_2$ ; and
- Waste – Solid Waste Disposal on Land  $\text{CH}_4$ .

The uncertainty when the LULUCF emissions and removals are included in the national total was found to be 7.1%.

The trend uncertainty, not including LULUCF, was found to be 0.71%. Therefore, the total increase in emissions since 1990 has a 95% chance of being in the range of 16.2–17.6%.

## 1.8. Completeness Assessment

The national GHG inventory, for the most part, is a complete inventory of the six GHGs required under the UNFCCC. The exclusion of some emissions typically relates to the unavailability of comprehensive activity data for certain subcategories of a source that are minor by nature. In some cases, the lack of appropriate and cost-effective methodologies has been the reason for exclusion of a minor source.

The Energy Sector has, since the 2007 UNFCCC in-country review, included biodiesel in transport as recommended by the expert review team. In the Industrial Processes Sector, emissions of  $\text{CH}_4$  from process sources are included in this submission. In the Agriculture Sector,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions from crop residue burning are estimated. In the LULUCF Sector, significant improvements have been implemented starting in 2006, but completeness has not yet been fully met due to data limitations. As for the Waste Sector,  $\text{CH}_4$  emissions from industrial wastewater are inventoried in this submission.  $\text{CH}_4$  emissions from unmanaged landfills, and municipal solid waste incineration, are under investigation through a study.

As part of the NIR improvement plans, efforts are continuously being made to identify and assess relevant new sources and sinks for which cost-effective estimation methods are available. Further details on the completeness of the inventory can be found in Annex 5 and in individual Sector chapters.

# Chapter 2

## Greenhouse Gas Emission Trends, 1990–2009

### 2.1. Summary of Emission Trends

In 2009, Canada's greenhouse gas (GHG) emissions, excluding the Land Use, Land-use Change and Forestry (LULUCF) Sector, were 690 Mt, which is a 16.9% increase over 1990 emissions. Between 2008 and 2009, emissions decreased by 5.7%.

Since 1990, major increases in oil and gas production—including export markets—a large increase in the number of motor vehicles and greater reliance on coal electricity generation, have resulted in a significant rise in emissions. More specifically, the 1990–2009 emission growth has resulted primarily from the categories of, Fossil Fuel Industries, Mining, Electricity and Heat Generation, Transportation, Commercial & Institutional, Consumption of Halocarbons and SF<sub>6</sub>, Enteric Fermentation, and Solid Waste Disposal on Land. There have been overall decreases in Manufacturing Industries and Construction, Chemical Industry, and Metal Production.

2009 GHG emissions were approximately 41 Mt lower than in 2005. During this period electric power emissions decreased by 25 Mt as demand fell and coal-fired generation dropped to its lowest level since 1990. Increased hydro-electric generating capacity as well as efforts made in the province of Ontario to decrease coal-fired power generation also contributed to reducing the amount of combustion-generated electricity and, consequently, emissions.

Industrial Process and Energy emissions from the manufacturing area sector resulted in a total drop of 17 Mt (15%) between 2005 and 2009, due to significantly lowered production evidenced by falling manufacturing GDP, particularly in last year of the period.

### 2.2. Emission Trends by Gas

CO<sub>2</sub> is the largest contributor to Canada's greenhouse gas (GHG) emissions. Figure 2–1 shows how the percent contributions of the six GHGs have changed between 1990 and 2009. The proportion of CO<sub>2</sub> has changed only slightly, rising from 77.7% of emissions in 1990 to 78.6% in 2009.

### 2.3. Emission Trends by Category

#### 2.3.1. Energy Sector (2009 GHG emissions, 566 Mt)

Energy-related activities are by far the largest source of GHG emissions in Canada. The Energy Sector includes emissions of all GHGs from the production of fuels and

Figure 2–1 Canada's GHG Emissions by Gas, 1990 and 2009 (excluding LULUCF)

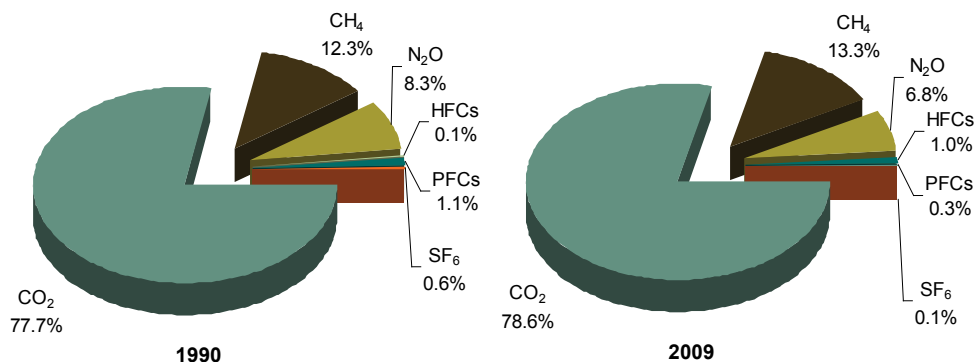


Table 2–1 GHG Emissions from Energy by UNFCCC CRF category, Selected Years

GHG Sources/Sinks	GHG Emissions (Mt CO <sub>2</sub> eq)						
	1990	2004	2005	2006	2007	2008	2009
1. Energy	468	605	595	583	613	597	566
A. Fuel Combustion (Sectoral Approach)	425	541	532	519	551	535	505
1. Energy Industries	143	197	189	181	193	185	161
2. Manufacturing Industries, Construction and Mining	64.5	72.3	68.8	70.4	79.4	72.6	74.9
3. Transport	146	189	193	192	197	196	190
4. Other Sectors	71.6	82.9	80.6	75.3	80.9	80.6	78.7
B. Fugitive Emissions	42.1	64.0	63.1	64.3	62.8	62.3	60.7
1. Solid Fuels (Coal)	1.9	0.7	0.7	0.7	0.8	0.8	0.7
2. Oil and Natural Gas	40.2	63.4	62.4	63.6	62.1	61.5	60.0

Note: Totals may not add up due to rounding.

their combustion for the primary purpose of delivering energy. Emissions in this sector are classified as either combustion-related releases or fugitive releases. Fugitive emissions are defined as intentional or unintentional releases of GHGs from the production, processing, transmission, storage and delivery of fossil fuels.

Overall, fuel combustion and fugitive emissions accounted for 82% of total Canadian GHG emissions in 2009 (505 Mt and 60.7 Mt, respectively). Between 1990 and 2009, fuel combustion-related emissions increased 19%, while emissions from fugitive releases rose by 44%. Emissions for both fuel combustion and fugitive emissions representing selected years are provided in Table 2–1.

The Energy Industries category, grouped in the Energy Sector, is the largest contributor to Canada's emissions. These industries, consisting of Fossil Fuel Production and Refining, Public Electricity and Heat Generation, generate both combustion emissions and fugitive emissions and are calculated as the sum of Fuel Combustion—Energy Industries and Fugitive Emissions in Table 2–1. Emissions from oil and gas extraction, as well as crude bitumen upgrading, are included in the Mining subsector due to how the energy data are collected and aggregated. Altogether, the Energy Industries category and the Fugitive Emissions subsector contributed 222 Mt or 32% of Canada's total and about 39% of the Energy Sector's emissions in 2009.

Table 2–1 divides energy sources by United Nations Framework Convention on Climate Change (UNFCCC) Common Reporting Format (CRF) category: Fuel Combustion is categorized separately from Fugitive Emissions. By this breakdown, fuel combustion in the Energy Industries ac-

counted for 161 Mt in 2009, while fugitive emissions were responsible for 60.7 Mt. In terms of relative growth, fugitive emissions from Oil and Natural Gas (including production, processing, transmission and distribution activities) have increased more rapidly than any other category (as shown in Table 2–1) in the Energy Sector. Between 1990 and 2009, these emissions rose by 49%.

### 2.3.1.1. Emissions from Fuel Combustion (2009 GHG emissions, 505 Mt)

GHG emissions from fuel combustion rose from 425 Mt in 1990 to 505 Mt in 2009, a 19% increase. Fuel combustion emissions are divided into the following subsectors: Energy Industries, Manufacturing Industries, Construction and Mining, Transport and Other Sectors. The Other Sectors subsector comprises emissions from the Residential and Commercial categories, as well as minor contributions of stationary fuel combustion emissions from the Agriculture and Forestry category.

#### Energy Industries (2009 GHG emissions, 161 Mt)

The Energy Industries subsector accounts for the second-largest portion of Canada's fuel combustion emissions (23% of Canada's total), behind Transport. Emissions included in this subsector are from stationary sources producing, processing and refining energy. This source includes Public Electricity and Heat Generation, Petroleum Refining, and Manufacture of Solid Fuels and Other Energy Industries. In 2009, combustion emissions from the Energy Industries category totalled 161 Mt, an increase of 13% from the 1990 level of 143 Mt.

Table 2–2 GHG Emissions from Public Electricity and Heat Generation, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)							Change (%)
	1990	2004	2005	2006	2007	2008	2009	
Electricity Generation	91	122	122	114	122	116	97	6%
Heat Generation	0.2	1.2	0.9	0.8	0.8	0.7	1.0	419%
Total	91.6	123.1	122.8	114.8	122.4	116.2	97.9	7%

Note: Totals may not add up due to rounding.

### Public Electricity and Heat Generation<sup>1</sup> (2009 GHG emissions, 97.9 Mt)

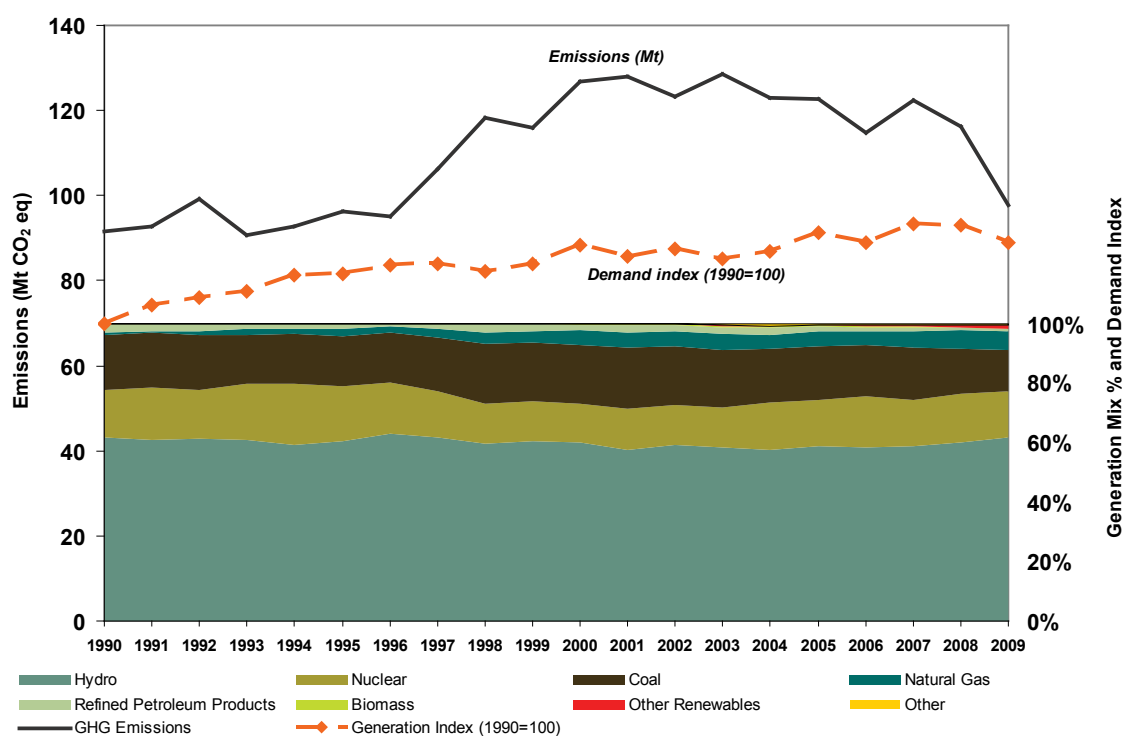
This category accounted for 14% (97.9 Mt) of Canada's 2009 GHG emissions (Table 2–2) and was responsible for 6.3% of the total emission growth between 1990 and 2009. Overall emissions from this category increased 7% (6.2 Mt) since 1990.

Emissions from electricity generation and distribution are unique in that electricity is generated to meet an instantaneous demand and depending on the characteristics

of that demand, the supply source can fluctuate between non-GHG-emitting to high-GHG-emitting sources. This observation is most evident in the last two years, as the economic recession has reduced demand significantly across the country, but specifically in Ontario, a province with a large manufacturing industry. Emissions growth, when compared to 1990, is significantly different than in previous years. In 2009, emissions grew by only 6 Mt (or 7%) compared to 1990, whereas from 2004 to 2008 the typical growth was between 23 and 32 Mt (25 to 34%). Notwithstanding the effects of the recession, rising electricity demand has played a key role in emissions growth due to increasing use of electrically driven manufacturing

1 The Public Electricity and Heat Generation category includes emissions from utilities and industrial generation.

Figure 2–2 Utility-Generated Electricity by Source and GHG Emissions, 1990–2009



#### Notes:

1. Generation statistics refer to utility-based generation only but contribute over 90% of the total supply.
2. Electricity emissions include only utility generation and do not include emissions associated with transmission.

processes, the rapid penetration of computers, increasing use of electronic equipment and a continued influx of electronic consumer goods (NRCan 2010a). Meanwhile, exports of electricity to the United States (mainly from Quebec and Manitoba) have more than doubled (Statistics Canada 2010a). The increase in domestic demand in conjunction with increasing exports has meant that the amount of electricity generated in Canada has increased by 27% from 1990 to 2009.<sup>2</sup>

Emissions, however, have not always followed the trend in electricity generation. During the early 1990s, even with rising demand, emissions from electricity generation oscillated above and below 1990 levels; then, from 1994 to 2000, emissions rose 37%, though generation increased by only about 9%. After a brief pause, emissions peaked in 2003, following which they decreased by 24% over the next six years—even though generation rose by more than 4%. Excluding 2009, the decrease in emissions between 2003 and 2008 was still about 10%, with generation up by over 9%.

Figure 2–2 illustrates the different sources and changes over time of electricity generation between 1990 and 2009.

In terms of electricity supply, the most significant driver relates to changes in the availability of energy sources that can be used to generate electricity. In 2009, approximately 62% of the electricity generated in Canada was from hydro

(about equal with 1990), while 16% was generated from nuclear fuel (about the same as 1990), 21% was from fossil fuels (22% in 1990), and the remainder was generated from renewables such as wind and biomass. Switching to fossil-fuel-generated electricity (the only source of electricity resulting in direct GHG emissions) will increase emissions, while the opposite holds true if other sources increase in proportion. Similarly, changes in the type of fuel being consumed (e.g. natural gas versus coal-fired generation) can increase or decrease emissions.

The impacts of these different drivers on electricity-related emissions in 2009 (compared to 1990 and 2005 base years) are shown in Figure 2–4 and Figure 2–3.

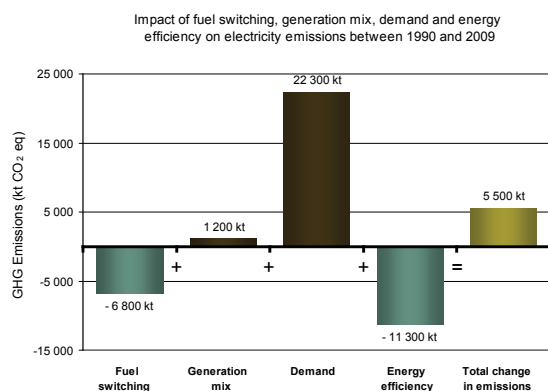
The trends illustrated by Figure 2-3 and Figure 2-4 can be summarized as follows:

**Fuel switching (combustion generation)** – between 1990 and 2009 the amount of electricity generated by natural-gas-fired units increased by almost 600%, while the amount generated by refined petroleum products (RPPs) decreased by over 60%. The switch from higher GHG-intense RPP fuels to natural gas has had the effect of lowering GHG emissions in 2009 as compared to 1990. Over the short term, the amount of fuel switching was less pronounced.

**Generation mix** – the generation mix refers to the shift between combustion and non-combustion (zero-GHG) sources to meet demand. Although hydro, nuclear and renewable generation increased over time, the proportion of

2 Although a number of provinces import some electricity from the United States, net imports only represent about 5% of the total amount of electricity generated in any given year. In all years since 1990, exports have been larger than imports, almost always by a considerable margin.

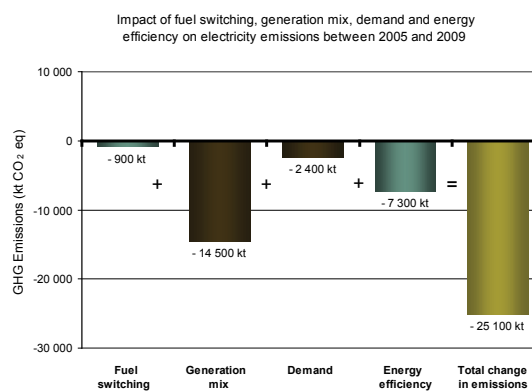
Figure 2–3 Impact of Factors on Change in Electricity Emissions, 1990–2009



Notes:

- 1) Fuel switching, generation mix, and demand and efficiency effects illustrate how these drivers have contributed to changes in emissions relative to a base year.
- 2) Emissions shown in the graph include those from electricity generation, but exclude SF<sub>6</sub> emissions from power transmission and distribution.

Figure 2–4 Impact of Factors on Change in Electricity Emissions, 2005–2009





electricity provided by combustion sources also increased compared to 1990. The increased level of non-combustion sources in 2009 is the biggest contributor to lower emissions compared to 2005.

**Demand** – the amount of electricity generated in 2009 was 33% higher than in 1990. This increase in demand by both the industrial and residential/commercial sectors is the main driver behind the overall net increase in emissions. The economic impact of the recession in Ontario reduced demand since 2005, although electricity consumption increased in Alberta and Saskatchewan over the same period.

**Energy efficiency** – improvements in energy efficiency, meaning the amount of energy required to generate electricity, helped reduce emissions compared to 1990 and 2009.

The main factor that explains the divergence between electricity demand and emissions from electricity generation is the generation source mix. In particular, in the mid-1990s, increased fossil fuel generation (mainly from coal plants) was used to support the growing demand for electricity, while nuclear and hydro-powered generation lagged. This consequently led to disproportionately higher increases in emissions relative to the early 1990s, when more nuclear generation capacity was available in Ontario. Since around 2003, although coal continued to be the fuel of choice in Alberta and Saskatchewan, Ontario initiated a program to shut down its coal-fired generators, while bringing a number of nuclear units back into service. In addition, precipitation since about 2004 was greater

than the 30-year average throughout many areas of the country and led to higher water levels and significantly greater hydro generation, while high oil prices also caused a significant shift from RPP fuels. In recent years, wind-generated electricity has begun also to have some impact on lowering emissions.<sup>3</sup> These events have all contributed to the decline in electricity industry emissions between 2003 and 2007.

The decrease in GHG emissions resulting from less electricity being generated from coal was further enhanced by continued fuel switching from higher- to lower-carbon fossil fuels and efficiency gains in fossil fuel-fired generators. In particular, the use of natural gas for electricity generation has increased significantly since 1990, and it now surpasses refined petroleum products (RPPs) in its contribution to total supply (natural gas is about half as carbon-intensive as coal, and approximately 25% lower than most RPPs). By 2009, the contribution of natural gas to the share of the generation mix was 6%—more than 7 times that of 1990. Aside from its environmental benefits, natural gas has also been price-competitive with oil. Natural gas electricity plants are now operating in most regions of the country, with Ontario and Alberta leading in gas-fired generation, followed by British Columbia and Saskatchewan. In Quebec and the Atlantic provinces, gas has been available only since 2000, but it is already being used in several new plants and a few retrofitted oil plants.

Emissions from electricity decreased significantly between 2008 and 2009 (Figure 2-2). The effects of the economic recession and mild weather in 2009 resulted in a drop in electricity consumption in Ontario, with demand reaching

### Renewable Incentives Programs

The federal government's Wind Power Production Incentive (WPPI), and ecoEnergy for Renewable Power, Ontario's Renewable Energy Standard Offer Program (RESOP) (or Ontario's Feed-In-Tariff program) are government incentives that have directly funded projects generating renewable power. These generators could be wind turbines, solar photovoltaic panels, tidal generators, biomass generators or geothermal generators. Wind power has enjoyed the most success of all of these types of renewable power. With over 4000 MW of capacity installed by 2010, wind energy now accounts for over 1% of Canada's total electricity generation. In total, the projects funded under all of these incentives have a generation capacity of over 6000 MW, which could be offsetting as much as 3.6 Mt CO<sub>2</sub> eq of GHG emissions annually.<sup>1</sup>

Two other federal initiatives that have encouraged renewable power generation are the Sustainable Development (SD) Technology Fund and the Pilot Emissions Removals, Reductions and Learning (PERRL) Initiative. The SD Technology Fund targeted many green prototype projects, providing funding to help them reach full-scale production. Several projects did achieve success and may be reducing GHG emissions by about 2–5Mt CO<sub>2</sub> eq annually.<sup>1</sup> The PERRL Initiative funded 13 projects in Canada including landfill gas capture and combustion technology, and wind power technology. These projects may be saving an additional 1 Mt CO<sub>2</sub> eq or more per year.

1. Environment Canada, *Complete Analysis of Notable Climate Change Incentives in Canada*. Internal document, Pollutant Inventories and Reporting Division, Environment Canada. Mar 2, 2011.



Table 2-3 GHG Emissions from Petroleum Refining and Manufacture of Solid Fuels and Other Energy Industries, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)							Change (%)
	1990	2004	2005	2006	2007	2008	2009	1990–2009
Petroleum Refining	17.7	20.2	18.8	18.6	21.2	20.3	19.9	13%
Manufacture of Solid Fuels and Other Energy Industries	33.7	53.2	47.5	47.9	49.4	48.8	43.7	29%
Total	51	73	66	66	71	69	64	24%

Note: Totals may not add up due to rounding.

139 TWh, the lowest level since 1997 (IESO 2010). As a result, coal fired generation (and emissions) in the province declined to levels well below those observed over the entire 1990–2009 time period and are the biggest contributing factor to the steep drop in emissions observed in 2009.

For more information on electricity generation and trends, see Annex 13 – Electricity Intensity Tables.

### **Petroleum Refining and Manufacture of Solid Fuels and Other Energy Industries<sup>4</sup>** (2009 GHG emissions, 64 Mt)

The Petroleum Refining category mainly includes emissions from the combustion of fossil fuels during the production of refined petroleum products (RPPs), whereas the Manufacture of Solid Fuels and Other Energy Industries category encompasses fuel combustion emissions associated with the upstream oil and gas (UOG) industry. Some combustion emissions from the upgrading of heavy oil and bitumen to produce synthetic crude oil are included in the Manufacture of Solid Fuels and Other Energy Industries category, with the remainder of upgrading emissions reported in the Petroleum Refining category and Mining subsector. The Mining subsector also includes a large amount of emissions associated with oil and gas extraction, particularly oil sands extraction. As shown in Table 2-3, between 1990 and 2009, emissions from the Petroleum Refining and Manufacture of Solid Fuels and Other Energy Industries categories increased by about 12 Mt, or 24%. This growth is due to increases in natural gas and oil production, particularly crude bitumen and heavy crude oil, largely for export.

Production rates of fossil fuels in the Manufacture of Solid Fuels and Other Energy Industries category are greatly

influenced by both export and domestic market demands (in 1990 and 2009, net exports<sup>5</sup> of crude oil and natural gas equalled 23% and 43% of total production, respectively). Figure 2-5 illustrates the production of fossil fuels in Canada from 1990 to 2009. During that period, production of crude oil and natural gas increased by 67% and 48%, respectively. Conventional crude oil production increased by 5%, although (after increasing steadily up until 1998) production was relatively constant until 2004, and has been generally declining since that time (Statistics Canada 2010b). In contrast, bitumen and synthetic crude oil production from Canada's oil sands has increased by 294%, with most of the growth occurring from 1996 onward. (CAPP 2010a).

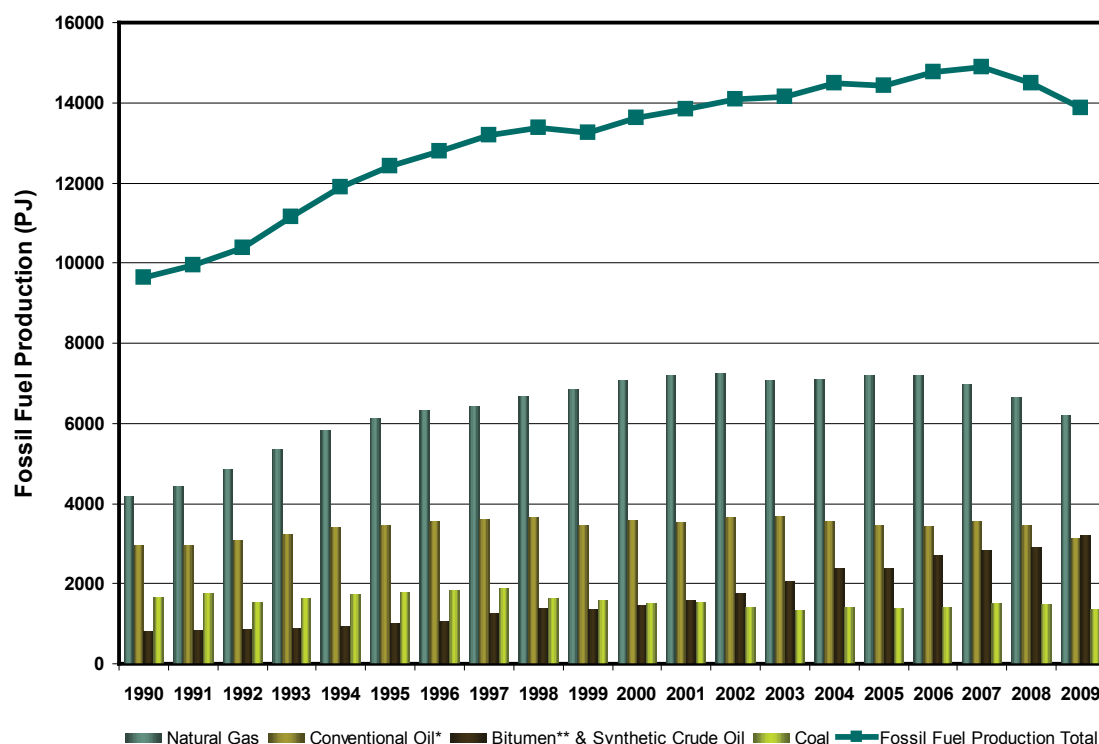
Increasingly, bitumen from the oil sands is being shipped to the United States where a much greater upgrading and refining capacity exists for heavier grades of oil (NEB 2006). This is supported by data from the Canadian Association of Petroleum Producers (CAPP), which show the ratio of bitumen to synthetic crude oil production in Canada rising by 9.0% between 2002 and 2009 (CAPP 2010a). As a result of this growing quantity of bitumen in the production mix, it appears that emissions associated with the upgrading and refining of bitumen were increasingly avoided in Canada.

Natural gas production increased rapidly from 1990 to peak production levels in 2002, a 73% increase over 1990 levels. However, from 2002 onwards, production decreased due to declines in output from the Alberta portion of the Western Sedimentary Basin, the largest gas-producing area in Canada (Nyboer and Tu 2010). Although total natural gas production started to decline after 2002, this was wholly from decreasing conventional natural gas

4 In the National Inventory Report (NIR), the Fossil Fuel Industries subsector encompasses both the Petroleum Refining and Upgrading category and the Fossil Fuel Production (also known as Manufacture of Solid Fuels and Other Energy Industries) category.

5 Although Canada exports significant volumes of oil and natural gas (mainly to the United States), it is also an importer of both crude oil and of refined oil products. This partially reflects historical events that helped ensure significant imports into Montréal and points east of the Ottawa Valley. Nonetheless, as a percentage of total production, the net export of crude oil is increasing.

Figure 2–5 Canadian Production of Fossil Fuels, 1990–2009



Notes:

\* Conventional oil includes light, medium and heavy oil production and pentanes plus and condensates.

\*\* Bitumen from oil sands operations.

production. In fact, unconventional natural gas production, including tight gas, coalbed methane and shale gas, has been increasing rapidly. In 2009, unconventional natural gas represented approximately 38% of total gas production, with tight gas, coalbed methane and shale gas accounting for 84%, 15% and 1% of total unconventional gas, respectively. In comparison, in 2002 unconventional gas accounted for 18% of total production with tight gas making up 99%. Since 2002 unconventional gas production has increased by 75%, while conventional gas has decreased by 35%.

Between 2002 and 2009, natural gas production decreased by approximately 14%, while crude oil production showed an overall growth of about 17%. This growth, which was accompanied by a 90% rise in the price of crude oil,<sup>6</sup> was almost wholly in the export market of oil sands products. While conventional crude oil exports decreased by 12%, exports of unconventional crude oil (crude bitumen and synthetic crude oil) increased by 109%. In addition, domestic crude consumption declined by 3.9% between 2002

and 2009. This decrease in domestic oil consumption was mainly the result of the economic downturn in 2008–2009 and affected not only GHG emissions from the fossil fuel industry, but emissions from other sectors as well.

### Manufacturing Industries, Construction and Mining (2009 GHG emissions, 74.9 Mt)

Emissions from the Manufacturing Industries, Construction and Mining subsector include the combustion of fossil fuels by the iron and steel; non-ferrous metals; chemicals; cement; pulp, paper and print; construction; mining; and all other manufacturing industries.<sup>7</sup> In 2009, GHG emissions were 74.9 Mt (Table 2-4). Overall, this subsector was responsible for 11% of Canada's total GHG emissions in 2009, up 10.4 Mt from 1990.

Between 1990 and 2009, there were changes in both directions in the emissions produced by the various categories within the Manufacturing Industries and Construction and Mining subsector. The majority of

6 Prices (Canadian dollars) rose from an average of about \$32 a barrel in 2002 to \$60 in 2009 (CAPP 2010b).

7 The NIR categories that constitute this UNFCCC sector are Manufacturing Industries, Construction, and Mining (refer to tables S-2 and S-3).

Table 2-4 GHG Emissions from Manufacturing, Construction, and Mining, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)							Change (%)
	1990	2004	2005	2006	2007	2008	2009	1990–2009
Iron and Steel	5.27	5.72	5.77	5.23	5.93	4.75	4.03	-24%
Non-ferrous Metals	3.26	3.26	3.29	3.25	3.51	3.68	3.12	-4.2%
Chemicals	8.22	7.89	7.04	8.14	8.57	7.27	7.57	-8.0%
Cement	3.82	4.76	5.02	5.20	4.84	4.63	3.61	-5.4%
Construction	1.87	1.34	1.36	1.30	1.29	1.26	1.08	-42%
Mining	6.65	17.78	18.60	20.50	27.20	27.62	31.30	371%
Pulp, Paper and Print	14.4	10.3	7.9	6.7	6.7	5.2	4.5	-69%
Other Manufacturing	21.0	21.2	19.8	20.1	21.4	18.2	19.7	-6.1%
Total	64.5	72.3	68.8	70.4	79.4	72.6	74.9	16%

Note: Totals may not add up due to rounding.

the overall increase can be attributed to the Mining category. This category includes oil sands mining and heavy oil and bitumen upgrading, as well as the mining of metals and minerals. Strong demand both domestically and internationally drove the long-term growth of GHG emissions in the Mining category, which has seen a 371% increase in emissions since 1990 (mainly as a result of increases in oil sands mining and heavy oil and bitumen upgrading). The remaining categories have all shown long-term decreases, from a high of 69% in the Pulp, Paper and Print category to 4.2% in the Non-ferrous Metals category. These decreases can be attributed to decreased demand due to effects from the 2008–2009 recession, fuel switching and changes in manufacturing operations.

Between 2005 and 2009, notable decreases in GHG emissions occurred in the Pulp, Paper and Print (43%), Iron and Steel (30%), and Cement (28%) categories. These decreases reflect the ongoing economic difficulties being felt by these sectors during the global recession and pressure from increased international competition.

### Transport (2009 GHG emissions, 190 Mt)

Transport is a large and diverse subsector, accounting for 28% of Canada's GHG emissions in 2009. This subsector includes emissions from fuel combustion for the transport of passengers and freight in five distinct categories:

- Road Transportation
- Civil Aviation (Domestic Aviation)
- Navigation (Domestic Marine)

- Railways
- Other Transportation (Off-road and Pipelines)

From 1990 to 2009, GHG emissions from transport, driven primarily by energy used for personal transportation, rose 30%, or 44 Mt. Overall, transport was the second-largest emission-producing category in 2009, contributing 190 Mt and accounting for 44% of Canada's emission growth from 1990 to 2009.

Emissions from light-duty gasoline trucks (LDGTs), the subcategory that includes sport utility vehicles (SUVs), pickups and minivans, increased 104% between 1990 and 2009 (from 20.3 Mt in 1990 to 41.3 Mt in 2009), while emissions from cars (light-duty gasoline vehicles or LDGVs) decreased 9% (from 45.5 Mt in 1990 to 41.4 Mt in 2009) (Table 2-5).

As shown in Table 2-5, the growth in road transport emissions is due not only to the 44% increase in the total vehicle fleet, but also to a shift in light-duty vehicle purchases from cars (LDGVs) to trucks (LDGTs), which, on average, emit 44% more GHGs per kilometre.

Over the 1990–2009 period, the increase of 21 Mt and 18 Mt for LDGTs and heavy-duty diesel vehicles (HDDVs), respectively, reflects the trend towards the increasing use of SUVs, minivans and pickups for personal transportation and heavy-duty trucks for freight transport (Table 2-6).

In 2009, emissions from HDDVs contributed 38 Mt to Canada's total GHG emissions (an increase of 91% from 1990 emissions). Emissions from heavy-duty gasoline vehicles (HDGVs) were lower for 2009, at 7 Mt; this figure represents a decrease of 6% over the 1990 level. While there are

Table 2–5 GHG Emissions from Transport, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)						
	1990	2004	2005	2006	2007	2008	2009
Transport (Total)	146	189	193	192	197	196	190
Civil Aviation (Domestic Aviation)	7.2	7.7	7.7	7.8	7.8	7.8	7.2
Light-duty Gasoline Vehicles	45.5	41.2	40.0	40.0	40.2	39.7	41.4
Light-duty Gasoline Trucks	20.3	41.4	42.5	42.7	43.0	42.6	41.3
Heavy-duty Gasoline Vehicles	7.44	6.62	6.54	6.65	6.81	6.84	6.99
Motorcycles	0.152	0.247	0.254	0.258	0.263	0.264	0.245
Light-duty Diesel Vehicles	0.469	0.569	0.574	0.580	0.617	0.652	0.663
Light-duty Diesel Trucks	0.702	1.830	1.930	1.960	2.020	2.020	1.940
Heavy-duty Diesel Vehicles	20.0	36.0	37.6	38.5	39.6	39.2	38.2
Propane & Natural Gas Vehicles	2.20	0.86	0.72	0.79	0.83	0.88	0.78
Railways	7	6	6	6	7	7	7
Navigation (Domestic Marine)	5.0	6.6	6.4	5.8	6.1	5.9	5.1
Off-road Gasoline	7.8	9.0	8.3	7.6	8.1	7.4	7.6
Off-road Diesel	16	23	24	24	26	28	26
Pipelines	6.85	8.47	10.1	9.61	8.94	7.46	6.32

Note: For full details on all years, please refer to Annex 12.

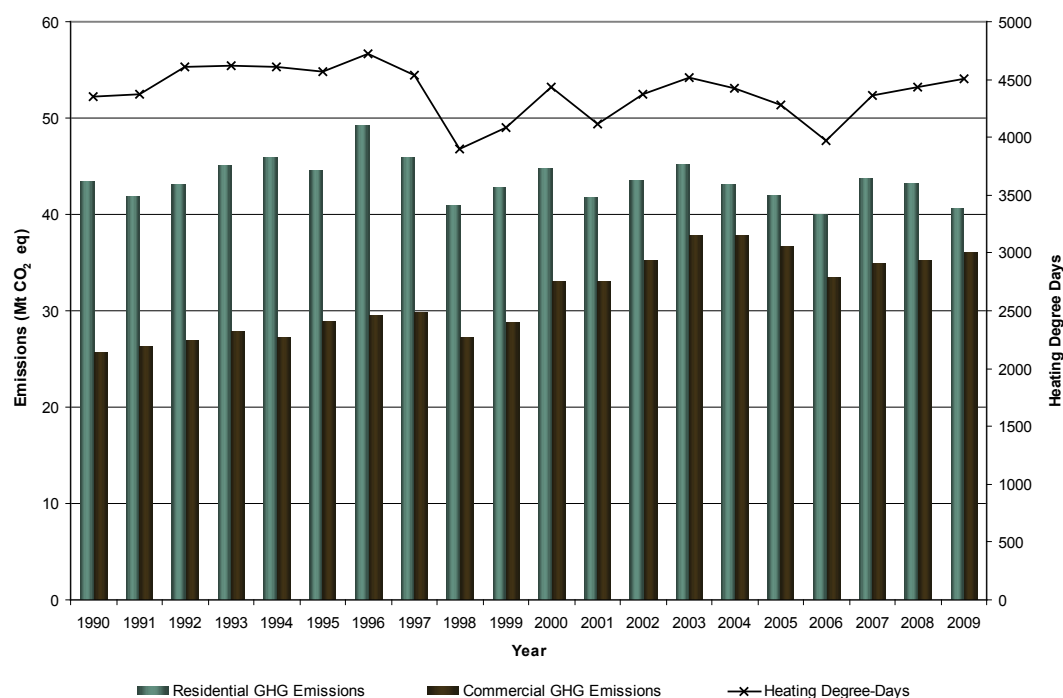
Table 2–6 Trends in Vehicle Populations for Canada, 1990–2009

Year	Number of Vehicles (000s)							Total
	LDGVs	LDGTs	HDGVs	MCs	LDDVs	LDDTs	HDDVs	
1990	10 646	3 308	518	261	109	112	402	15 356
1991	10 677	3 496	463	255	110	117	394	15 512
1992	10 674	3 712	432	248	109	126	397	15 698
1993	10 761	4 019	425	247	111	145	442	16 150
1994	10 694	4 305	428	234	108	165	487	16 421
1995	10 590	4 395	387	226	104	183	513	16 398
1996	10 273	4 517	383	213	99	174	498	16 157
1997	10 420	4 939	388	225	101	188	537	16 798
1998	10 250	5 347	395	263	107	204	629	17 195
1999	10 696	5 787	349	257	114	205	616	18 024
2000	10 863	6 065	376	288	123	224	649	18 588
2001	10 969	6 266	407	327	131	231	713	19 044
2002	10 929	6 421	394	359	138	234	724	19 199
2003	10 940	6 688	410	390	142	243	742	19 555
2004	10 931	6 959	429	417	153	254	801	19 944
2005	10 961	7 386	435	437	159	277	856	20 511
2006	11 195	7 551	445	446	162	284	876	20 959
2007	11 429	7 715	455	456	166	291	897	21 409
2008	11 663	7 879	465	465	170	298	918	21 858
2009	12 204	7 687	477	436	163	288	887	22 142

Notes:

HDDVs = Heavy-duty Diesel Vehicles; HDGVs = Heavy-duty Gasoline Vehicles; LDDTs = Light-duty Diesel Trucks; LDDVs = Light-duty Diesel Vehicles; LDGTs = Light-duty Gasoline Trucks; LDGVs = Light-duty Gasoline Vehicles; MCs = Motorcycles.

Figure 2-6 GHG Emissions and Heating Degree Days (HDDs) from Residential and Commercial Categories, 1990–2009



difficulties in obtaining accurate and complete data for the freight transport mode, the trends in data from major for-hire truck haulers in Canada show conclusively that freight hauling by truck has increased substantially and that this activity is the primary task performed by HDGVs and HDDVs.

Off-road fuel combustion emissions<sup>8</sup> in the Other Transportation subsector increased by 41% between 1990 and 2009, when the contribution from pipelines is not included.

The pipeline emissions included in the Other Transportation subsector are combustion emissions primarily from natural gas transport. Emissions have been steadily decreasing since 2005 most likely due to less demand for natural gas as newer, more efficient heating products are installed in residential and commercial buildings.

### Other Sectors (2009 GHG emissions, 78.7 Mt)

The Other Sectors subsector comprises fuel combustion

<sup>8</sup> Off-road emissions include those from the combustion of diesel and gasoline in a variety of widely divergent activities. Examples include the use of heavy mobile equipment in the construction, mining and logging industries, recreational vehicles such as snowmobiles and all-terrain vehicles (ATVs) and residential equipment such as lawnmowers and trimmers.

emissions from the Residential and Commercial categories, as well as stationary fuel combustion emissions from the Agriculture and Forestry category.<sup>9</sup> Overall, this subsector exhibited increases in GHG emissions of 10% from 1990 to 2009, while individual categories within it demonstrated a variety of changes.

### Residential and Commercial

Emissions in these categories arise primarily from the combustion of fuel to heat residential and commercial buildings. Fuel combustion in the Residential and Commercial and Institutional categories<sup>10</sup> accounted for 5.9% (41 Mt) and 5.2% (36 Mt), respectively, of all GHG emissions in 2009.

As shown in Figure 2-6, residential emissions fluctuate on an annual basis and overall, have decreased by 3 Mt between 1990 and 2009. Over the short term, emissions

<sup>9</sup> The UNFCCC Other Sectors subsector comprises the following NIR categories: Residential; Commercial and Institutional; and Agriculture and Forestry (listed under Energy, Stationary Combustion Sources in Annex 12).

<sup>10</sup> Commercial category emissions are based on fuel use as reported in the *Report on Energy Supply-Demand in Canada (RES-D)* (Statistics Canada #57-003) for the Commercial and Other Institutional, and Public Administration categories. The former is a catch-all category that includes fuel used by service industries related to mining, wholesale and retail trade, financial and business services, education, health and social services, and other industries that are not explicitly included elsewhere.

decreased by 2.5 Mt from 2008, 3.0 Mt from 2007 and increased 0.7 Mt from 2006. Commercial and Institutional emissions, however, increased 10.3 Mt or 40% between 1990 and 2009 and show a trend similar to that illustrated by the Residential category. Combined, the two categories exhibited an overall increase of 7.5 Mt or 11% between 1990 and 2009.

GHG emissions, particularly in the Residential category, track heating degree-days (HDDs)<sup>11</sup> closely (as shown in Figure 2-6). This close tracking indicates the important influence weather can have on space heating requirements and, therefore, on the demand for natural gas, home heating oil and biomass fuels.

There are several major factors that influenced the changes in energy-related GHG emissions in the Residential category. For example, the number of homes in Canada has increased by about 3 million since 1990 (NRCan 2010a), causing a 45% increase in floor space.<sup>12</sup> This increased floor space translates to an increase in GHG emissions of approximately 13 Mt between 1990 and 2009. In addition,

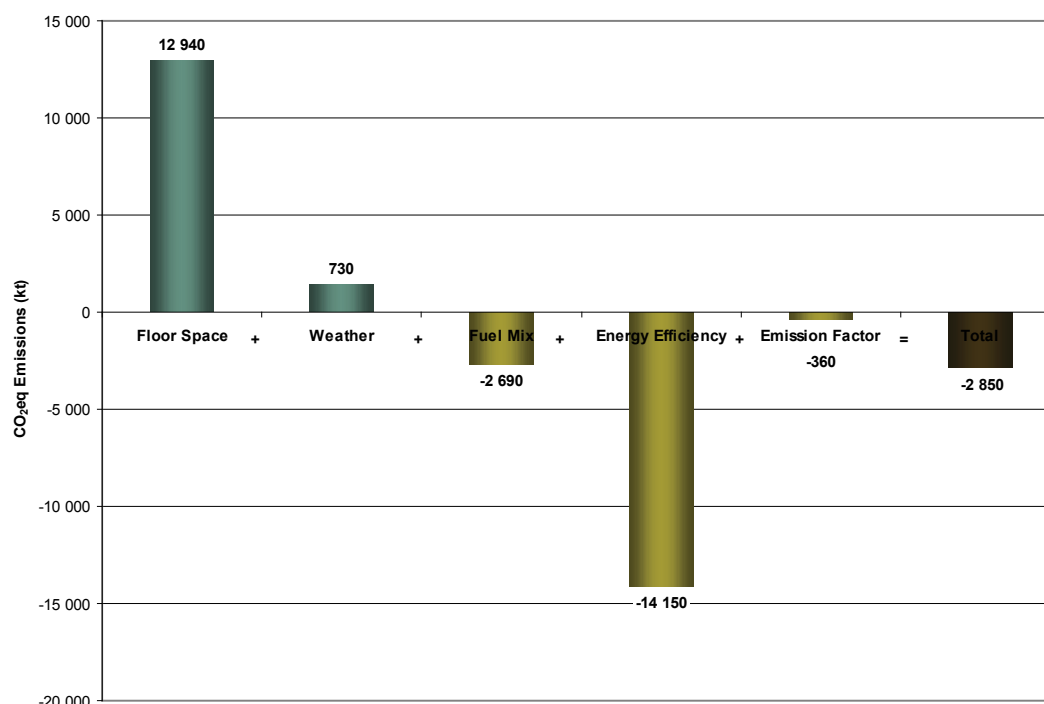
the climactic changes between 1990 and 2009 (as measured by heating degree days) caused an additional increase in GHG emissions of approximately 0.7 Mt. (See box, “Reducing Heating Requirements in Commercial and Residential Buildings,” for links between temperature and energy demand).

The GHG emission increases from increased floor space and climactic conditions have been offset by improvements made by owners and residents; specifically, changes in the fuel mix from heavy oil to natural gas (resulting in a 3-Mt reduction) and energy efficiency improvements (better construction methods, increased insulation and higher-efficiency heating systems). Energy-efficient new homes have been encouraged through programs such as the R-2000 initiative and residential home improvement incentive programs such as the EnerGuide for Homes (replaced by the ecoENERGY Retrofit program in 2007). These have been important because residential space heating requires the most energy of any end-use in Canadian homes, meaning that these changes led to significant reductions in GHG emissions. These improvements have resulted in a total emission reduction of about 14 Mt.

11 HDDs are calculated by determining the average cross-Canada number of days below 18.0°C and multiplying this value by the corresponding number of degrees below 18.0°C.

12 Data on number of homes and floor space were available to 2008 at time of publishing.

Figure 2-7 Major Influences on Stationary GHG Emissions from the Residential category between 1990 and 2009



The final influence on the overall energy-related GHG emissions were changes in residential emission factors, resulting in a decrease in GHG emissions of approximately

0.4 Mt between 1990 and 2009.<sup>13</sup>

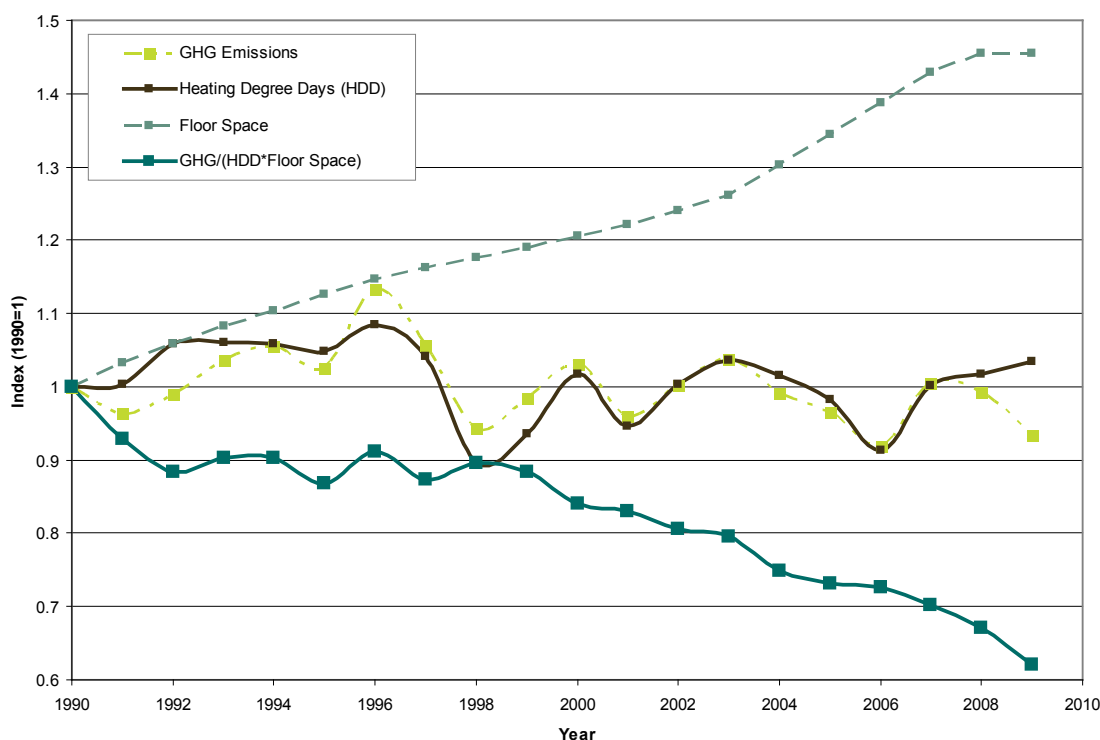
13 Since the emission factor effect is an aggregated factor on an energy basis, it is influenced by changes in emission factor or energy content over time. The change between 1990 and 2009 relates to variations in the natural gas emission factor by province, as well as variations in the emission factor and the energy content by type of coal and the variation in the fuelwood emission factor by appliance type. Energy content variations over time for light and heavy fuel oil and propane also have an influence.

### Reducing Heating Requirements in Commercial and Residential Buildings

The amount of energy required to heat and cool a dwelling is closely related to the outside ambient air temperature. Two common indicators that are used to determine the impacts of weather on energy requirements and GHG emissions are annual heating degree-days (HDDs) and annual cooling degree-days (CDDs). Annual HDDs are the annual sum of the days when the average daily temperature is below 18°C multiplied by the number of degrees the temperature is below 18°C on each of those days. Annual CDDs are the annual sum of days when the average daily temperature is over 18°C multiplied by the number of degrees above 18°C on each of those days. Since Canada is a northern country, home heating consumes a much greater amount of energy for the average home on an annual basis compared with other countries.

In general, there is a strong correlation between HDDs in Canada and the energy-related GHG emissions originating from the Residential category (see Figure 2-8). This indicates the close relationship between outside air temperatures and how much energy is required to heat the home. Another important relationship that can be seen is the decrease in GHG emissions per amount of floor space requiring heating (as indicated by the product of floor space and HDDs). This decoupling has been the result of increases in the efficiency of heating and the thermal envelope of buildings, as well as some changes in the mix of heating fuels.

Figure 2-8 Relationship between HDDs and Residential GHG Emissions, 1990–2009



Notes:

1. To remove the effect of climatic and population variables, the trend is measured in terms of GHG emissions divided by the product of floor space and HDDs over the period. The curve is shown indexed to 1990. If efficiency or fuel use patterns had not changed, the graph would likely have shown a horizontal line. However, the resultant GHG emission rate shows a declining trend between 1990 and 2009. This illustrates how building efficiency improvements and fuel switching produced GHG emission reductions.
2. Residential floor space data taken from NRCan (2010b)



## Energy Efficiency Programs for the Commercial and Residential Categories

Within the Residential and Commercial categories, the implementation of government incentives such as the *Energy Efficiency Act*, the ecoEnergy Retrofit and the ecoEnergy for Buildings and Houses program have coincided with noticeable GHG emission reductions due to efficiency improvements. Of these government actions, the most notable is the *Energy Efficiency Act*, as shown by the amount of GHGs that these programs have offset. These improvements have come from several areas including producing more efficient appliances, better insulation, and better construction techniques.

The Residential category is also a large consumer of electricity; therefore, efforts to increase efficiency in electricity use can have significant indirect impacts on reducing the requirements for electricity generation. The most significant of these changes have occurred with large appliances used in Canadian households. For example, thanks to programs like ENERGY STAR®, the energy efficiency of refrigerators has improved by up to 40% since 1990 (NRCan 2010a).

### Agriculture and Forestry

Stationary fuel combustion related emissions from the Agriculture and Forestry categories amounted to 2.1 Mt in 2009, a decrease of 14% from 1990. Emissions from these categories contributed less than 0.3% of the total for 2009.

#### 2.3.1.2. Fugitive Emissions from Fuels (2009 GHG emissions, 60.7 Mt)

As stated above, fugitive emissions from fossil fuels are the intentional or unintentional releases of GHGs from the production, processing, transmission, storage and delivery of fossil fuels. Released gases that are combusted before disposal (e.g. flaring of natural gases at oil and gas production and processing facilities) are also considered fugitive emissions. Fugitive emissions have two sources: coal mining and handling, and activities related to the oil and natural gas industry. They constituted 8.8% of Canada's total GHG emissions for 2009 and contributed 19% to the growth in emissions between 1990 and 2009.

Table 2-7 summarizes the changes in fugitive emissions from the Solid Fuels and the Oil and Natural Gas categories. In total, fugitive emissions grew by about 44% between 1990 and 2009, from 42.1 to 60.7 Mt, with emissions from the Oil and Natural Gas category contributing 99% of the total fugitive emissions in 2009, far overshadowing the 1% contribution from Coal Mining. Although fugitive releases from the Solid Fuels category (i.e. coal mining) decreased by 1.2 Mt (63%) between 1990 and 2009 as a result of the closing of many mines in eastern Canada, emissions

from oil and natural gas increased 49% during the same period.

This rise in emissions is a result of the increased production of natural gas and heavy oil (including crude bitumen) since 1990, largely due to increased worldwide demand for energy products. Since 1990, there has been a very large increase in the net energy exported from Canada (refer to section ES.4.1 of the Executive Summary for a discussion of emissions associated with the export of oil and natural gas), accompanied by a 185% increase in GHG emissions associated with those net energy exports.

Although overall fugitive emissions associated with oil and gas production have increased substantially since 1990, the overall fugitive emissions intensity (emissions per unit of energy produced) of upstream oil and gas production has decreased by 5.0% (see Table 2-7). This reduction is due to a decrease in oil sands fugitive emissions intensity of 4%, which was somewhat offset by a 12% increase in conventional oil production intensity. The increase in conventional oil intensity is indicative of the fact that easily removable reserves of conventional crude oil are being replaced with more high energy- and GHG-intensive sources, including heavier and/or more-difficult-to-obtain conventional oils such as those from offshore sources and enhanced oil recovery (EOR) operations.

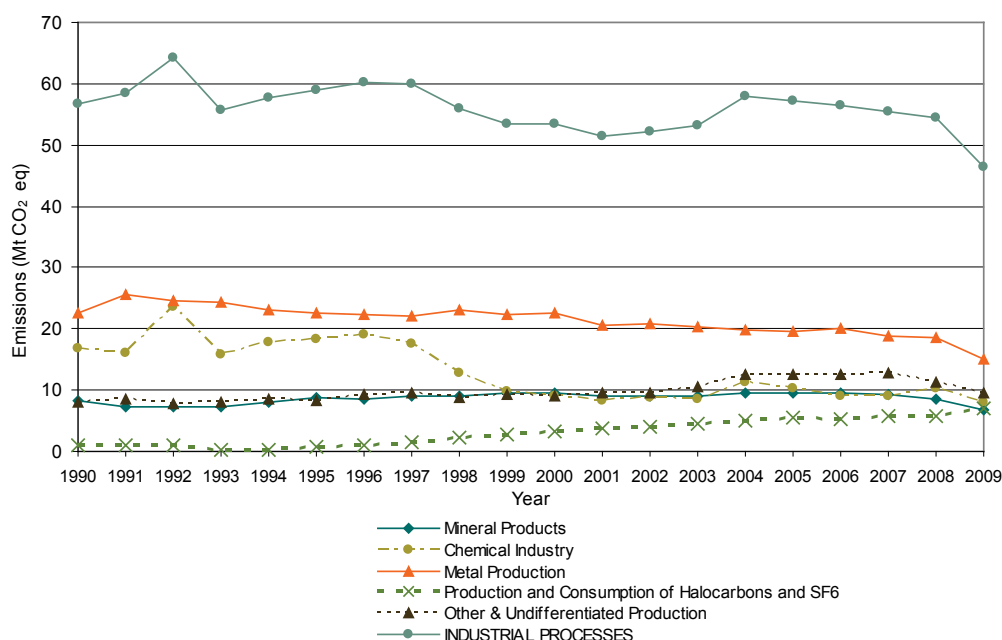
From 2008 to 2009, the emissions intensity of upstream oil and gas production has increased slightly from 4.48 to 4.53 kt CO<sub>2</sub> eq / PJ (1.1%). During this same time period, the emissions intensity of the oil sands industry also increased (~14.5%), mainly the result of a new integrated mining and upgrading facility as well as a new integrated *in-situ* bitumen extraction and upgrading facility. Both of these facilities ramped up production in 2009 and were not operating at peak efficiencies due to the required tuning of complex processes during the ramp-up to full-scale production.

Table 2–7 Fugitive GHG Emissions Intensity of Fossil Fuel Production by Category, Selected Years

	1990	2004	2005	2006	2007	2008	2009
<b>COAL PRODUCTION</b>							
Emissions (Mt CO <sub>2</sub> eq)	1.9	0.7	0.7	0.7	0.8	0.8	0.7
% Change since 1990	NA	-66%	-62%	-63%	-60%	-61%	-63%
Production (PJ)	1 673	1 416	1 401	1 419	1 506	1 490	1 361
% Change since 1990	NA	-15%	-16%	-15%	-10%	-11%	-19%
Intensity (kt CO <sub>2</sub> eq / PJ)	1.14	0.47	0.52	0.50	0.51	0.51	0.52
% Change since 1990	NA	-59%	-55%	-56%	-56%	-56%	-54%
<b>UPSTREAM OIL AND GAS PRODUCTION</b>							
Emissions (Mt CO <sub>2</sub> eq)	38.0	60.0	59.1	60.3	58.6	58.2	56.7
% Change since 1990	NA	58%	56%	59%	54%	53%	49%
Production (PJ)	7 969	13 056	13 017	13 343	13 388	12 991	12 518
% Change since 1990	NA	64%	63%	67%	68%	63%	57%
Intensity (kt CO <sub>2</sub> eq / PJ)	4.77	4.60	4.54	4.52	4.38	4.48	4.53
% Change since 1990	NA	-3.6%	-4.8%	-5.2%	-8.2%	-6.0%	-5.0%
<b>Conventional Oil Production</b>							
Emissions (Mt CO <sub>2</sub> eq)	16.2	24.3	23.3	23.3	22.1	20.8	18.9
% Change since 1990	NA	50%	44%	44%	37%	29%	17%
Production (PJ)	2 973	3 567	3 459	3 434	3 542	3 447	3 115
% Change since 1990	NA	20%	16%	16%	19%	16%	5%
Intensity (kt CO <sub>2</sub> eq / PJ)	5.44	6.82	6.73	6.79	6.23	6.03	6.07
% Change since 1990	NA	25%	24%	25%	15%	11%	12%
<b>Oil Sands Mining, Extraction and Upgrading</b>							
Emissions (Mt CO <sub>2</sub> eq)	2.5	6.3	5.9	6.6	7.0	7.3	9.2
% Change since 1990	NA	155%	141%	169%	187%	197%	277%
Production (PJ)	813	2 371	2 375	2 712	2 852	2 892	3 203
% Change since 1990	NA	192%	192%	234%	251%	256%	294%
Intensity (kt CO <sub>2</sub> eq / PJ)	3.02	2.64	2.48	2.43	2.46	2.52	2.88
% Change since 1990	NA	-12%	-18%	-19%	-18%	-17%	-4%
<b>Natural Gas Production and Processing</b>							
Emissions (Mt CO <sub>2</sub> eq)	15.1	23.8	24.3	24.7	23.8	24.4	22.8
% Change since 1990	NA	58%	61%	63%	58%	62%	51%
Production (PJ)	4 184	7 118	7 183	7 196	6 994	6 652	6 200
% Change since 1990	NA	70%	72%	72%	67%	59%	48%
Intensity (kt CO <sub>2</sub> eq / PJ)	3.61	3.34	3.38	3.43	3.40	3.67	3.67
% Change since 1990	NA	-7.4%	-6.4%	-5.0%	-5.8%	1.8%	1.9%
<b>Natural Gas Transmission</b>							
Emissions (Mt CO <sub>2</sub> eq)	4.3	5.7	5.7	5.7	5.8	5.7	5.8
% Change since 1990	NA	33%	32%	33%	34%	34%	35%
Pipeline Length (km)	64 222	83 619	83 245	83 865	84 362	84 077	84 731
% Change since 1990	NA	30%	30%	31%	31%	31%	32%
Intensity (kt CO <sub>2</sub> eq/km)	0.067	0.068	0.068	0.068	0.068	0.068	0.068
% Change since 1990	NA	1.8%	1.7%	1.9%	2.0%	2.1%	2.2%
<b>DOWNSTREAM PRODUCTION</b>							
Emissions (Mt CO <sub>2</sub> eq)	2.1	3.3	3.3	3.3	3.4	3.3	3.2
% Change since 1990	NA	56%	53%	53%	60%	54%	52%
Production (PJ)	3 907	4 848	4 755	4 710	4 813	4 638	4 539
% Change since 1990	NA	24%	22%	21%	23%	19%	16%
Intensity (kt CO <sub>2</sub> eq / PJ)	0.55	0.68	0.69	0.69	0.71	0.71	0.71
% Change since 1990	NA	25.3%	26.0%	27.3%	29.8%	30.0%	30.6%
<b>Petroleum Refining</b>							
Emissions (Mt CO <sub>2</sub> eq)	0.9	1.5	1.5	1.4	1.5	1.4	1.3
% Change since 1990	NA	78%	69%	65%	75%	60%	51%
Production (PJ)	3 907	4 848	4 755	4 710	4 813	4 638	4 539
% Change since 1990	NA	24%	22%	21%	23%	19%	16%
Intensity (kt CO <sub>2</sub> eq / PJ)	0.22	0.32	0.31	0.30	0.31	0.30	0.29
% Change since 1990	NA	44%	39%	36%	42%	35%	30%
<b>Natural Gas Distribution</b>							
Emissions (Mt CO <sub>2</sub> eq)	1.3	1.8	1.8	1.9	1.9	1.9	1.9
% Change since 1990	NA	40%	43%	46%	49%	51%	52%
Pipeline Length (km)	168 813	236 404	241 344	246 317	252 371	254 512	256 634
% Change since 1990	NA	40%	43%	46%	49%	51%	52%
Intensity (kt CO <sub>2</sub> eq / km)	0.008	0.008	0.008	0.008	0.008	0.008	0.008
% Change since 1990	NA	0.0%	0.0%	0.0%	0.0%	0.0%	0.0%

Notes: NA = Not applicable.

Figure 2–9 GHG Emissions from Industrial Processes by Subsector, 1990–2009



## 2.3.2. Industrial Processes Sector

### (2009 GHG emissions, 46.3 Mt)

The Industrial Processes Sector includes GHG emissions that are direct by-products of processes, including Mineral Products, Chemical Industry, Metal Production, Production and Consumption of Halocarbons and SF<sub>6</sub>, and Other and Undifferentiated Production. GHG emissions from the Industrial Processes Sector contributed 46.3 Mt to the 2009 national GHG inventory, compared with 56.8 Mt in 1990. Figure 2-9 illustrates the changes in each of the subsectors over the period 1990–2009, and Table 2-8 provides an emission breakdown by category for selected years.

Between 1990 and 2009, the overall Sector emissions decreased by approximately 10.2 Mt (18.4%). This change could be explained by significant emission reductions in adipic acid production (N<sub>2</sub>O), aluminium production (PFCs), magnesium production (SF<sub>6</sub>), and iron and steel production (CO<sub>2</sub>), which were only partly offset by growths in emissions from consumption of HFCs, other and undifferentiated production (CO<sub>2</sub>)<sup>14</sup>, aluminium production (CO<sub>2</sub>) and ammonia production (CO<sub>2</sub>).

14 Other and Undifferentiated Production is an emission subsector composed mainly of petrochemical production that uses hydrocarbons as feedstock. Also, included here is a portion of CO<sub>2</sub> emissions from fuels used as reductants (other than coke) in iron and steel production, and in base-metal smelting.

### 2.3.2.1. Mineral Products

Between 1990 and 2009, the decrease in emissions from the category of Cement Production of 6.0% (0.33 Mt CO<sub>2</sub> eq) could be explained by the decrease in clinker production<sup>15</sup> (–6.0%) which in turn was associated with the drop in the domestic demand for cement products (–7.7%) (Statistics Canada 1991, 2009a). This trend is, of course, atypical for the category and is due to the recessionary year of 2009. In fact, the 1990–2008 trend shows an increase of 22% in emissions. The economic downturn and considerable decreases in construction activities in the United States<sup>16</sup> and in Canada,<sup>17</sup> which started in 2008 and continued in 2009, were responsible for the overall decrease in cement and clinker production seen in 2009. During the period 2004–2009, production of clinker hit its lowest point in 2009 (9.9 Mt in 2009 vs 14 Mt in 2004). Similar to the long-term trend, the downward short-term

15 Clinker is an intermediate component in the cement manufacturing process. During the production of clinker, limestone, which is mainly (95%) calcium carbonate (CaCO<sub>3</sub>), is first heated (calcined) to produce lime (CaO). The CaO then reacts with silica, aluminium and iron oxides in the raw materials to make the clinker minerals that are dominantly hydraulic calcium silicates (IPCC 2000).

16 Data on residential construction activities in the U.S. can be found at: <http://www.census.gov/const/www/newresconstindex.html>

17 Data on residential and non-residential building permits were provided in a presentation by Statistics Canada (Manufacturing and Energy Division) on August 25, 2010.

Table 2–8 GHG Emissions from Industrial Processes by Category, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Total - Industrial Processes	56.8	53.5	58.1	57.2	56.5	55.5	54.5	46.3
Mineral Products	8.3	9.6	9.5	9.5	9.6	9.3	8.6	6.8
Cement Production	5.4	6.7	7.1	7.2	7.3	7.3	6.6	5.1
Lime Production	1.8	1.9	1.8	1.7	1.6	1.6	1.5	1.2
Limestone and Dolomite Use	0.73	0.62	0.24	0.25	0.31	0.16	0.29	0.28
Soda Ash Use	0.21	0.22	0.15	0.16	0.17	0.17	0.15	0.10
Magnesite Use	0.147	0.181	0.184	0.175	0.149	0.068	0.057	0.069
Chemical Industry	17	9.0	11	10	9.1	8.9	10	8.1
Ammonia Production	5.0	6.8	6.8	6.3	6.6	6.2	6.7	6.2
Nitric Acid Production	1.01	1.23	1.23	1.25	1.23	1.13	1.23	1.15
Adipic Acid Production	11	0.9	3.1	2.6	1.2	1.5	2.4	0.7
Petrochemical Production	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Metal Production	22.6	22.5	19.9	19.6	20.1	18.8	18.5	15.0
Iron and Steel Production	10.2	11.5	10.4	10.1	11.0	11.0	10.6	7.65
Aluminium Production	9.3	8.2	7.3	8.2	7.7	7.3	7.4	7.2
Magnesium Production	2.87	2.31	2.00	1.09	1.20	0.325	0.181	0.000
Magnesium Casting	0.236	0.471	0.191	0.201	0.190	0.198	0.280	0.193
Production and Consumption of Halocarbons	0.77	3.0	4.7	5.2	5.0	5.4	5.5	6.8
SF <sub>6</sub> Use in Electric Utilities and Semiconductors	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Other and Undifferentiated Production	8.0	9.2	13	12	13	13	11	9.4

Note: Totals may not add up due to rounding.

trend was caused by the decrease in domestic and foreign demands.<sup>18</sup>

There was also a decrease in process CO<sub>2</sub> emissions of 31% (0.55 Mt CO<sub>2</sub> eq) for the category of Lime Production between 1990 and 2009. During that period, the national calcining capacity decreased by 20%, with most of the plant downsizing or closures happening in the province of Ontario (NRCan 1990; personal communication<sup>19</sup>). Moreover, according to the Canadian Lime Institute, lime output in Canada was down in 2008–2009 due to a general market downturn associated with the global financial crisis and particularly a downturn in the steel market. Also, more lime plant closures in the U.S. than in Canada may explain the increase in lime exports from Canada during the 2004–2008 period.<sup>20</sup> However, the increase in exports

(+0.2 Mt between 2004 and 2009)<sup>21</sup> did not suffice to compensate for the falling domestic needs driven by steel and pulp and paper industries.

The category of Mineral Product Use (which is made up of uses of limestone and dolomite, soda ash and magnesite) experienced a decrease in emissions of 0.64 Mt CO<sub>2</sub> eq (59%) from 1990 to 2009. Pulp and paper mills, which are consumers of limestone and soda ash, have faced challenges since the mid-1990s. Decline in newsprint demand and foreign competition have caused some plant closures (NRCan 2011). The decrease in pulp and paper manufacturing was partly responsible for decreases in the use of certain minerals and their resulting emissions. Moreover, limestone and soda ash, along with silica, are raw materials for making glass. The increase in use of recycled glass (cullet) has reduced the need for virgin raw material in the glass batch (NRCan 2007) from which CO<sub>2</sub> emissions could be generated. Since 2008, the decline in iron and steel production, a user of mineral products, further contributed to the decrease of emissions from Mineral Product Use. The closure of Norks Hydro in 2007 accounted for the majority of the decrease in magnesite use and the associated emissions.

18 Downing T. Personal communication (email from Tim Downing to Paula Charbonneau, Pollutant Inventories and Reporting Division, dated February 16, 2011). Cement Association of Canada.

19 Panagapko D. Personal communication (email from Doug Panagapko to Shanta Chakrovorty, Pollutant Inventories and Reporting Division, dated August 17, 2010). Natural Resources Canada.

20 Kenefick W. Personal communication (emails from Wayne Kenefick to Paula Charbonneau, Pollutant Inventories and Reporting Division, dated October 26 and 29, 2010). Canadian Lime Institute.

21 Data on export of lime (commodity code 252210) were obtained from [http://cansim2.statcan.gc.ca/cgi-win/CNSMCGI.PGM?Lang=E&CMT\\_Action=Sections&ResultTemplate=CII\\_CIMT5](http://cansim2.statcan.gc.ca/cgi-win/CNSMCGI.PGM?Lang=E&CMT_Action=Sections&ResultTemplate=CII_CIMT5)

### 2.3.2.2. Chemical Industry

Between 1990 and 2009, a significant emission reduction of 10.1 Mt CO<sub>2</sub> eq (94%) was noted in the Adipic Acid Production category. This resulted from the installation of an emission abatement system at Canada's only adipic acid facility in 1997. According to the plant's environmental manager, since the abatement unit was implemented, emissions were to a large degree dependent on the efficiency of the abatement system and the site's success in maximizing the uptime of the system.<sup>22</sup> The plant has, however, become indefinitely idled as of the spring of 2009.

Over the long term (1990–2009), production of ammonia increased by 0.78 Mt (24%), resulting in an emission growth of 24% (1.2 Mt CO<sub>2</sub> eq). Because ammonia is mainly used as fertilizer or in the production of fertilizers, the rise in ammonia production could be attributed to the increase in agricultural activities (of approximately 36%, in terms of gross domestic product [GDP] value).<sup>23</sup> During the period of 2004–2009, the production of ammonia has been relatively stable, with inter-annual changes of around 5–8%.

A decrease of 0.04 Mt CO<sub>2</sub> eq (42%) between 1990 and 2009 was observed for the category of Petrochemical Production. This long-term decreasing trend reflected closures of certain petrochemical (i.e. carbide, methanol and ethylene dichloride) facilities. According to Statistics Canada (2010c) CANSIM 304-0014, manufacturing sales for the subsector of Chemical Industry has dropped by 16% between 2005 and 2009. The chemical manufacturing sales value reached its lowest point in 2009 (amongst the years 2005–2009).

It is important to note that the sudden emission increase in 1992 (shown in Figure 2-9) for the chemical industry and for the Industrial Processes Sector as a whole were due to a file linkage error that will be corrected in the new inventory. The correct value for 1992 should be in similar range of magnitude covering the 1991 and 1993 values. For further explanations, please see Section 4.12.4 in Chapter 4.

### 2.3.2.3. Metal Production

Steel production steadily increased during the period 1990–2004 (+33%), after which production began to drop and reached its lowest point in 2009. In addition to foreign competition, the economic recession (which started in late 2008 and continued in 2009) had repercussions on the iron and steel industry in 2009. For example, the export of steel products (mainly to the U.S.) has significantly dropped, by 48%, in 2009 from its 2008 level. The sales to domestic consumers of steel products (including the motor vehicle industry) also fell by 39% (Statistics Canada 2008–2009). As a result, there were indefinite or temporary closures of some plants (e.g. ArcelorMittal in Contrecoeur, Quebec) in 2009. The emission reduction seen between 2004 and 2009 of 27% (2.8 Mt CO<sub>2</sub> eq) for the Iron and Steel Production category as a whole corresponded to decreases of 3.5 Mt (40%) observed in the national production levels for pig iron and of 7.0 Mt (43%) for steel (Statistics Canada 2004, 2008–2009) due to considerable decreases in domestic and foreign demands.

The aluminium industry has succeeded in bringing down its perfluorocarbon (PFC) emissions by 4.4 Mt CO<sub>2</sub> eq (67%), while increasing production by 94% (1.5 Mt), between 1990 and 2009. The PFC reductions have been achieved through incorporation of computerized sensors and automated alumina feeders, which have helped reduce the occurrence of anode effects. In addition, the provided data showed that the industry continued to increase its production in more modern plants (i.e. with prebaked technology), rather than in older plants (i.e. with Soderberg technology). However, the increase in aluminium production also gave rise to increases in CO<sub>2</sub> emissions of 2.3 Mt CO<sub>2</sub> eq (or 85%), since CO<sub>2</sub> comes from the reduction of alumina with carbon anodes, an essential reaction in the production that cannot be avoided. During the period 2004–2009, production of aluminium steadily increased and reached its highest level in 2008. However, there was a 3% drop in production between 2008 and 2009, which could be explained by changes at two facilities. The RioTintoAlcan's facility in Beauharnois, Quebec, ceased its smelting operations in spring 2009 but continued its casting activities.<sup>24</sup> One of the pot rooms at Alcoa's facility in Baie-Comeau, Quebec, also stopped producing

<sup>22</sup> Lauridsen S. 2007. Personal communication (email dated October 30, 2007). Invista Canada.

<sup>23</sup> Sonnen C. Personal communication (email from Carl Sonnen to Alice Au, Pollutant Inventories and Reporting Division, dated December 15, 2010). Informetrica Ltd.

<sup>24</sup> Information obtained from [http://www.riotintoalcan.com/ENG/media/media\\_releases\\_1798.asp](http://www.riotintoalcan.com/ENG/media/media_releases_1798.asp)



in late 2008,<sup>25</sup> as the facility has been undergoing a modernization process (i.e. to convert part of the production from Soderberg type to pre-baked type).<sup>26</sup>

Between 1990 and 2005, the category of magnesium production had also shown significant emission decreases because of progressive replacement of SF<sub>6</sub> with alternatives used as cover gas. For the years 2005–2008, gradual production reductions in anticipation of plant closures in Norsk Hydro and Timminco brought about considerable

decreases in SF<sub>6</sub> emissions. Norsk Hydro was finally shut down in the spring of 2007, and Timminco in August 2008.

The number of facilities involved in magnesium casting increased since 1990 (four) and reached the highest (nine) in 2000. This corresponded to the upward SF<sub>6</sub> emission trend seen between 1990 and 2000. For the years after 2000, the SF<sub>6</sub> emission level fell gradually because of the replacement of SF<sub>6</sub> with alternatives, such as SO<sub>2</sub> and Novec 612, and the decline in the magnesium casting business (i.e. closure of plants). In 2009, with the closures of two plants, one in June and another in December, there were only five magnesium casting plants operating in Canada.

#### 2.3.2.4. Production and Consumption of Halocarbons and SF<sub>6</sub>

There has been an emission growth of 6.3 Mt CO<sub>2</sub> eq (1300%) for consumption of HFCs since 1995. This could be explained by the fact that more ozone-depleting substances (ODSs) have been replaced by hydrofluorocarbons (HFCs) within the refrigeration and air conditioning (AC) markets since the Montreal Protocol came into effect in 1996. Although a 1990 value is shown for Production and Consumption of Halocarbons in Table 2-8, this value represents only HFC-23 emissions from the production of HCFC-22, because emissions from the consumption of halocarbons were considered negligible in 1990. The 2000 and 2004 to 2009 values in Table 2-8 account for emissions coming from the consumption of halocarbons (PFCs and HFCs) only, because the last HCFC-22-producing plant was shut down in 1993. It is also important to note that the 1999–2008 values will be corrected in the new inventory, as there was a file linkage error detected during the quality

control process that could not be corrected in this inventory. The transcription errors did not affect the overall long-term trend, but the short-term (i.e. 2004–2009) increasing emission trend should be less steep, given that the correct (2004–2008) values would be closer to the 2009 value.

#### 2.3.2.5. Other and Undifferentiated Production<sup>27</sup>

The increase in emissions of 1.4 Mt CO<sub>2</sub> eq (17%), compared to the 1990 level, for the Other and Undifferentiated Production subsector was partly due to the growth in non-energy use of butane (217%) and ethane (130%).<sup>28</sup> This could be due to the significant increase in ethylene production (% change cannot be divulged due to data confidentiality) between 1990 and 2009. However, for other types of petrochemical products, such as butadiene, butylene and propylene, the decrease in production started in 2005 and continued until 2009. The Other and Undifferentiated Production subsector also accounts for CO<sub>2</sub> emissions coming from use of reductants during the smelting of base metals. The growth of base metal production between 1990 and 2009 (of 23% in terms of GDP<sup>29</sup>) contributed as well to the 1990–2009 emission increase. However, with increased competition in export markets and the effect of the recession in 2009, the short-term production trend has been downward for the base-metal industry, resulting in lesser use of reductants and lower CO<sub>2</sub> emissions.

#### 2.3.3. Solvent and Other Product Use Sector (2009 GHG emissions, 0.26 Mt)

The Solvent and Other Product Use Sector accounts for emissions related to the use of N<sub>2</sub>O as an anaesthetic in medical applications and as a propellant in aerosol products. It contributed 260 kt CO<sub>2</sub> eq to the 2009 national GHG inventory, compared to 179 kt CO<sub>2</sub> eq in 1990. The emission trends were primarily driven by the domestic demand for N<sub>2</sub>O for anaesthetic or propellant purposes.

25 Dubois C. Personal communication (email from Christine Dubois to Etienne Renaud, Pollutant Inventories and Reporting Division, dated December 6, 2010). Alcoa.

26 Dubois C. Personal communication (email from Christine Dubois to Alice Au, Pollutant Inventories and Reporting Division, dated Feb 21, 2011). Alcoa.

27 Other and Undifferentiated Production is an emission subsector composed mainly of petrochemical production activities that use hydrocarbons as feedstock. Also, included here is a portion of CO<sub>2</sub> emissions from fuels used as reductants (other than coke) in base-metal smelting and iron and steel production.

28 Percentages come from a comparison between 1990 and 2009 data obtained from Statistics Canada's RESD (#57-003).

29 Personal communication (email from Carl Sonnen to Alice Au, Pollutant Inventories and Reporting Division, dated December 15, 2010). Informetrica Ltd.

## 2.3.4. Agriculture Sector (2009 GHG emissions, 56 Mt)

The main sectors in Canadian agriculture are livestock and crop production. The livestock sector is dominated by the beef and swine industries, while crop production is mainly dedicated to the production of cereals and oil seeds. Dairy and poultry production are controlled to meet national demand. A wide variety of specialty crops and animals are also produced, but represent a very small portion of the overall agricultural economy. The sectors are highly regionalized; approximately 70% of beef cattle and more than 90% of wheat, barley and canola are produced in the Prairies<sup>1</sup> in a semi-arid to sub-humid ecozone. On the other hand, approximately 70% of dairy cattle, 60% of swine and poultry, 95% of corn and 90% of soybeans are produced in the humid mixed-wood plains ecozone in eastern Canada. Traditionally Canada's Agriculture Sector has been composed of small family farms, but over the past 30 years, intensification has occurred in the Agricultural Sector and as a consequence, the number of farms has decreased and farm size and productivity have increased.

Emissions directly related to animal and crop production accounted for 56 Mt CO<sub>2</sub> eq or 8.1% of total 2009 GHG emissions for Canada, an increase of 9 Mt CO<sub>2</sub> eq or 20% since 1990. Agriculture accounted for 24% and 72% of the national CH<sub>4</sub> and N<sub>2</sub>O emissions, respectively. All these emissions are from non-energy sources; N<sub>2</sub>O accounted for 61% of estimated sectoral emissions in 2009 and CH<sub>4</sub> for 39%.

The processes and activities that produce GHG emissions in the Agriculture Sector are enteric fermentation (digestion by ruminant animals); the application of nitrogen fertilizer to agricultural soils; and manure management, storage and application to soils. These emissions can be attributed to either the livestock sector, which includes enteric fermentation emissions (CH<sub>4</sub>) and all manure-related emissions (CH<sub>4</sub> and N<sub>2</sub>O), or the crop production sector, which consists of N<sub>2</sub>O emissions from the application of mineral fertilizers, crop residue decomposition and other management practices (Table 2-9).

In 2009, livestock emissions consist of 19 Mt CO<sub>2</sub> eq from enteric fermentation and 15 Mt CO<sub>2</sub> eq from manure management and storage (57% and 43% of livestock emissions, respectively). Crop production produces N<sub>2</sub>O emissions (Table 2-9) during the application of synthetic nitrogen fertilizers (13 Mt CO<sub>2</sub> eq,) and from crop residue decomposition (8.2 Mt CO<sub>2</sub> eq), representing 61% and 38 % of crop production emissions, respectively.

A discussion of GHG trends in agricultural production must also take into account the complex interconnections between the two dominant branches of agriculture: livestock and crop production. These two sub-industries both compete for the same landbase and contribute resources to and from that landbase. Over the past decades, agriculture has undergone a continual process of intensification. In the crop production industry intensification has involved an increased reliance on off-farm inputs such as fertilizers, herbicides and pesticides and has resulted in very important increases in productivity per hectare. In the

Table 2-9 GHG Emissions from Agriculture by Production Systems for Selected Years<sup>1</sup>

Production System	GHG Emissions (Mt CO <sub>2</sub> eq)							
	1990	1996	2001	2005	2006	2007	2008	2009
Livestock	29	34	36	39	38	37	36	34
Dairy Cows	5.9	5.6	4.9	4.7	4.7	4.5	4.5	4.5
Beef Cattle	19	24	26	28	27	27	26	25
Swine	2.4	2.7	3.2	3.5	3.4	3.3	2.9	2.7
Other Livestock <sup>2</sup>	1.5	1.7	2.2	2.2	2.2	2.2	2.2	2.2
Crop	18	20	18	19	19	21	23	22
Synthetic Nitrogen Fertilizers	9.2	11	12	11	11	13	13	13
Crop Residue Decomposition	6.9	7.0	5.5	7.6	7.9	7.6	9.1	8.2
Other Management Practices <sup>3</sup>	2.0	1.6	0.9	0.6	0.5	0.4	0.2	0.1
Agriculture (Total)	47	54	54	58	57	57	58	56

1. Totals may not add up due to rounding.

2. Other livestock includes sheep, lamb, goat, horse, bison, poultry, llamas and alpacas.

3. Other management practices includes summerfallow, conservation tillage practices, irrigation, cultivation of organic soils and field burning of crop residues.



livestock industry this has also involved increased reliance on processed feeds and medicinal and non-medicinal supplements that have also increased output per animal. At the same time, over the past 30 years there has been an increased focus on soil conservation through conservation tillage and crop rotation as well as increasing efficiency through the reduction of nutrient losses to the environment through improved fertilizer application to crops and more precise feeding techniques for animals. For these reasons, a comprehensive discussion of trends in emissions from agricultural production must at least touch on the dominant emissions from production practices, farm inputs and land management practices and land-use change.

The main drivers of the emission trend in the Agriculture Sector are the expansion of the beef cattle and swine populations, and increases in the application of synthetic nitrogen fertilizers in the Prairies. Beef, swine and poultry populations in Canada are 23%, 19% and 31% higher, respectively, than in 1990. The significant growth in animal populations largely accounts for the 19% increase in emissions, from 29 to 34 Mt CO<sub>2</sub> eq in emissions associated with animal production over the 1990–2009 period (Table 2-9). In the case of beef cattle, emissions increased at greater rates than cattle populations as herd improvements resulted in an increase in live weight; consequently, an average animal now consumes more feed and also emits more GHGs.

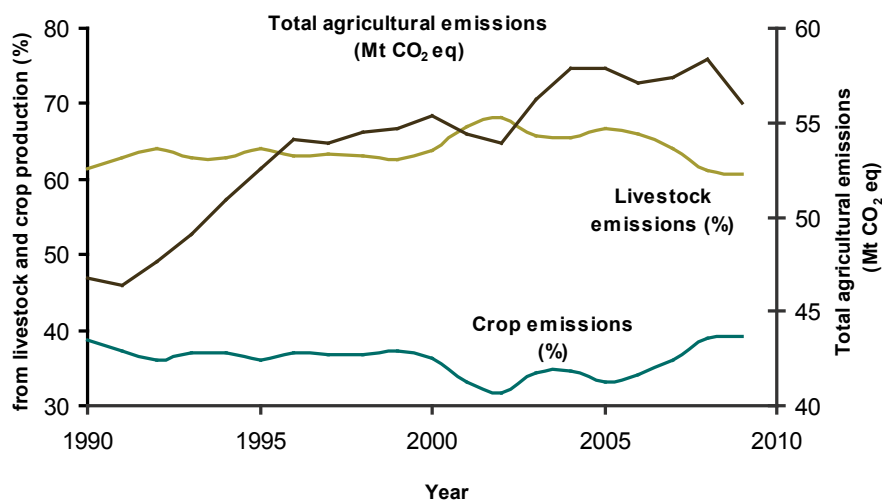
Increases from beef production were, however, partially offset by a 28% reduction of the dairy population. Histori-

cally, Ontario and Quebec have been the location of most of Canada's dairy industry. The dairy quota systems in these and other provinces encouraged the dairy industry to invest in herd improvement in order to increase industry profitability. Emissions associated with dairy cows have fallen by approximately 14% since 1990, as the decline in the dairy herd has been partly offset by a 26% increase in average milk productivity, due to improved genetics and changes in feeding and/or management practices. Therefore, even though the drop in dairy population is driving the emission decline in this category, as was the case with non-dairy cattle, an average cow produces more milk today than in 1990, and also emits more GHGs.

Emissions strictly from crop production are due mainly to either the application of synthetic nitrogen fertilizers or to crop residue decomposition, which is directly proportional to crop yields. There are about two dozen major crops grown in Canada. Corn, wheat, barley and canola require more fertilizers to sustain high levels of production. Emissions from synthetic nitrogen fertilizer consumption have increased substantially, from 9.2 Mt CO<sub>2</sub> eq in 1990 to 13.5 Mt CO<sub>2</sub> eq in 2009. The increase in synthetic fertilizer nitrogen use has jumped from 1.2 Mt N to 1.9 Mt N over the same time period due mainly to a reduction in summerfallow and an intensification of cropping systems in western Canada.

Over the past two decades, emissions from crop residue decomposition varied between 4.9 Mt CO<sub>2</sub> eq and 9.1 Mt CO<sub>2</sub> eq. Severe drought for most regions of the Canadian Prairies in 2001 and 2002 resulted in very poor crop pro-

Figure 2-10 Relative GHG Contribution from Livestock and Crop Production and Total Agricultural Emissions, 1990–2009



duction and, in turn, lower emissions for these years. The impact of the drought is observed in both the emission trend and the relative proportion of emissions attributed to crop or animal growth (Figure 2-10). On the other hand, since 2005, favourable weather conditions along with good commodity prices resulted in record production for soybean, corn, pulse and canola and consequently greater emissions of nitrous oxide (N<sub>2</sub>O).

Overall, through the 1990s, the combination of increased livestock populations and increasing emissions per animal in some animal categories resulted in a change in the relative proportion of GHGs coming from the livestock sector from 63 to 67%, increasing to a high of 69% during the drought years of 2001 and 2002 (Figure 2-10).

### Recent Trends

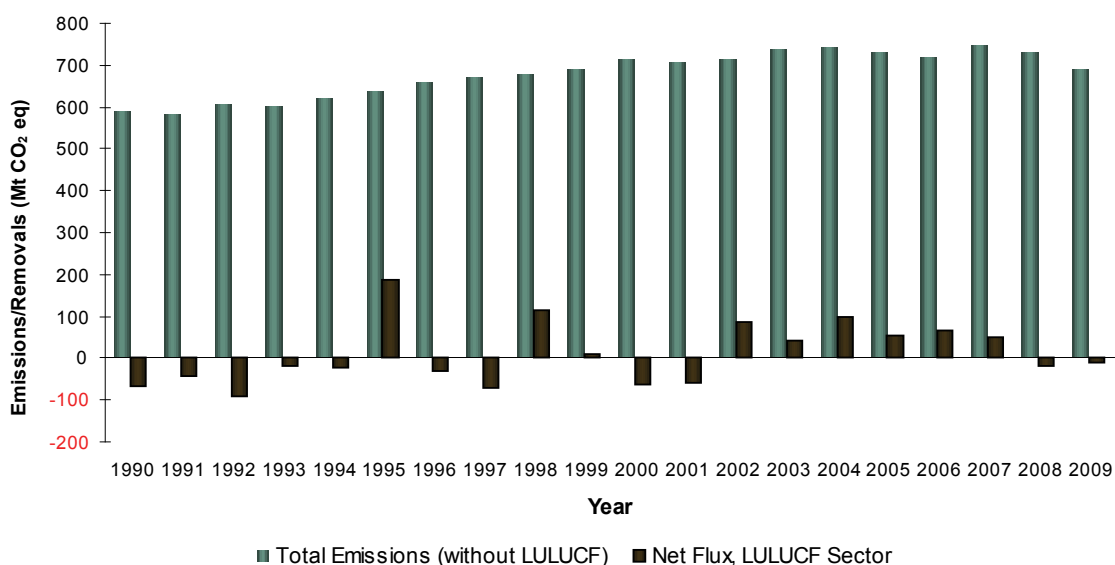
Beef prices were strong from 1990 until 2003, when the occurrence of bovine spongiform encephalopathy (BSE, or mad cow disease) resulted in a worldwide ban on Canadian beef products. A sudden increase in domestic animal populations occurred (9% increase between January 2003 and January 2004). The BSE crisis was not completely resolved until 2005, and since the peak of the crisis in 2005, beef populations have decreased by 13%.

The prices of hogs were also strong from 1990 to 2003 (Statistics Canada 2009b), and large increases in population numbers occurred. However, prices have also suffered important decreases in recent years, and as a result,

populations have also decreased by 21% since their peak in 2005. These population decreases, combined with continued decreasing trends in dairy cattle populations, have decreased emissions from livestock by 12%, or roughly 4.6 Mt CO<sub>2</sub> eq since 2005. At the same time, since 2005, due to improved crop yields and strong grain commodity prices, emissions from crop production have increased by 13%, roughly 2.5 Mt CO<sub>2</sub> eq.

Since 2005, emissions from agriculture, overall, have stabilized, and there appears to be some reversal in the trend of the increasing proportion of emissions from livestock production (Figure 2-10). Total emissions in 2009 for the Agriculture Sector were 2.5 Mt CO<sub>2</sub> eq lower than in 2008 and 2 Mt CO<sub>2</sub> eq lower than in 2005. In 2009, the relative proportion of emissions coming from livestock sector compared to the crop production sector, decreased once again to 61% of total emissions, the same as in 1990. Overall, between 2005 and 2008, decreases in emissions from livestock production have been compensated for by increases in emissions from crop production, resulting in no net changes in agricultural emissions. In 2009 a continued reduction in emissions from livestock production and a reduction emission from crop production, resulted in an apparent decrease in emissions, but this reduction in emissions may be interannual variability.

Figure 2-11 GHG Emissions from LULUCF Relative to Total Canadian Emissions, 1990–2009



### 2.3.5. Land Use, Land-use Change and Forestry Sector

(2009 net GHG removals, 12 Mt, not included in national totals)

The LULUCF Sector reports GHG fluxes between the atmosphere and Canada's managed lands, as well as those associated with land-use changes.

The net LULUCF flux, calculated as the sum of CO<sub>2</sub> emissions and removals and non-CO<sub>2</sub> emissions, displays high interannual variability over the reporting period. In 2009, this net flux amounted to removals of 12 Mt (Figure 2-11).

All emissions and removals in the LULUCF Sector are excluded from the national totals. In 2009, the estimated -12 Mt would, if included, decrease the total Canadian GHG emissions by about 2%.

GHG emissions from sources and removals by sinks are estimated and reported for four categories of managed lands: Forest Land, Cropland, Wetlands and Settlements.

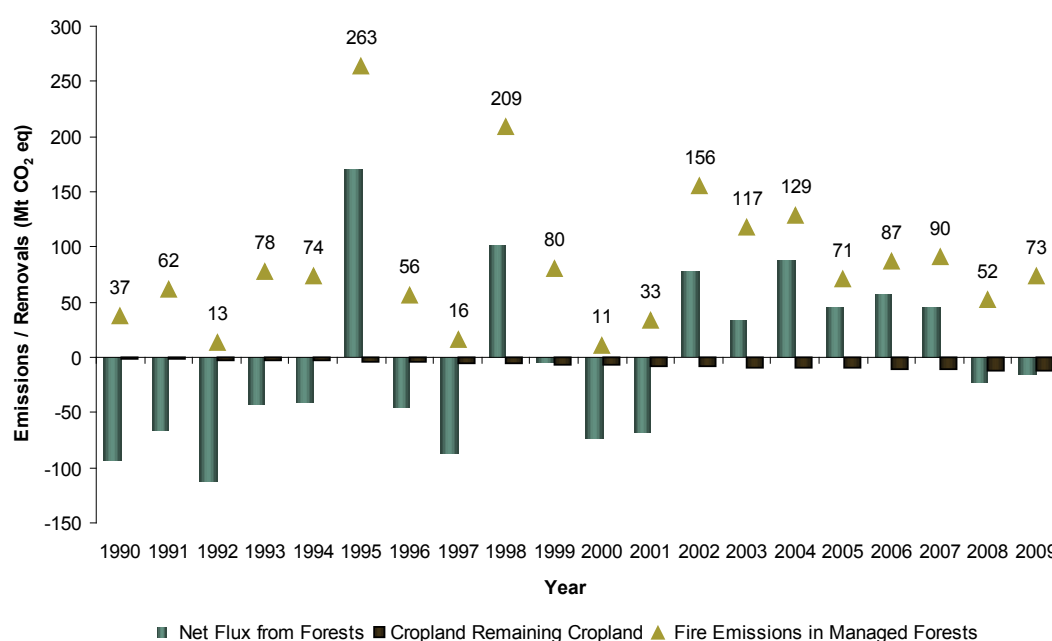
The Forest Land category includes GHG emissions from and removals by Canada's managed forests. The net flux in forest land displays an important annual variability due to the erratic pattern of forest wildfires, which masks underly-

ing patterns of interest in the sector. Important subsectoral trends associated with human activities in managed forests include a 28% increase in the carbon removed in harvested wood biomass between 1990 and the peak harvest years of 2004–2005. Since then, significant reductions in forest management activities have occurred, with a 39% decline in harvest levels, which in 2009 have reached their lowest point for the two decades covered by this report (32 Mt C). This trend reflects a deep restructuring of the Canadian forest economic sector, aggravated by the consequences of the economic recession in the United States, Canada's main export market.

The high variability in the net flux from managed forests is associated with the immediate impact of wildfires, which alone accounted for annual emissions of between 11 and 263 Mt over the period from 1990 to 2009 (Figure 2-12). Likewise, the immediate and long-term effect of the catastrophic Mountain Pine Beetle infestation in Western Canada will undoubtedly continue to influence the GHG trends.

Note that the current default approach ignores long-term carbon storage in wood products. Taking into account this storage, emission estimates from harvesting in the year 2009 alone could be reduced by 13 to 54 Mt, depending on the approach used to account for the fate of this carbon.

Figure 2-12 Selected GHG Emissions and Removals in LULUCF, 1990–2009



The Cropland subcategory includes the effect of agricultural practices on CO<sub>2</sub> emissions from, and removals by, arable soils and the immediate and long-term impacts of forest and grassland conversion to cropland. The steady decline in emissions from cropland is noteworthy, from 11 Mt CO<sub>2</sub> eq in 1990 to a net removal of 6.9 Mt CO<sub>2</sub> eq in 2009. This pattern largely results from changing agricultural land management practices in western Canada, such as the extensive adoption of conservative tillage practices (over 11 Mha of cropland since 1990), reduction in summerfallow by more than 60% and an increase in perennial forage crops. A decline in forestland conversion to cropland has also contributed to this trend in emissions/removals. As a result, the net CO<sub>2</sub> removals due to the management of Cropland Remaining Cropland increased from about 1.5 Mt in 1990 to 12 Mt in 2009.

CO<sub>2</sub> emissions from peatlands managed for peat extraction and from land flooding are reported under the Wetlands category. Emissions from managed peatlands have increased 34% since 1990, amounting to 1.2 Mt in 2009. Land conversion to flooded lands (reservoirs) accounted for 4.2 Mt in 1990, decreasing to 1.3 Mt in 2009. Note that reservoirs flooded for more than 10 years are excluded from the accounting (IPCC 2003).

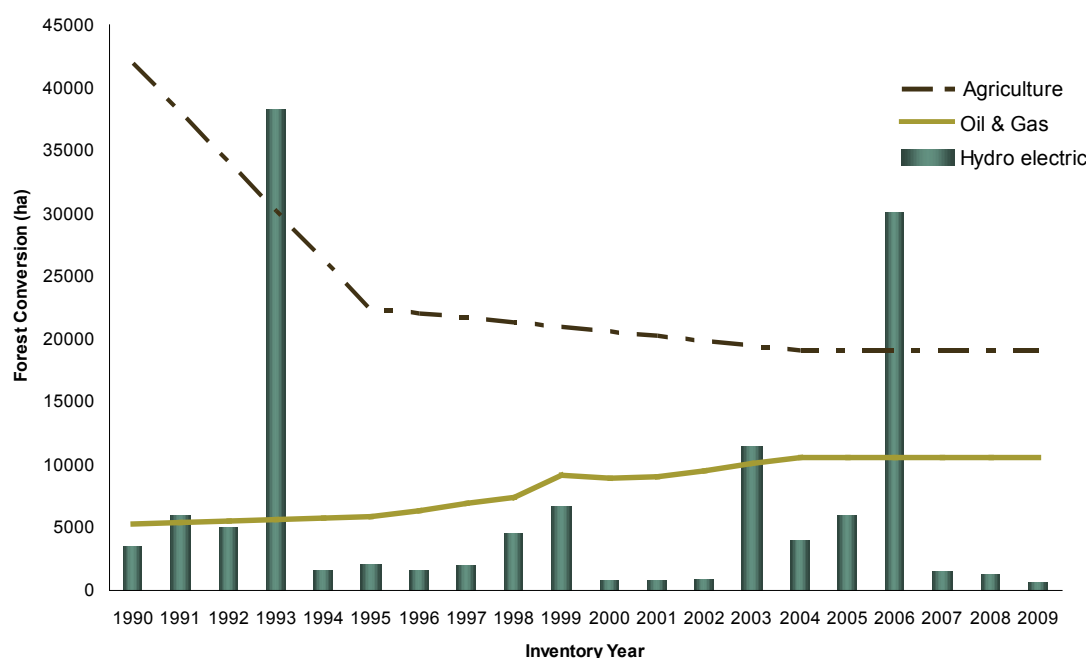
The conversion of forests to other land is a prevalent practice in Canada, driven by a great variety of circum-

stances across the country, including policy and regulatory frameworks, market forces and resource endowment. The economic drivers of forest conversion are diverse and result in heterogeneous spatial and temporal patterns of forest conversion. Since 1990, approximately one million hectares of forest have been lost in Canada. GHG emissions from forest conversion have dropped from 26 Mt CO<sub>2</sub> eq in 1990 to 18 Mt CO<sub>2</sub> eq in 2009. Primary drivers of forest conversion include agricultural expansion, resource extraction and hydroelectric development. Geographically, the highest average rates of forest conversion occur in the Boreal Plain (24 kha/year) and the Boreal Shield East (9 kha/year), which accounts for 45% and 17% of the total forest area lost in Canada since 1990, respectively.

The largest driver of forest conversion is agricultural expansion, which since 1990 has replaced over 470 kha of forests and accounts for 45% of the cumulative area of forest conversion since 1990. However, annual rates of deforestation to agriculture have dropped from 42 kha in 1990 to 19 kha in 2009 (Figure 2-13). This decrease predominantly took place in the Boreal Plains, Sub-humid Prairies and Montane Cordillera of western Canada, following a period of active agricultural expansion in the previous decades.

Forest clearing for resource extraction, which includes mining, oil and gas extraction, forestry roads, and peat

Figure 2-13 Trends in Annual Rates of Forest Conversion due to Agricultural Expansion, Oil and Gas Extraction and Hydroelectric Developments



extraction is the second-largest driver of forest conversion. Resource extraction expanded at the expense of over 310 kha of forests and accounts for 29% of the cumulative area of forest conversion since 1990. Forest clearing for oil and gas extraction have steadily increased, from to 5.3 kha per year in 1990 to 10.6 kha per year in 2009 (Figure 2-13) and has largely occurred in the Boreal Plains of the northern Prairies.

Forest conversion due to hydroelectric development is episodic, corresponding to the occasional impoundment of large reservoirs (e.g. La Forge 1 in 1993 and Eastmain-1 in 2006) (Figure 2-13). Cumulative areas of forests converted for the creation of hydro reservoirs and associated infrastructure equal 130 kha, which accounts for 12% of forest conversion over the time period. Hydroelectric development occurs mainly in the Taiga Shield East and the Boreal Shield East.

Other drivers of forest conversion such as development of built-up lands and transportation routes have remained relatively constant. Development of built-up lands resulted in the conversion of 116 kha of forest in the 1990 to 2009 time period and accounts for 11% of the cumulative area of forest conversion. Annual rates of forest clearing for this activity were approximately 5.5 kha for both 1990 and 2009 with only slight variation in the time series.

### 2.3.6. Waste Sector (2009 GHG emissions, 22 Mt)

From 1990 to 2009, GHG emissions from the Waste Sector increased 13%, which is less than the population growth of 22%, while over the same period total national GHG emissions grew by 16.9%. In 2009, these emissions represented 3.1% of the total national GHG emissions, compared with a 3.2% contribution in 1990. Of the 22 Mt total emissions from this Sector in 2009, Solid Waste Disposal on Land, which includes municipal solid waste (MSW) landfills and

wood waste landfills, accounted for 20 Mt. CH<sub>4</sub> emissions produced by the decomposition of biomass in MSW landfills were responsible for 83% of the emissions from the Waste Sector. Emissions from municipal wastewater treatment and incineration of waste (excluding emissions from incineration of biomass material) contributed 1.01 Mt and 0.26 Mt, respectively, to the total from the Waste Sector (Table 2-10). Figure 2-14 presents the emission trends for each of the three subsectors compared with the total emissions for the Waste Sector between 1990 and 2009. The tables in Annex 12 summarize this information nationally by CO<sub>2</sub> equivalent and by category (i.e. individual gas and source).

CH<sub>4</sub> emissions from MSW landfills increased by 16% between 1990 and 2009, despite an 81% increase in landfill gas capture and combustion over the same period.. Approximately 349 kt of CH<sub>4</sub> (or 7334 kt CO<sub>2</sub> eq) were captured by the 68 landfill gas collection systems operating in Canada (Environment Canada 2011).<sup>30</sup> Of the total amount of CH<sub>4</sub> collected in 2009, 51% (179 kt) was utilized for various energy purposes and the remainder was flared. As shown in Figure 2-15, Canada has considerably increased the number of landfill sites collecting gas since facility data collection was initiated in 1997.

Over the time series the distribution of the collected landfill gas being utilized vs. flared has changed as presented in Figure 2-16. Although the percentage of gas being utilized drops from 70% to 51% from 1997 to 2009, this is due to a greater number of facilities initiating gas collection and which are flaring in preparation to starting up utilization units.

30 Four landfill gas capture facilities did not provide data for the 2007 landfill gas inventory by December 31, 2008. Since two of these facilities had provided data for the 2005 landfill gas inventory, it was assumed that the quantities of landfill gas collected were constant for 2005, 2006 and 2007. The other two sites were new installations where gas collection was being constructed and for which no landfill gas collection quantity information was available.

Table 2-10 GHG Emissions from Waste, Selected Years

GHG Source Category	GHG Emissions (Mt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Waste Sector	19	20	21	21	22	22	21	22
Solid Waste Disposal on Land	18	19	20	20	21	20	20	20
Wastewater Handling	0.78	0.93	0.97	0.98	0.99	1.00	1.00	1.00
Waste Incineration	0.40	0.25	0.23	0.24	0.24	0.25	0.25	0.26

Note: Totals may not add up due to rounding.

Figure 2–14 GHG Emissions from Waste, 1990–2009

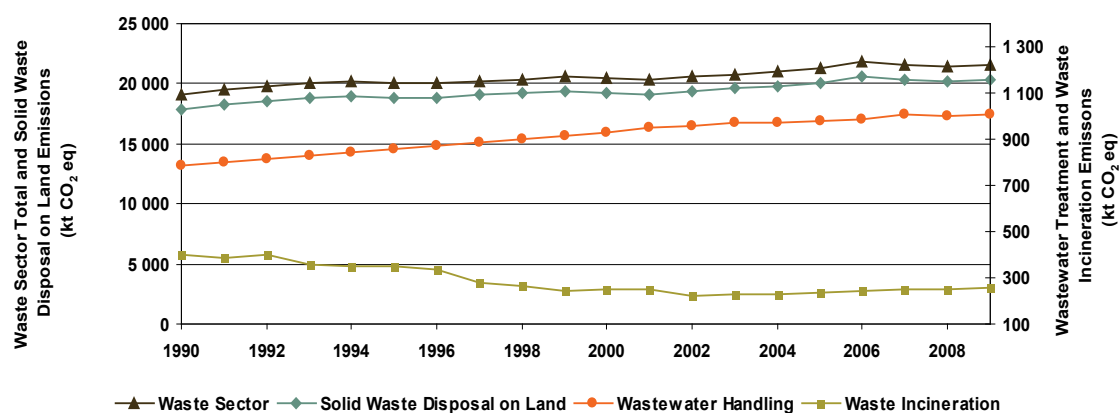


Figure 2–15 Number of Active Gas Collection Landfill Sites in Canada

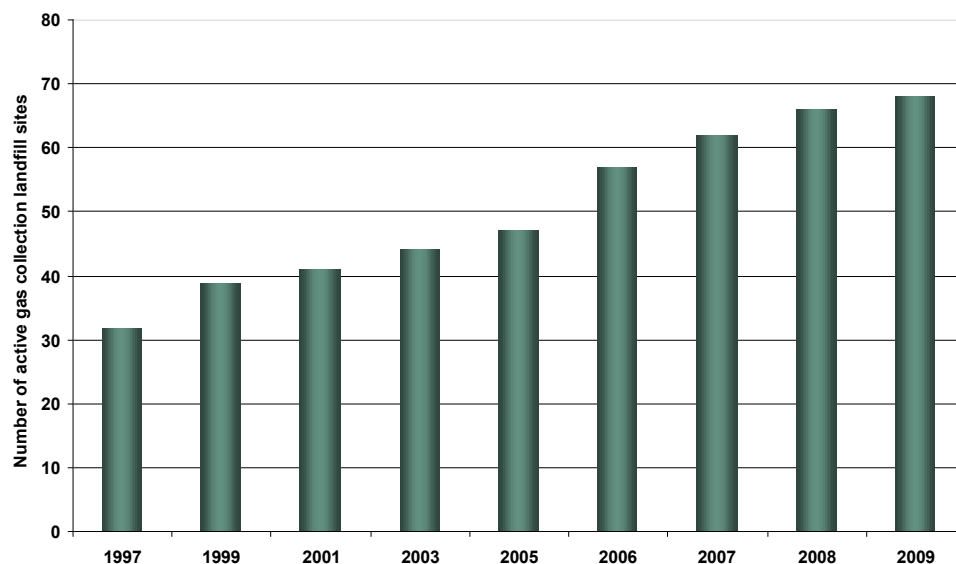
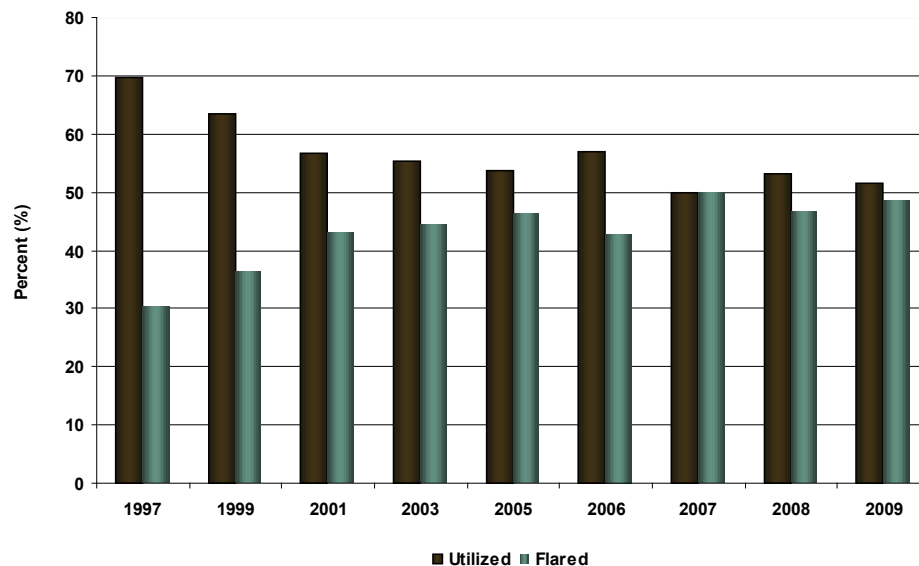


Figure 2–16 Proportion of Landfill Gas Utilized vs Flared



GHG emissions from landfills were estimated for two solid waste types: MSW and wood waste landfills, both of which produce CH<sub>4</sub> anaerobically.<sup>31</sup> The CH<sub>4</sub> production rate at a landfill is a function of several factors, including the mass and composition of biomass being landfilled, the landfill temperature, and the moisture entering the site from rainfall.

CH<sub>4</sub> capture and waste diversion programs at landfills have made significant contributions to reductions in emissions during this period. The quantity of CH<sub>4</sub> captured at MSW landfills for flaring or combustion for energy recovery purposes in 2009 amounted to 29% of the total generated emissions from this source, as compared to 21% in 1990. Per capita emissions from the Waste Sector decreased by 7.1% from 1990 to 2009, owing primarily to the increasing quantities of CH<sub>4</sub> captured at landfill sites (Figure 2-17). The amount of CH<sub>4</sub> captured increased by 81% from 1990 to 2009, and the amount of waste diverted as a percentage of the waste generated has fluctuated from 21% to 25% over the period between 1998 and 2006 (Statistics Canada 2000, 2003, 2004, 2007, 2008). Although the quantity of waste placed in MSW landfills increased by 29% from 1990 to 2009, the landfilled quantity per capita increased by only 5.7%. The amount of waste exported from Canada to the United States for the years 1990 and 2009 was estimated at 100 kt and 3490 kt, respectively, giv-

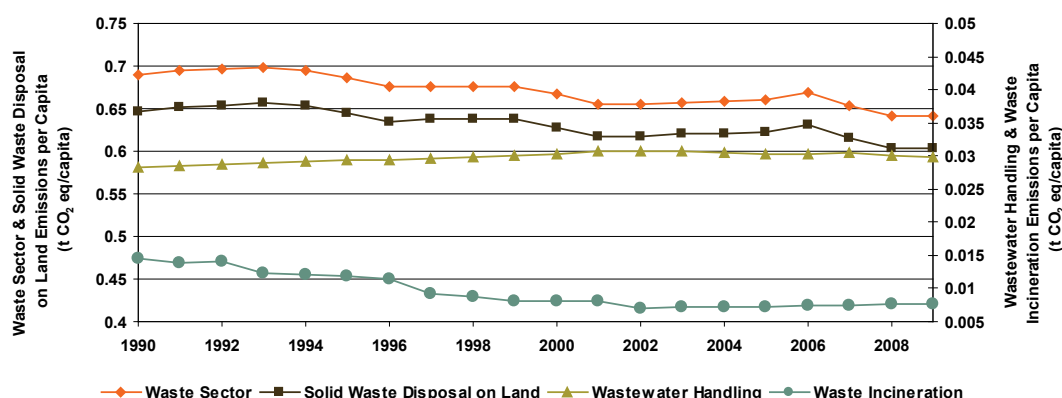
ing a 3 390% increase in the amount of waste (residential and non-residential) exported over this period. However, emissions from MSW landfills are expected to increase in subsequent years as a result of restrictions on the exportation of solid waste from Ontario. An agreement was signed between the State of Michigan and the Province of Ontario that calls for a 20% reduction in municipally managed exported waste by the end of 2007, 40% by the end of 2008 and 100% by the end of 2010 (Ontario Ministry of the Environment 2006). This is all based on an estimated figure of 1.34 million tonnes of municipally managed wastes reported for 2005.<sup>32</sup> Municipally managed wastes do not include institutional, commercial or industrial wastes.

From 1990 to 2009, the population growth trend (22%) exceeded that of the Sector emissions (13%). The decline in the growth of emissions per capita observed in the mid-1990s, shown in Figure 2-17, is directly attributable to landfill CH<sub>4</sub> capture and waste diversion programs. However, from 1997 to 1999, there was a reduction followed by an increase in the quantities of landfill gas captured. These changes have an inversely proportional influence on the emissions per capita, which is apparent in Figure 2-17. The 2006–2009 drop in emissions from Solid Waste Disposal on Land may be more attributable to the extrapolation method used to derive the waste quantities landfilled for those years rather than a true reflection of the situation. A more accurate determination will be made for 2007–2009 emissions in the subsequent NIR submission, when the results from the next biennial MSW waste quantity landfilled data collection survey will be included. It was expected that this information would be incorporated in the present

31 When waste consists of biomass, the CO<sub>2</sub> produced from burning or aerobic decomposition is not accounted for in the Waste Sector. This is because, in the case of agricultural biomass, it is deemed to be a sustainable cycle (carbon in CO<sub>2</sub> will be sequestered when the biomass regenerates in crop reproduction). In the case of biomass from forest products, the emissions of CO<sub>2</sub> are accounted for as part of the LULUCF Sector (forest harvests). However, waste that decomposes anaerobically produces CH<sub>4</sub>, which is not used photosynthetically and therefore does not sequester carbon in biomass regeneration and is not accounted for in forest harvest estimates. The production and release of unburned CH<sub>4</sub> from waste are therefore accounted for in GHG inventories.

32 Personal communication between Jim Hiraish, Ontario Ministry of the Environment and Craig Palmer, Environment Canada, Nov. 30, 2007.

Figure 2-17 Per Capita GHG Emission Trend for Waste, 1990–2009





submission; however, the survey report from Statistics Canada was delayed and was not available in time for this year's emission estimations.

In terms of trends in emissions per capita from the Wastewater Handling subsector, there was an overall slight increase of 6% from 1990 to 2009. From 1990 to 2001 there was a constant slow growth, which coincided with the trend in N<sub>2</sub>O emissions from human sewage that peaked in 2001. The Wastewater Handling subsector emissions then held relatively steady as the human sewage emissions dropped slightly due to lower protein consumption, which was offset, to some extent, by population-growth-driven CH<sub>4</sub> emissions from the Municipal Wastewater subsector. In contrast, the Waste Incineration subsector showed a significant decrease in GHG emissions over the 1990–2009 time series (Figure 2-17). Total incineration emissions per capita decreased by 47% over the time series, with the greatest decline in emissions per capita for incineration occurring between 1992 and 1997, due mainly to the closure of aging incinerators.

## 2.4. Emission Trends for Ozone and Aerosol Precursors

Emissions of ozone and aerosol precursors fell over the 1990–2009 period. Sulphur oxides (SO<sub>x</sub>) decreased by 54%, non-methane volatile organic compound (NMVOC) emissions declined by 28%, nitrogen oxides (NO<sub>x</sub>) emissions were down by 21%, and carbon monoxide emissions fell by 43%. (See Annex 10 for data tables).

# Chapter 3

## Energy (CRF Sector 1)

### 3.1. Overview

Overall, the Energy Sector contributed about 82% (or 566 Mt) of Canada's total greenhouse gas (GHG) emissions in 2009 (Table 3–1). The Energy Sector accounts for all GHG (CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O) emissions from stationary and transport fuel combustion activities as well as fugitive emissions from the fossil fuel industry. Fugitive emissions associated with the fossil fuel industry are the intentional (e.g. venting) or unintentional (e.g. leaks, accidents) releases of GHGs that may result from production, processing, transmission and storage activities. Emissions from flaring activities by the oil and gas industry are reported in the Fugitive category, since their purpose is not to produce heat or to generate mechanical work (IPCC/OECD/IEA 1997).

Emissions resulting from stationary fuel combustion include, for example, the use of fossil fuels by the electricity generating industry, the oil and gas industry, the manufacturing and construction industry, and the residential and commercial sector. Only CH<sub>4</sub> and N<sub>2</sub>O emissions resulting from the combustion of biomass fuels by the pulp and

paper industry and by the residential sector are accounted for in the Energy Sector, whereas CO<sub>2</sub> emissions resulting from the use of biomass are reported as a memo item in the Common Reporting Format (CRF) tables.

GHG emissions from the combustion (and evaporation) of fuel for all transport activities, such as Civil Aviation (Domestic Aviation), Road Transportation, Railways, Navigation (Domestic Marine) and Other Transportation (Off-road and Pipelines), are included in the Transport subsector. Usage of transport fuels (such as gasoline and diesel) by the mining industry, by the oil and gas extraction industry, and by agriculture and forestry is also included under Other Transportation. Emissions from international bunker activities (only in regards to aviation and marine) are reported as a memo item in the CRF tables.

### 3.2. Fuel Combustion (CRF Category 1.A)

Fuel combustion sources include all emissions from the combustion of fossil fuels. Major subsectors include Energy Industries, Manufacturing Industries and Construction, Transport, and Other Sectors (which include the residential and commercial categories). Methods used to calculate emissions from fuel combustion are consistent throughout and are presented in Annex 2: Methodology and Data for Estimating Emissions from Fossil Fuel Combustion; the estimation methodologies are consistent with the revised 1996 Intergovernmental Panel on Climate Change (IPCC) Tier 2 approach, with country-specific emission factors and parameters.

Table 3–1 GHG Emissions from Energy, Selected Years

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Energy Sector	468 000	586 000	605 000	595 000	583 000	613 000	597 000	566 000
Fuel Combustion (1.A)	425 000	523 000	541 000	532 000	519 000	551 000	535 000	505 000
Energy Industries (1.A.1)	143 026	193 000	197 000	189 000	181 000	193 000	185 000	161 000
Manufacturing Industries and Construction (1.A.2)	64 500	69 500	72 300	68 800	70 400	79 400	72 600	74 900
Transport (1.A.3)	146 000	180 000	189 000	193 000	192 000	197 000	196 000	190 000
Other Sectors (1.A.4)	71 600	80 500	82 900	80 600	75 300	80 900	80 600	78 700
Fugitive Emissions from Fuels (1.B)	42 100	63 000	64 000	63 100	64 300	62 800	62 300	60 700

Note: Totals may not add up due to rounding.

In 2009, about 505 Mt (or 73%) of Canada's GHG emissions were from the combustion of fossil fuels (Table 3–1). The overall GHG emissions from fuel combustion activities have increased by 10% since 1990 and have decreased by 6% since 2008. Between 1990 and 2009, combustion-related emissions from the Energy Industries and from the Transport category increased by about 13% and 30%, respectively (Figure 3–1).

## 3.2.1. Energy Industries (CRF Category 1.A.1)

### 3.2.1.1. Source Category Description

The Energy Industries subsector is divided into the following three categories: Public Electricity and Heat Production, Petroleum Refining, and Manufacture of Solid Fuels and Other Energy Industries (which consists primarily of crude oil, coal, natural gas, bitumen and synthetic crude oil production).

In 2009, the Energy Industries subsector accounted for 161 Mt (or about 23%) of Canada's total GHG emissions, an increase of about 13% since 1990. The Public Electricity and Heat Production subsector accounted for 61% (or

98 Mt) of the Energy Industries' GHG emissions, while Petroleum Refining, and the Manufacture of Solid Fuels and Other Energy Industries contributed 12% (20 Mt) and 27% (44 Mt), respectively (Table 3–2). Additional discussions on trends in emissions from the Energy Industries subsector are to be found in the Emission Trends chapter (Chapter 2).

The Energy Industries subsector includes all emissions from stationary fuel combustion sources related to utility electricity generation and the production, processing and refining of fossil fuels. All of the emissions associated with the fossil fuel industry are estimated, although a portion of emissions from coal mining and from oil and gas extraction associated with the Petroleum Refining and the Manufacture of Solid Fuels and Other Energy Industries categories have been allocated to the Manufacturing Industries and Construction—Mining and Transport—Other subsectors, because fuel consumption data at a lower level of disaggregation are not available. Combustion emissions associated with the pipeline transmission of oil and natural gas are included under Other Transportation according to the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997).

Although actually associated with the Energy Industries, emissions from venting and flaring activities related to the production, processing and refining of fossil fuels are

Figure 3–1 GHG Emissions from Fuel Combustion, 1990–2009

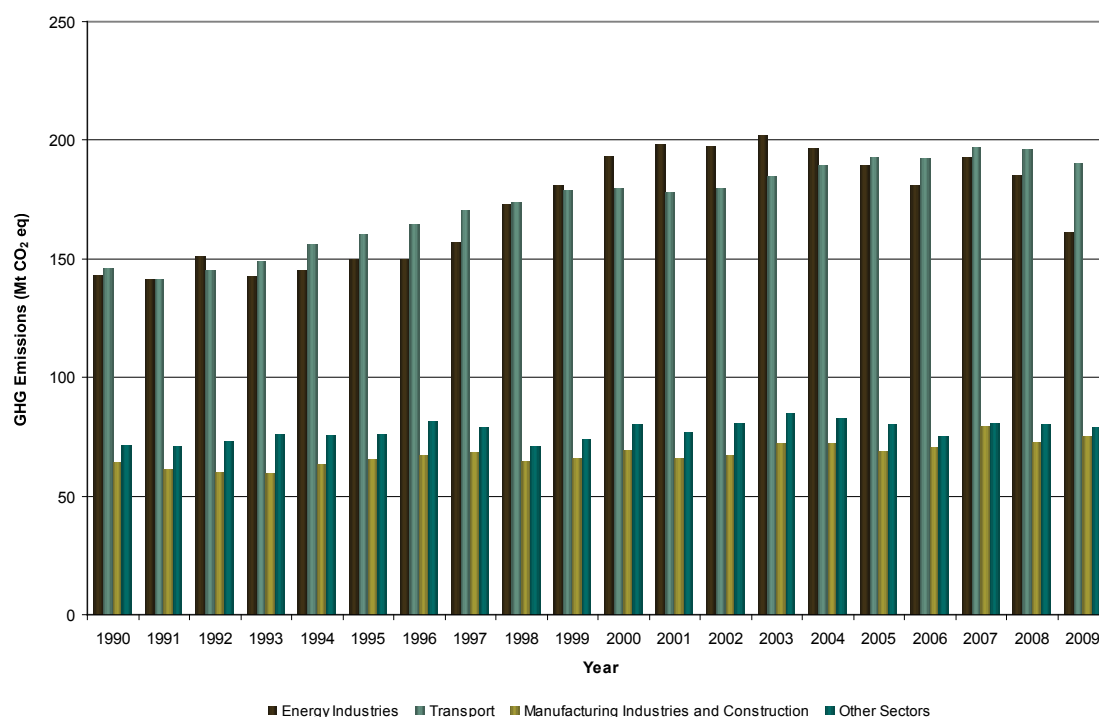


Table 3–2 Energy Industries GHG Contribution

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Energy Industries TOTAL (1.A.1)	143 000	193 000	197 000	189 000	181 000	193 000	185 000	161 000
Public Electricity and Heat Production	91 600	127 000	123 000	123 000	115 000	122 000	116 000	97 900
Electricity Generation—Utilities	91 400	126 000	122 000	122 000	114 000	122 000	116 000	96 800
Heat/Steam Generation	200	500	1 200	900	800	800	700	1 000
Petroleum Refining	18 000	16 000	20 000	19 000	19 000	21 000	20 000	20 000
Manufacture of Solid Fuels and Other Energy Industries	34 000	50 000	53 000	48 000	48 000	49 000	49 000	44 000

Note: Totals may not add up due to rounding.

reported as fugitive emissions (refer to Section 3.2, Fuel Combustion (CRF Category 1.A)).

### Public Electricity and Heat Production (CRF Category 1.A.1.a)

The Public Electricity and Heat Production category includes emissions associated with the production of electricity and heat from the combustion of fuel in public utility thermal power plants. The estimated GHG emissions from this sector do not include emissions from industrial generation; rather, these emissions have been allocated to the specific industrial sectors.

The electricity supply grid in Canada includes combustion-derived electricity as well as hydro, nuclear and other renewables (wind, solar and tidal power). Total power generated from wind, tidal and solar resources is relatively small compared with that from Canada's significant hydro and nuclear installations. Nuclear, hydro, wind, solar and tidal electricity generators are not direct emitters of GHGs; therefore, GHG estimates reflect emissions from combustion-derived electricity only. Steam generation and internal combustion engines are the primary systems used to generate electricity through thermal processes.

- Steam turbine boilers are fired with coal, petroleum coke, heavy fuel oil, natural gas or biomass. For turbine engines, the initial heat may be generated from natural gas and refined petroleum products (RPPs—e.g. light fuel oil or diesel fuel). Reciprocating engines can use natural gas and/or a combination of RPPs, whereas gas turbines are also fired with natural gas or RPPs.

Emissions of CH<sub>4</sub> and N<sub>2</sub>O from the combustion of landfill gas (LFG) for heat, steam and electricity generation are included, while CO<sub>2</sub> emissions are excluded from totals

but reported separately in the United Nations Framework Convention on Climate Change (UNFCCC) CRF tables as a memo item.

### Petroleum Refining (CRF Category 1.A.1.b)

The Petroleum Refining category includes direct emissions from the production of petroleum products from a raw feedstock. Conventional or synthetic crude oil is refined by distillation and other processes into petroleum products such as heavy fuel oil, residential fuel oil, jet fuel, gasoline and diesel oil. The heat required for these processes is created by combusting either internally generated fuels (such as refinery fuel gas) or purchased fuels (such as natural gas). CO<sub>2</sub> generated as a by-product during the production of hydrogen in the steam reforming of natural gas is reported in the Fugitive category (Section 3.3).

The Petroleum Refining category also includes a small portion of combustion emissions that result from the upgrading of heavy oil from oil sands mining and *in situ* extraction to produce synthetic crude oil and/or other refined products such as diesel oil for sale. Also, owing to the level of aggregation of the fuel consumption data and the assumptions used to report the emissions associated with the downstream (petroleum refining) and the upstream (consisting of solid, oil and gas production) industry, a small portion of emissions associated with Petroleum Refining (such as CH<sub>4</sub>) are included in the Manufacture of Solid Fuels and Other Energy Industries category (and vice versa for a portion of the emissions associated with bitumen upgrading in the oil sands industry). Refer to Annex 2 for additional details on the method used to disaggregate the activity data.

## **Manufacture of Solid Fuels and Other Energy Industries (CRF Category 1.A.1.c)**

The Manufacture of Solid Fuels and Other Energy Industries category comprises fuel combustion emissions associated with the crude oil, natural gas, oil sands mining, bitumen extraction and upgrading, and coal mining industries. A portion of emissions associated with coal mining and oil and gas extraction (which includes bitumen upgrading) are reported in the Manufacturing Industries and Construction–Mining category, whereas emissions associated with pipeline transmission and with the use of transport fuels (such as gasoline and diesel oil) in off-road applications in the mining and the oil & gas mining and extraction industry are reported under Other Transportation, since the fuel data cannot be further disaggregated in the national energy balance as compiled by Statistics Canada.

Upgrading facilities are responsible for producing synthetic crude oil based on a feedstock of bitumen produced by oil sands mining, extraction and *in situ* recovery activities (i.e. thermal extraction). The synthetic (or upgraded) crude oil has a hydrocarbon composition similar to that of conventional crude oil, which can be refined to produce RPPs such as gasoline and diesel oil. Upgrading facilities also rely on internally generated fuels such as process gas and natural gas for their operation, which result in both combustion- and fugitive-related emissions.

### **3.2.1.2. Methodological Issues**

Emissions for all source categories are calculated following the methodology described in Annex 2 and are primarily based on fuel consumption statistics reported in the *Report on Energy Supply–Demand in Canada* (RES—Statistics Canada #57-003), with additional information from the *Electric Power Generation, Transmission and Distribution* (EPGTD) publication (Statistics Canada #57-202). LFG utilization estimates are provided by the Waste Sector. The method is consistent with the IPCC Tier 2 approach, with country-specific emission factors.

## **Public Electricity and Heat Production (CRF Category 1.A.1.a)**

The Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) require the Public Electricity and Heat Production category to include only emissions generated by public utilities. Emissions associated with industrial generation are allocated to the industry that produces the energy under the appropriate industrial category within the Energy Sector,

regardless of whether the energy is for sale or for internal use. The rationale for this is that the IPCC recognizes that it is difficult to disaggregate emissions in cogeneration facilities (i.e. to separate the electricity component from the heat component of fuel use). Statistics Canada fuel-use data in the RESD do distinguish industrial electricity generation data, but aggregate the data into one category titled industrial electricity generation. Industrial electricity generation emissions were reallocated to their respective industrial subsectors using the RESD input data. The methodology is described in greater detail in Annex 2.

## **Petroleum Refining (CRF Category 1.A.1.b)**

Emissions for this category are calculated using all fuel use attributed to the petroleum refining industry and include all petroleum products (including still gas, petroleum coke and diesel) reported as producer-consumed/own consumption as well as purchases of natural gas for fuel use by refineries. The fuel-use data in the RESD include volumes of flared fuels; however, flaring emissions are calculated and reported separately in the Fugitive category (refer to Section 3.3.2). The fuel-use and emission data associated with flaring are subtracted to avoid double counting.

## **Manufacture of Solid Fuels and Other Energy Industries (CRF Category 1.A.1.c)**

Emissions for this category are calculated using all fuel use attributed to fossil fuel producers (including petroleum coke, still gas, natural gas, natural gas liquids [NGLs] and coal data). The fuel-use data in the RESD include volumes of flared fuels; however, flaring emissions are calculated and reported separately in the Fugitive category. The fuel-use and emission data associated with flaring are subtracted to avoid double counting.

### **3.2.1.3. Uncertainties and Time-Series Consistency**

The estimated uncertainty for the Energy Industries subsector ranges from -4% to 6% for all gases and from -6% to 2% for CO<sub>2</sub> alone (ICF Consulting 2004).

The uncertainties for the Energy Industries subsector are largely dependent on the collection procedures used for the underlying activity data as well as on the representativeness of the emission factors for specific fuel properties. Commercial fuel volumes and properties are generally well known, whereas there is greater uncertainty surrounding both the reported quantities and properties of

non-marketable fuels (e.g. *in situ* use of natural gas from the producing wells and the use of refinery fuel gas). For example, in the Petroleum Refining category, the CO<sub>2</sub> emission factors for non-marketable fuels as consumed, such as refinery still gas, petroleum coke and catalytic coke, have a greater influence on the uncertainty estimate than the CO<sub>2</sub> factors for commercial fuels. As well, new coal CO<sub>2</sub> emission factors were developed using statistical methods and 95% confidence intervals. The use of the 95% confidence intervals resulted in large uncertainties for these fuels ( $\pm 20\%$ ) even though the values are considered more representative and of better overall quality.

Approximately 84% of the 2009 emissions from the Manufacture of Solid Fuels and Other Energy Industries category are associated with the consumption of natural gas in the natural gas production and processing industry. The uncertainty for this category is influenced by the CO<sub>2</sub> emission factors ( $\pm 6\%$ ) and CH<sub>4</sub> emission factors (0% to 240%) for the consumption of unprocessed natural gas. Provincially weighted natural gas emission factors were used to estimate emissions for the natural gas industry due to a lack of plant-level information, such as the physical composition of unprocessed natural gas (which will vary from plant to plant). Thus, the overall uncertainty estimate is based on a rather broad assumption as well.

The estimated uncertainty for CH<sub>4</sub> (1% to 230%) and N<sub>2</sub>O (-23% to 800%) emissions for the Energy Industries subsector is influenced by the uncertainty associated with the emission factors. Additional expert elicitation is required to improve the CH<sub>4</sub> and N<sub>2</sub>O uncertainty estimates for some of the emission factor uncertainty ranges and probability density functions developed by ICF Consulting, since insufficient time was available to have these assumptions reviewed by industry experts.

The estimates for the Energy Industries subsector are consistent over time and calculated using the same methodology.

### 3.2.1.4. QA/QC and Verification

Quality control (QC) checks were done in a form consistent with the IPCC's *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000). Elements of a Tier 1 QC check include a review of the estimation model, activity data, emission factors, time-series consistency, transcription errors, reference material, conversion factors and unit labelling, as well as sample emission calculations.

Minor activity data revisions affecting historical data were identified during the review and corrected. The level of trend analysis and assessment has been improved through the use of additional sources of data for comparison purposes, such as facility-level GHG emissions reporting via Environment Canada's mandatory reporting program for major emitters. No mathematical errors were found during the QC checks. The data, methodologies and changes related to the QC activities are documented and archived in both paper and electronic form.

### 3.2.1.5. Recalculations

Several improvement activities have contributed to increased accuracy of the data, as well as to their comparability and consistency with that of the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) and UNFCCC requirements. As discussed below in more detail, revised activity data and improved emission factors contributed to recalculations along with the reallocation of emissions in the Public Electricity and Heat Production, the Petroleum Refining, and the Manufacturing of Solids and Other Energy Industries categories.

*Activity Data:* The 2008 fuel-use data were revised by Statistics Canada, and estimates were recalculated accordingly.

*Emission Factors:* Revisions to CO<sub>2</sub> emission factors for coal resulted in recalculations. These new emission factors were developed using significantly larger data sets across the entire time series and are expected to better reflect the properties and variability of coals consumed in Canadian power plants.

*Public Electricity and Heat Production:* A method was developed to properly allocate the 1990 to 2008 fuels used to generate electricity by industry (as discussed in Annex 2) to their respective industrial sources. This new approach ensures a representative trend and better comparability of emissions data, and is consistent with the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997). This method involves a reallocation of emissions; the overall emission estimates are therefore not affected by this update.

*Petroleum Refining and the Manufacturing of Solids and Other Energy Industries:* As mentioned in Section 3.2.1.2, the fuel-use data in the RESD include volumes of flared fuels; however, flaring emissions are calculated and reported separately in the Fugitive category. The fuel-use and emission data associated with flaring are subtracted to avoid double counting. The method used to reallocate



1990 to 2008 flaring emissions was revised. The previous method subtracted all oil-related flaring estimates (i.e. oil production, well drilling, refining) from CRF category 1.A.1.b Petroleum Refining, and the remainder (non-oil-related flaring) from CRF category 1.A.1.c Manufacture of Solid Fuels and Other Energy Industries. The new method subtracts only petroleum refining flaring emissions from 1.A.1.b Petroleum Refining and the remainder from the Manufacture of Solid Fuels category. This results in an increase in gaseous fuel emission estimates for 1.A.1.b Petroleum Refining and a decrease in 1.A.1.c Manufacture of Solid Fuels and Other Energy Industries for the full time series. As this is just a reallocation of emissions, the overall emission estimates are not changed by this revision.

In addition, emissions from producer consumption of natural gas liquids (i.e. butane, propane) were incorrectly allocated to 1.A.1.b Petroleum Refining. This resulted in petroleum refining emissions in provinces that do not have petroleum refineries. This was corrected by allocating producers' own use of those fuels to 1.A.1.c Manufacture of Solid Fuels and Other Energy Industries. This correction resulted in changes to gaseous fuel emission estimates for both CRF categories mentioned for the full time series. As above, the overall emission estimates were not changed by this revision as it was only a reallocation exercise.

### 3.2.1.6. Planned Improvements

With the proliferation of publicly reported data, Tier 3 methods for the Public Electricity and Heat Production category are being investigated with the eventual goal of developing a bottom-up inventory. Increases in the usage of combined heat and power plants (and co-generation systems) require additional research and investigation to ensure that emissions are appropriately allocated

## 3.2.2. Manufacturing Industries and Construction (CRF Category 1.A.2)

### 3.2.2.1. Source Category Description

This subsector is composed of emissions from the combustion of fossil fuels by all mining, manufacturing and construction industries. The UNFCCC has assigned six categories under the Manufacturing Industries and Construction subsector, and these are presented separately in the following subsections.

In 2009, the Manufacturing Industries and Construction subsector accounted for 74.9 Mt (or 11%) of Canada's total GHG emissions, with a 16% (10.4 Mt) increase in overall

Table 3–3 Manufacturing Industries and Construction GHG Contribution

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Manufacturing Industries and Construction TOTAL (1.A.2)	64 500	69 500	72 300	68 800	70 400	79 400	72 600	74 900
Iron and Steel	5 270	6 330	5 720	5 770	5 230	5 930	4 750	4 030
Non-Ferrous Metals	3 260	3 220	3 260	3 290	3 250	3 510	3 680	3 120
Chemicals	8 220	10 000	7 890	7 040	8 140	8 570	7 270	7 570
Pulp, Paper and Print	14 400	12 000	10 300	7 920	6 660	6 670	5 160	4 510
Food Processing, Beverages and Tobacco <sup>1</sup>	IE	IE	IE	IE	IE	IE	IE	IE
Others	33 300	37 900	45 100	44 800	47 100	54 700	51 800	55 700
Cement	3 820	4 240	4 760	5 020	5 200	4 840	4 630	3 610
Mining	6 650	12 400	17 800	18 600	20 500	27 200	27 600	31 300
Construction	1 870	1 070	1 340	1 360	1 300	1 290	1 260	1 080
Other Manufacturing	21 000	20 200	21 200	19 800	20 100	21 400	18 200	19 700

Notes:

1. Note that Food Processing, Beverages and Tobacco emissions are included under Other Manufacturing.

IE = included elsewhere.

Totals may not add up due to rounding.



emissions since 1990 (refer to Table 3–3 for more details). Within the Manufacturing Industries and Construction subsector, 55.7 Mt (or 74%) of the GHG emissions are from the Others category. The Others category is made up of cement, mining, construction and other manufacturing activities. This category is followed by (in order of decreasing contributions) the Chemical Industries; Pulp, Paper and Print; Iron and Steel; and Non-ferrous Metals categories, at 7.6 Mt (or 10.1%), 4.5 Mt (or 6.0%), 4.0 Mt (or 5.4%), and 3.1 Mt (or 4.2%), respectively. Emissions from Food Processing, Beverages and Tobacco are included in the Other Manufacturing subcategory due to fuel-use data not being available at the appropriate level of disaggregation.

Industrial emissions resulting from fuel combustion for the generation of electricity or steam for sale have been assigned to the appropriate industrial subsector. Emissions generated from the use of fossil fuels as feedstocks or chemical reagents such as for use as metallurgical coke during the reduction of iron ore are reported under the Industrial Processes Sector to ensure that the emissions are not double counted.

### 3.2.2.2. Methodological Issues

Fuel combustion emissions for each category within the Manufacturing Industries and Construction subsector are calculated using the methodology described in Annex 2, which is consistent with an IPCC Tier 2 approach. Emissions generated from the use of transportation fuels (e.g. diesel and gasoline) are reported under the Transport subsector (Section 3.2.3, Transport (CRF Category 1.A.3)). Methodological issues specific to each manufacturing category are identified below.

#### Iron and Steel (CRF Category 1.A.2.a)

Canada has four integrated iron and steel facilities that manufacture all the coal-based metallurgical coke. All these facilities are structured in such a way that by-product gases from the integrated facilities (e.g. coke oven gas, blast furnace gas) are used in a variety of places throughout the facility (e.g. boilers, blast furnace, coke oven). As such, emissions from coke production are included in the Iron and Steel category. Since the plants are integrated, all the produced coke oven gas is used in the mills and reported in the RESD. Due to the way the fuel consumption is reported by the iron and steel industry, determining the amount of coke oven gas lost as fugitive emissions through flaring is difficult. However, Statistics Canada

indicates that the amount of fuel flared is included in the energy statistics, indicating that fugitive emissions are being captured as well.

Emissions associated with the use of metallurgical coke as a reagent for the reduction of iron ore in blast furnaces have been allocated to the Industrial Processes Sector.

#### Non-ferrous Metals (CRF Category 1.A.2.b)

All fuel-use data for this category were obtained from the RESD.

#### Chemicals (CRF Category 1.A.2.c)

Emissions resulting from fuels used as feedstocks are reported under the Industrial Processes Sector.

#### Pulp, Paper and Print (CRF Category 1.A.2.d)

Fuel-use data include industrial wood wastes and spent pulping liquors combusted for energy purposes. Emissions of CH<sub>4</sub> and N<sub>2</sub>O from the combustion of biomass are included in the pulp and paper industrial category. CO<sub>2</sub> emissions from biomass combustion are not included in totals but are reported separately in the UNFCCC CRF tables as a memo item.

#### Others (Other Manufacturing and Construction) (CRF Category 1.A.2.f)

This category includes the remainder of industrial sector emissions, including construction, cement, vehicle manufacturing, textiles, mining, food, beverage and tobacco sectors. Consumption of diesel fuels associated with on-site off-road vehicles in mining (which also includes oil and gas mining and extraction use of diesel) have been allocated to the Other Transportation category.

### 3.2.2.3. Uncertainties and Time-Series Consistency

The estimated uncertainty for the Manufacturing Industries and Construction subsector ranges from -3% to 6% for all gases and from -3% to 2% for CO<sub>2</sub> (ICF Consulting 2004).

The underlying fuel quantities and CO<sub>2</sub> emission factors have low uncertainty because they are predominantly commercial fuels, which have consistent properties and a more accurate tracking of quantity purchased for consumption. Coal CO<sub>2</sub> emission factor uncertainties

were recently updated with 95% confidence intervals (as discussed in Section 3.2.1.3), while new uncertainty values were identified for coke oven gas and biomass (spent pulping liquor and industrial fuelwood).

As stated in the Energy Industries subsector uncertainty discussion, additional expert elicitation is required to improve the CH<sub>4</sub> and N<sub>2</sub>O uncertainty estimates for some of the emission factor uncertainty ranges and probability density functions developed by the ICF Consulting study, since these assumptions were not reviewed by industry experts, owing to a lack of available time in the study's preparation.

The estimates for the Manufacturing Industries and Construction subsector have been prepared in a consistent manner over time using the same methodology.

#### 3.2.2.4. QA/QC and Verification

QC checks were done in a form consistent with IPCC Good Practice Guidance (IPCC 2000). Elements of a Tier 1 QC check include a review of the estimation model, activity data, emission factors, time-series consistency, transcription errors, reference material, conversion factors and unit labelling, as well as sample emission calculations.

Tier 1 QC checks were completed on the entire stationary combustion GHG estimation model, which included checks of emission factors, activity data and CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O estimates for the entire time series. No mathematical or reference errors were found during the QC checks. The data, methodologies and changes related to the QC activities are documented and archived in both paper and electronic form.

#### 3.2.2.5. Recalculations

*Activity Data:* The 2008 fuel-use data were revised by Statistics Canada and estimates were recalculated accordingly.

*Emission Factors:* Revisions to CO<sub>2</sub> emission factors for coal resulted in recalculations. These new emission factors were developed using significantly larger data sets across the entire time series and are expected to better reflect the properties and variability of coals consumed in Canada. The CO<sub>2</sub> emission factor for coke oven gas was compared with values obtained from Canadian industries and other reporting parties, and was found to be incorrect since its carbon content was assumed to be the same as coke. This was changed to match the default emission factor in the

Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997). It resulted in a change in the gaseous emission estimates for all years in CRF category 1.A.2.a – Iron and Steel. Updated CO<sub>2</sub> emission factors for biomass consumed in the Pulp, Paper and Print category were also applied. These resulted in minor recalculations (which are reported under memo items).

#### 3.2.2.6. Planned Improvements

As a continuous improvement activity, Environment Canada, Natural Resources Canada and Statistics Canada are working jointly to improve the underlying quality of the national energy balance and to further disaggregate fuel-use information.

### 3.2.3. Transport (CRF Category 1.A.3)

Transport-related emissions account for 28% of Canada's total GHG emissions (190 Mt—refer to Table 3–4 for more details). The greatest emission growth since 1990 has been observed in light-duty gasoline trucks (LDGTs) and heavy-duty diesel vehicles (HDDVs); this growth amounts to 104% (21.1 Mt) for LDGTs and 91% (18.2 Mt) for HDDVs. A long-term decrease in some Transport categories has also been registered: specifically, reductions in emissions from light-duty gasoline vehicles (LDGVs, i.e. cars), propane and natural gas vehicles, pipelines, heavy-duty gasoline vehicles (HDGVs), off-road gasoline, domestic aviation and railways, for a combined decrease of 6.8 Mt since 1990. Generally, emissions from the Transport subsector have increased 30% and have contributed the equivalent of 44% of the total overall growth in emissions observed in Canada.

#### 3.2.3.1. Source Category Description

This subsector comprises the combustion of fuel by all forms of transportation in Canada. The subsector has been divided into five distinct categories:

- Civil Aviation (Domestic Aviation);
- Road Transportation;
- Railways;
- Navigation (Domestic Marine); and
- Other Transportation (Off-road and Pipelines).

Table 3–4 Transport GHG Contribution

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Transport TOTAL (1.A.3.)	146 000	180 000	189 000	193 000	192 000	197 000	196 000	190 000
Civil Aviation (Domestic Aviation)	7 200	7 500	7 700	7 700	7 800	7 800	7 800	7 200
Road Transport	96 700	118 000	129 000	130 000	131 000	133 000	132 000	131 000
<i>Light-duty Gasoline Vehicles</i>	45 500	41 900	41 200	40 000	40 000	40 200	39 700	41 400
<i>Light-duty Gasoline Trucks</i>	20 300	36 300	41 400	42 500	42 700	43 000	42 600	41 300
<i>Heavy-duty Gasoline Vehicles</i>	7 440	5 460	6 620	6 540	6 650	6 810	6 840	6 990
<i>Motorcycles</i>	152	161	247	254	258	263	264	245
<i>Light-duty Diesel Vehicles</i>	469	466	569	574	580	617	652	663
<i>Light-duty Diesel Trucks</i>	702	1 660	1 830	1 930	1 960	2 020	2 020	1 940
<i>Heavy-duty Diesel Vehicles</i>	20 000	30 900	36 000	37 600	38 500	39 600	39 200	38 200
<i>Propane &amp; Natural Gas Vehicles</i>	2 200	1 100	860	720	790	830	880	780
Railways	7 000	7 000	6 000	6 000	6 000	7 000	7 000	7 000
Navigation (Domestic Marine)	5 000	5 100	6 600	6 400	5 800	6 100	5 900	5 100
Other Transport	30 000	43 000	41 000	43 000	41 000	43 000	43 000	40 000
<i>Off-road Gasoline</i>	7 800	8 800	9 000	8 300	7 600	8 100	7 400	7 600
<i>Off-road Diesel</i>	16 000	23 000	23 000	24 000	24 000	26 000	28 000	26 000
<i>Pipelines</i>	6 850	11 200	8 470	10 100	9 610	8 940	7 460	6 320

Note: Totals may not add up due to rounding.

### 3.2.3.2. Methodological Issues

Fuel combustion emissions associated with the Transport subsector are calculated using various adaptations of Equation A2-1 in Annex 2. However, because of the many different types of vehicles, activities and fuels, the emission factors are numerous and complex. In order to cope with the complexity, transport emissions are calculated using Canada's Mobile Greenhouse Gas Emission Model (MGEM) and the Aviation Greenhouse Gas Emission Model (AGEM). These models incorporate a version of the IPCC-recommended methodology for vehicle modelling (IPCC/OECD/IEA 1997) and are used to calculate all transport emissions with the exception of those associated with pipelines (energy necessary to propel oil or natural gas).

#### Civil Aviation (Domestic Aviation) (CRF Category 1.A.3.a)

This category includes all GHG emissions from domestic air transport (commercial, private, agricultural, etc.). In accordance with the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997), military air transportation emissions generated by consuming aviation turbo fuel are reported in the Other (Non-specified) subsector (CRF category 1.A.5). However,

military emissions generated by consuming aviation gasoline are included in this category (1.A.3.a) since the current data source for this fuel type does not disaggregate military from civil fuel use. Emissions from transport fuels used at airports for ground transport and stationary combustion applications are reported under Other Transportation. Emissions arising from flights that have their origin in Canada and destination in another country are considered to be international in nature and are reported separately under memo items – International Bunkers (CRF category 1.C.1.a).

The methodologies for the Civil Aviation category are fuel-type dependant. They follow a modified IPCC Tier 1 approach for aviation gasoline and a modified IPCC Tier 3 approach for aviation turbo fuel. Emissions estimates employ a mix of country-specific, plane-specific and IPCC default emission factors. The estimates attributed to aviation gasoline consumption are performed within MGEM, while those attributed to aviation turbo fuel are generated using AGEM. The estimates are calculated based on the reported quantities of aviation gasoline and turbo fuel consumed (IPCC/OECD/IEA 1997), as published in the RESD (Statistics Canada #57-003). Aviation fuel sales are reported

in the RESD; these figures represent aviation fuels sold to Canadian airlines, foreign airlines, public administration and commercial/institutional sectors. All aviation gasoline use is designated domestic, other than that reported under foreign airlines. (refer to Annex 2 for a description of the methodology).

### **Road Transportation (CRF Category 1.A.3.b)**

The methodology used to estimate road transportation GHG emissions is a detailed IPCC Tier 3 method (except for propane and natural gas vehicles, for which an IPCC Tier 1 method is followed), as outlined in IPCC/OECD/IEA (1997). MGEM disaggregates vehicle data and calculates emissions of CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O from all mobile sources except pipelines (refer to Annex 2 for a description of the methodology).

### **Railways (CRF Category 1.A.3.c)**

The procedure used to estimate GHG emissions from railways adheres to an IPCC Tier 1 methodology (IPCC/OECD/IEA 1997). Emission estimates are performed within MGEM. Fuel sales data from the RESD (Statistics Canada #57-003), reported under railways, are multiplied by country-specific emission factors (refer to Annex 2 for a description of the methodology).

### **Navigation (Domestic Marine) (CRF Category 1.A.3.d)**

This category includes all GHG emissions from domestic marine transport. Emissions arising from fuel sold to foreign marine vessels are considered to be international bunkers and are reported separately under memo items (CRF Category 1.C.1.b). Comprehensive activity data that would enable the accurate disaggregation of domestic and international marine emissions are being investigated (refer to Section 3.2.3.6, Planned Improvements).

The methodology complies with IPCC Tier 1 techniques (IPCC/OECD/IEA 1997), and emission estimates are performed within MGEM. Fuel consumption data from the RESD, reported as domestic marine, are multiplied by country-specific emission factors (refer to Annex 2 for a description of the methodology).

### **Other Transportation (CRF Category 1.A.3.e)**

This category comprises vehicles and equipment that are not licensed to operate on roads or highways, and includes GHG emissions from the combustion of fuel used to propel

products in long-distance pipelines.

### *Off-road Transport*

Non-road or off-road transport<sup>1</sup> (ground, non-rail vehicles and equipment) includes GHG emissions resulting from both gasoline and diesel fuel combustion. Vehicles in this category include farm tractors, logging skidders, tracked construction vehicles and mobile mining vehicles as well as off-road recreational vehicles. Equipment in this category includes residential and commercial lawn and garden combustion machines, generators, pumps and portable heating devices.

Industry uses a considerable amount of diesel fuel in non-road vehicles. The mining and construction industries (including coal, oil and natural gas drilling and extraction activities) both operate significant numbers of heavy non-road vehicles and are the largest diesel fuel users in the group.

Off-road vehicle emissions are calculated using a modified IPCC Tier 1 approach (IPCC/OECD/IEA 1997). For these estimates, emissions are based on country-specific emission factors and total fuel consumed (refer to Annex 2 for a description of the methodology).

### *Pipeline Transport*

Pipelines<sup>2</sup> represent the only non-vehicular transport in this sector. They use fossil-fuelled combustion engines to power motive compressors that propel their contents. The fuel used is primarily natural gas in the case of natural gas pipelines. Oil pipelines tend to use electric motors to operate pumping equipment, but some refined petroleum, such as diesel fuel, is also consumed as a backup during power failures.

The methodology employed is considered an IPCC Tier 2 sectoral approach, with country-specific emission factors. Fuel consumption data from the RESD, reported as pipelines, are multiplied by country-specific emission factors (refer to Annex 2 for a description of the methodology).

<sup>1</sup> Referred to as non-road or off-road vehicles. The terms “non-road” and “off-road” are used interchangeably.

<sup>2</sup> Consisting of both oil and gas types.

### 3.2.3.3. Uncertainties and Time-Series Consistency

The Transport subsector employs a Monte Carlo uncertainty analysis, established upon the recommendations and results reported in *Quantitative Assessment of Uncertainty in Canada's National GHG Inventory Estimates for 2001* (ICF Consulting 2004). Several modifications were introduced into the original model in order to more accurately reflect uncertainties in the latest Transport subsector emissions estimates.

Modifications to the original assessment include the addition of biofuel emission factor uncertainties, based on the assumption of similarities in emission control technologies between conventional transport fuels and biofuels. Biofuel activity data uncertainties were based on expert judgement. The uncertainties associated with the CH<sub>4</sub> (Lister and Norman 2003) and N<sub>2</sub>O (IPCC 2006) emission factors pertaining to aviation turbo fuel have been updated from those recommended in the ICF Consulting report to better reflect the improvements made by implementing AGEM. Uncertainties relating to the updated on-road CH<sub>4</sub> and N<sub>2</sub>O emission factors described in Section 3.2.3.5 are also incorporated into the latest Monte Carlo simulation. Additionally, a thorough verification of the 2004 ICF Consulting report revealed a number of discrepancies in referenced uncertainty ranges. In these instances, the discrepancy was corrected to coincide with the original reference.

#### Transport Subsector Fossil Fuel Combustion

The Transport subsector comprises 1) the mobile sources of transport, including on-road and off-road vehicles, railways, civil aviation and navigation; and 2) pipeline transport. The overall uncertainty of the 2009 estimates for the mobile Transport subsector (not including pipelines) was estimated to be between -1.9% and +4.7%.

The uncertainty for Transport fuel combustion CO<sub>2</sub> emissions was  $\pm 0.4\%$ . In contrast, and similar to the stationary fuel combustion sources, CH<sub>4</sub> and N<sub>2</sub>O emission uncertainty ranges were two to three orders of magnitude greater than that of CO<sub>2</sub>. Hence, the overall uncertainty for the mobile Transport subsector reflects the predominance of CO<sub>2</sub> in total GHG emissions.

#### Emissions from Civil Aviation (Domestic Aviation)

The uncertainty associated with overall emissions from

domestic aviation was estimated to be within the range of -1% to +5%. This implied that the source category was more likely underestimated than overestimated. The high uncertainties associated with jet kerosene CH<sub>4</sub> (-50% to +50%) and N<sub>2</sub>O emission factors (-70% to +150%) resulted in a downward bias on the inventory. These effects were somewhat reduced by the large contribution of jet kerosene CO<sub>2</sub> emissions and its comparatively low emission factor uncertainty. The Civil Aviation category only contributed approximately 4% to total Transport GHG emissions and therefore did not greatly influence overall uncertainty levels.

#### Emissions from Road Transportation

The uncertainty related to the overall emissions from on-road vehicles was estimated to be within the range of  $\pm 1\%$ , driven primarily by the relatively low uncertainties in gasoline and diesel activity data and their related CO<sub>2</sub> emissions. Conversely, the high uncertainties associated with CH<sub>4</sub> and N<sub>2</sub>O emissions, as well as biofuel activity data, did not greatly influence the analysis due to their comparatively minor contributions to the inventory. Approximately 69% of the Transport subsector's GHG emissions were attributable to on-road transportation. Accordingly, the Transport subsector's relatively low inventory uncertainty is justified through the results of the Road Transportation category uncertainty analysis.

#### Emissions from Railways

The uncertainty associated with emissions from rail transport was estimated to be between -11% and +32%, indicating that this category was potentially underestimated. The greatest influence was exerted by the high N<sub>2</sub>O emission factor uncertainty (-90% to +900%), whereas the relatively low uncertainties in diesel activity data and CO<sub>2</sub> emission factors contributed very little. It is important to note that railway emissions only accounted for approximately 4% of the Transport subsector GHG inventory and therefore did not greatly influence the overall uncertainty results.

#### Emissions from Navigation (Domestic Marine)

The uncertainty associated with emissions from the domestic marine source category ranged from -8% to +16%, suggesting that GHGs were potentially underestimated. The high N<sub>2</sub>O emission factor uncertainty (-90% to +900) represented the largest contribution to uncertainty, while CO<sub>2</sub> emission factor uncertainties were insignificant. Since



domestic marine emissions only made up 3% of the Transport subsector GHG inventory, they did not substantially alter the overall uncertainty results.

### Emissions from Other Transportation (Off-road)

The Off-road Transport subcategory includes both off-road gasoline and off-road diesel consumption. The uncertainty associated with the off-road mobile transport sources ranged from –8% to +25%, indicating that the 2011 submission likely underestimates total emissions from this subcategory. Consistent with the inventory estimation methodology for this source category, off-road diesel fuel consumption is calculated from the on-road diesel fuel consumption residual, and likewise for off-road gasoline consumption. Consequently, activity data uncertainties from road transportation were employed in the off-road uncertainty analysis and did not greatly contribute to the results mentioned above since they were relatively low. Of greater influence was the N<sub>2</sub>O emission uncertainty for gasoline and diesel fuels (–90% to +900%) which indicated a downward bias in the GHG estimate. Approximately 17% of the Transport subsector's GHG emissions were attributable to off-road transportation and therefore had a significant effect on the overall uncertainty analysis.

### Summary

Generally, for the Transport subsector, the ICF Consulting study incorporated uncertainty values for CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O emission factors from two other reports—McCann (2000) and SGA Energy Ltd. (2000). The ICF Consulting study included values determined in these reports, along with limited expert elicitations addressing the uncertainty of the activity data contributing to the Transport subsector estimates within its Monte Carlo analysis. A number of incremental improvements have been incorporated into the original analysis, as described in the opening paragraphs of Section 3.2.3.3.

Some of the weaker components of the uncertainty analysis surround the acquisition of expert opinions on non-fuel-quantity-type activity estimates (e.g. vehicle populations, kilometres travelled, motorcycle numbers). Although it was suggested that the vehicle population data supplied by an outside consultant to Environment Canada are 100% accurate, this is unlikely, and there are indications that compilation errors exist. Presently, inventory practitioners are conducting a study to re-establish the time series for the Canadian fleet. The current fleet uncertainty will intro-

duce only marginal errors in a fuel-constrained model, but it has considerable impact on the attribution of that fuel to specific vehicle types.

### 3.2.3.4. QA/QC and Verification

Tier 1 QC checks as elaborated in the framework for the QA/QC plan (see Annex 6) were performed on all categories in Transport, not just those designated as “key.” No significant mathematical errors were found. The QC activities are documented and archived in paper and electronic form.

In addition, certain verification steps were performed during the model preparation stage. Since MGEM uses national fuel data defined by type and region combined with country-specific emission factors, primary scrutiny is applied to the vehicle population profile, as this dictates the fuel demand per vehicle category and, hence, emission rates and quantities. Interdepartmental partnerships have been developed among Environment Canada, Transport Canada and Natural Resources Canada to facilitate the sharing of not only raw data but also derived information such as vehicle populations, fuel consumption ratios (FCRs), vehicle kilometres travelled (VKTs) and kilometre accumulation rates (KARs). This broader perspective fosters a better understanding of actual vehicle use and subsequently should promote better modelling and emission estimating. With support from Transport Canada and Natural Resources Canada, Statistics Canada publishes the *Canadian Vehicle Survey* (CVS), a quarterly report that provides both vehicle population and VKTs in aggregated regional classes. It provides alternative interpretation of provincial registration files and can therefore corroborate the commercially available data sets mentioned above. Unfortunately, the resolution necessary for emission modelling is unavailable from the CVS, and therefore it cannot replace the annually purchased data sets.

### 3.2.3.5. Recalculations

Transportation estimates were revised for the entire time series due to the following factors:

*Statistics Canada fuel consumption data:* A revised data set for 2008 energy consumption was released, resulting in minor adjustments of estimates for that year.

*Statistics Canada taxed fuel sales data:* A revised data set for 2008 was received. Minor adjustments resulted for that year.

*Implementation of AGEM for aviation turbo fuel:* This impacted the domestic/international split for aviation turbo fuel due to the increased clarity of flight-by-flight data for applying the definitions of domestic and international aviation. Assumptions that were needed previously to estimate the allocation of turbo fuel between domestic and international are no longer needed. The result is increased emissions allocated to the domestic component in the earlier years of the time series and a shift to decreased allocation in the later years.

*Emission factors:* Twelve on-road CH<sub>4</sub> and N<sub>2</sub>O emission factors were updated based on a review of the most recent emissions testing results, sourced primarily from Environment Canada's laboratories, leading to slight adjustments in estimates across the entire time series. Aviation turbo fuel CH<sub>4</sub> and N<sub>2</sub>O emission factors were updated in conjunction with AGEM.

*Kilometre accumulation rates (KARs):* A revised KAR data set, averaged by vehicle age and vehicle class, was integrated into MGEM. Minor adjustments resulted for all inventory years.

*Fuel consumption ratios (FCRs):* A revised light-duty vehicle FCR data set, covering model years 1966–2009, was incorporated into MGEM, leading to slight adjustments in estimates across the entire time series.

*Biofuels in National Energy Balance:* For many years, Statistics Canada was ambiguous about the inclusion of biofuels in their annually reported National Energy Balance (RES-D – 57-003). In order to ensure completeness of the NIR, known quantities of the biofuels consumed were modelled for emissions over and above the amounts of base fuels reported in the RES-D. This year, however, Statistics Canada has stated that, in fact, ethanol has always been included in the RES-D. As a result of this new understanding, known quantities of ethanol are first removed from the quantity of base fuel (gasoline) reported in the RES-D prior to emission modelling, and subsequently, all historic estimates of Transport subsector emissions from gasoline consumption reflect this reduction (1990–2008). The biofuel-based emissions are then reported as memo items.

### 3.2.3.6. Planned Improvements

The transportation model (MGEM) was upgraded in 2010–2011 and continuously evolves to accommodate an increasing number of higher-resolution data sets being made available through partnerships and reporting.

Future improvements will concentrate on the following:

- The development of better on-road activity data. Interdepartmental partnerships have been established among Environment Canada, Transport Canada and Natural Resources Canada to develop common and better on-road activity data. The decoding of vehicle identification numbers (VINs) progressed during the last year but there is still work to be done before it can be introduced into the inventory. This work will hopefully allow us to use provincial registration files and obtain a better representation of the Canadian fleet. Fuel consumption ratios (FCRs) are also being evaluated to make sure that what we use represents the Canadian situation accurately.
- The development of an improved model to estimate marine emissions. Interdepartmental partnerships have been established between Environment Canada, Transport Canada and the marine industry in order to develop a bottom-up marine model. Progress has been made; however, the model is currently in the developmental stage. Environment Canada's focus for this model is to better represent the disaggregation between domestic and international marine traffic.

## 3.2.4. Other Sectors (CRF Category 1.A.4)

### 3.2.4.1. Source Category Description

The Other Sectors subsector consists of three categories: Commercial/Institutional, Residential and Agriculture/Forestry/Fisheries. Emissions consist primarily of fuel combustion related to space and water heating. Emissions from the use of transportation fuels in these categories are allocated to Transport (Section 3.2.3).

Biomass<sup>3</sup> combustion is a significant source of emissions in the residential sector, and CH<sub>4</sub> and N<sub>2</sub>O emissions are included in the subsector estimates. However, CO<sub>2</sub> emissions from biomass combustion are reported separately in the CRF tables as memo items and are not included in Energy Sector totals. This method is consistent with the treatment of biomass in the Pulp, Paper and Print category.

In 2009, the Other Sectors subsector contributed 78.7 Mt (or 11.4%) of Canada's total GHG emissions, with an overall growth of about 10% (7.2 Mt) since 1990. Within the Other Sectors subsector, residential emissions contributed about 41 Mt (or 52%), followed by a 36 Mt (or 46%) contribution from the Commercial/Institutional category, which also in-

3 Typically firewood.



Table 3–5 Other Sectors GHG Contribution

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Other Sectors TOTAL (1.A.4)	71 600	80 500	82 900	80 600	75 300	80 900	80 600	78 700
Commercial/Institutional	25 700	33 100	37 700	36 700	33 400	34 900	35 200	36 000
<i>Commercial and Other Institutional</i>	23 600	30 600	35 400	34 400	31 400	32 700	33 000	33 700
<i>Public Administration</i>	1970	2230	2040	2040	1810	2000	1910	2060
Residential	43 000	45 000	43 000	42 000	40 000	44 000	43 000	41 000
Agriculture/Forestry/Fisheries	2 390	2 540	2 090	1 970	1 910	2 240	2 260	2 050
<i>Forestry</i>	60	70	130	120	100	100	110	50
<i>Agriculture</i>	2300	2500	1900	1800	1800	2100	2100	2000

Note: Totals may not add up due to rounding.

cludes emissions from the public administration sector (i.e. federal, provincial and municipal establishments). Since 1990, GHG emissions have grown by about 40% in the Commercial/Institutional category, while GHG emissions in the Residential category have decreased by about 6.5%. Refer to Table 3–5 for additional details. Additional trend discussion for the Other Sectors subsector is presented in the Emission Trends chapter (Chapter 2).

### 3.2.4.2. Methodological Issues

Emissions from these source categories are calculated consistently according to the methodology described in Annex 2, which is considered to be an IPCC Tier 2 approach, with country-specific emission factors. Methodological issues specific to each category are described below. Emissions from the combustion of transportation fuels (e.g. diesel and gasoline) are all allocated to the Transport subsector.

#### Commercial/Institutional (CRF Category 1.A.4.a)

Emissions are based on fuel-use data reported as commercial and public administration in the RESD.

#### Residential (CRF Category 1.A.4.b)

Emissions are based on fuel-use data reported as residential in the RESD. The methodology for biomass combustion from residential firewood is detailed in the CO<sub>2</sub> Emissions from Biomass section (Section 3.4.2); although CO<sub>2</sub> emissions are not accounted for in the national residential GHG total, the CH<sub>4</sub> and N<sub>2</sub>O emissions are reported here.

#### Agriculture/Forestry/Fisheries (CRF Category 1.A.4.c)

This source category includes emissions from stationary fuel combustion in the agricultural and forestry industries. However, emission estimates are included for the agriculture and forestry portion only. Fishery emissions are reported typically under either the Transportation subsector or the Other Manufacturing (i.e. food processing) category. Mobile emissions associated with this category were not disaggregated and are included as off-road or marine emissions reported under Transport (3.2.3). Emissions from on-site machinery operation and heating are based on fuel-use data reported as agriculture and forestry in the RESD.

### 3.2.4.3. Uncertainties and Time-Series Consistency

The estimated uncertainty for the Other Sectors subsector ranges from -4% to 41% for all gases and from -3% to 2% for CO<sub>2</sub> (ICF Consulting 2004).

The underlying fuel quantities and CO<sub>2</sub> emission factors have low uncertainties, since they are predominantly commercial fuels, which have consistent properties and accurate tracking. Although the non-CO<sub>2</sub> emissions from biomass combustion contributed only 5% to the total Residential category, its CH<sub>4</sub> (-90% to 1500%) and N<sub>2</sub>O (-65% to 1000%) uncertainties are high due to the uncertainty associated with their emission factors. As stated in the Energy Industries subsector, additional expert elicitation is required to improve the CH<sub>4</sub> and N<sub>2</sub>O uncertainty estimates for some of the emission factor uncertainty ranges and probability density functions developed by the ICF

Consulting study, since insufficient time was available to have these assumptions reviewed by industry experts.

These estimates are consistent over the time series.

### 3.2.4.4. QA/QC and Verification

The Other Sectors subsector underwent Tier 1 QC checks in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). No mathematical or referencing errors were observed during the QC checks, while minor data errors were discovered and corrected. The data, methodologies, and changes related to the QC activities are documented and archived in both paper and electronic form.

### 3.2.4.5. Recalculations

The 2008 fuel-use data were revised by Statistics Canada, and revised emission factors (coal and biomass) were applied resulting in recalculations to the entire time series. Residential biomass consumption values were also revised following a review of the model and subsequent correction to an assumption in provincial allocation.

### 3.2.4.6. Planned Improvements

Future improvement plans for the Other Sectors subsector include a review of the activity data used by the residential biomass model.

## 3.2.5. Other: Energy–Fuel Combustion Activities (CRF Category 1.A.5)

The UNFCCC reporting guidelines assign military fuel combustion to this subsector. The introduction of AGEM in this submission allowed us to report military turbo fuel air transportation emissions under this category. However, military emissions generated by consuming aviation gasoline are included under Civil Aviation (CRF Category 1.A.3.a), since the current data source for this type of fuel consolidates military and civil fuel use. As in previous submissions, emissions related to military vehicles have been included in the Transport subsector, whereas stationary military fuel use has been included under the Commercial/Institutional category (Section 3.2.4) due to fuel data allocation in the RESD (Statistics Canada #57-003). This is a small source of emissions.

## 3.3. Fugitive Emissions (CRF Category 1.B)

Fugitive emissions from fossil fuels are intentional or unintentional releases of GHGs from the production, processing, transmission, storage and delivery of fossil fuels.

Released gas that is combusted before disposal (e.g. flaring of natural gases at oil and gas production facilities) is considered a fugitive emission. However, if the heat generated during combustion is captured for use (e.g. heating) or sale, then the related emissions are considered fuel combustion emissions.

Table 3–6 Fugitive GHG Contribution

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Fugitive Emissions from Fuels (1.B)	42 100	63 000	64 000	63 100	64 300	62 800	62 300	60 700
Solid Fuels—Coal Mining (1.B.1)	2 000	900	700	700	700	800	800	700
Oil and Natural Gas (1.B.2)	40 200	62 100	63 400	62 400	63 600	62 100	61 500	60 000
<i>a. Oil<sup>1</sup></i>	4 190	5 440	5 940	5 650	5 730	5 830	5 550	5 530
<i>b. Natural Gas<sup>1</sup></i>	11 400	17 700	18 700	19 200	19 700	19 700	19 700	19 400
<i>c. Venting and Flaring<sup>2</sup></i>	24 600	38 900	38 800	37 600	38 100	36 600	36 300	35 100
<i>Venting</i>	20 200	33 500	33 100	32 100	32 100	31 300	30 700	28 700
<i>Flaring</i>	4 400	5 400	5 600	5 500	6 000	5 300	5 500	6 400

Notes:

1. All other fugitives except venting and flaring.

2. Both oil and gas activities.

Totals may not add up due to rounding.

The two categories considered in the inventory are fugitive releases associated with solid fuels (coal mining and handling) and releases from activities related to the oil and natural gas industry.

In 2009, the Fugitives category accounted for about 60.7 Mt (or 8.8%) of Canada's total GHG emissions, with about a 44% growth in emissions since 1990. Between 1990 and 2009, fugitive emissions from oil and natural gas increased 49% to 60.0 Mt, and those from coal decreased by about 1.3 Mt from 2 Mt in 1990. The oil and gas production, processing, transmission and distribution activities contributed 99% of the fugitive emissions. Refer to Table 3–6 for more details.

### 3.3.1. Solid Fuels (CRF Category 1.B.1)

#### 3.3.1.1. Source Category Description

Coal in its natural state contains varying amounts of CH<sub>4</sub>. In coal deposits, CH<sub>4</sub> is either trapped under pressure in porous void spaces within the coal formation or adsorbed to the coal. The pressure and amount of CH<sub>4</sub> in the deposit vary depending on the grade, the depth and the surrounding geology of the coal seam. During coal mining, post-mining activities and coal-handling activities, the natural geological formations are disturbed, and pathways are created that release the pressurized CH<sub>4</sub> to the atmosphere. As the pressure on the coal is lowered, the adsorbed CH<sub>4</sub> is released until the CH<sub>4</sub> in the coal has reached equilibrium with the surrounding atmospheric conditions.

Emissions from mining activities are from exposed coal surfaces, coal rubble and the venting of CH<sub>4</sub> from within the deposit. Post-mining activities such as preparation, transportation, storage and final processing prior to combustion also release CH<sub>4</sub>.

Fugitive emissions from solid fuel transformation (e.g. fugitive losses from the opening of metallurgical coking oven doors) are not estimated owing to a lack of data. Other sources of solid fuel transformation emissions are not known. These sources are thought to be insignificant.

#### 3.3.1.2. Methodological Issues

In the early 1990s, King (1994) developed an inventory of fugitive emissions from coal mining operations, which is the basis for the coal mining fugitive emissions estimated.

Emission factors were calculated by dividing the emission estimates from King (1994) by the appropriate coal production data.

The method used by King (1994) to estimate emission rates from coal mining (emission factors in Annex 3) was based on a modified procedure from the Coal Industry Advisory Board. It consists of a hybrid of IPCC Tier 3– and Tier 2–type methodologies, depending on the availability of mine-specific data. Underground mining activity emissions and surface mining activity emissions were separated, and both include post-mining activity emissions. A detailed description of the methodology is located in Annex 3: Additional Methodologies.

#### 3.3.1.3. Uncertainties and Time-Series Consistency

The CH<sub>4</sub> uncertainty estimate for fugitive emissions from coal mining is estimated to be in the range of -30% to 130% (ICF Consulting 2004). The production data are known to a high degree of certainty ( $\pm 2\%$ ). On the other hand, a very significant uncertainty (-50% to 200%) was estimated for the emission factors. It is our view that further expert elicitation is required to validate assumptions made by the study in the development of the probability density functions and uncertainty ranges of emission factors and activity data from surface and underground mining activities. IPCC default uncertainty values were assumed for Canada's country-specific emission factors, and these will need to be reviewed. The use of IPCC default values will not result in a representative uncertainty estimate when country-specific information is used.

#### 3.3.1.4. QA/QC and Verification

The CH<sub>4</sub> emissions from coal mining were identified as a key category and underwent Tier 1 QC checks in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). Checks included a review of activity data, time-series consistency, emission factors, reference material, conversion factors and units labelling, as well as sample emission calculations. No mathematical errors were found during the QC checks. The data and methods related to the QC activities are documented and archived in paper and electronic form.

### 3.3.1.5. Recalculations

There were no recalculations required for fugitive emissions from coal mining.

### 3.3.1.6. Planned Improvements

In the long-term, a comprehensive study of coal mining in Canada is planned, with the goal of improving aspects of the model, such as new emission factors.

## 3.3.2. Oil and Natural Gas (CRF Category 1.B.2)

### 3.3.2.1. Source Category Description

The Oil and Natural Gas category of fugitive emissions includes emissions from oil and gas production, processing, oil sands mining, bitumen extraction, *in-situ* bitumen production, heavy oil/bitumen upgrading, petroleum refining, natural gas transmission and natural gas distribution. Fuel combustion emissions from facilities in the oil and gas industry (when used for energy) are included under the Petroleum Refining and the Manufacture of Solid Fuels and Other Energy Industries categories (Section 3.2.1).

The Oil and Natural Gas source category has three main components: Upstream Oil and Gas (UOG), Oil Sands / Bitumen, and Downstream Oil and Gas.

### Upstream Oil and Gas

Upstream oil and gas (UOG) includes all fugitive emissions from exploration, production, processing and transmission of oil and natural gas, excluding those from oil sands mining, bitumen extraction and upgrading activities. Emissions may be the result of designed equipment leakage (bleed valves, fuel gas-operated pneumatic equipment), imperfect seals on equipment (flanges and valves), use of natural gas to produce hydrogen, and accidents, spills and deliberate vents.

The sources of emissions have been divided into major groups:

*Oil and Gas Well Drilling and Associated Testing:* Oil and gas well drilling is a minor emission source. The emissions are from drill stem tests, release of entrained gas in drilling fluids and volatilization of invert drilling fluids.

*Oil and Gas Well Servicing and Associated Testing:* Well servicing is also a minor emission source. The emissions are

mainly from venting, flaring and fuel combustion, which is included in the Stationary Combustion subsector. Venting results from conventional service work, such as the release of solution gas from mud tanks and blow down treatment for natural gas wells. It is assumed that there is no significant potential for fugitive emissions from leaking equipment. Fugitive emissions from absolute open flow tests are assumed to be negligible.

*Natural Gas Production:* Natural gas is produced exclusively at gas wells or in combination with conventional oil, heavy oil and crude bitumen production wells with gas conservation schemes. The emission sources associated with natural gas production are wells, gathering systems, field facilities and gas batteries. The majority of emissions result from equipment leaks, such as leaks from seals; however, venting from the use of fuel gas to operate pneumatic equipment and line-cleaning operations are also significant sources.

*Light/Medium Oil Production:* This type of production is defined by wells producing light- or medium-density crude oils (i.e. density < 900 kg/m<sup>3</sup>). The emissions are from the wells, flow lines and batteries (single, satellite and central). The largest sources of emissions are the venting of solution gas and evaporative losses from storage facilities.

*Heavy Oil Production:* Heavy oil is defined as having a density above 900 kg/m<sup>3</sup>. Production of this viscous liquid requires a special infrastructure. There are generally two types of heavy oil production systems: primary and thermal. The emission sources from both types are wells, flow lines, batteries (single and satellite) and cleaning plants. The largest source is venting of casing and solution gas.

*In-situ Bitumen Production:* Crude bitumen is a highly viscous, dense liquid that cannot be removed from a well using primary production means. Enhanced heavy oil recovery is required to recover the hydrocarbons from the formation, including primary production (or cold flow), cyclic steam stimulation (CSS), steam-assisted gravity drainage (SAGD); and experimental methods such as toe-to-heel air injection (THAI), vapour extraction process (VAPEX) and combustion overhead gravity drainage (COGD). The sources of emissions are wells, flow lines, satellite batteries and cleaning plants. The main source of emissions is the venting of casing gas.

*Natural Gas Processing:* Natural gas is processed before entering transmission pipelines to remove water vapour, contaminants and condensable hydrocarbons. There are

four different types of natural gas plants: sweet plants, sour plants that flare waste gas, sour plants that extract elemental sulphur, and straddle plants. Straddle plants are located on transmission lines and recover residual hydrocarbons. They have a similar structure and function and are considered in conjunction with gas processing. The largest source of emissions is equipment leaks.

*Natural Gas Transmission:* Virtually all of the natural gas produced in Canada is transported from the processing plants to the gate of the local distribution systems by pipelines. The volumes transported by truck are insignificant and assumed to be negligible. The gas transmission system emission sources are from equipment leaks and process vents. Process vents include activities such as compressor start-up and purging of lines during maintenance. The largest source of emissions is equipment leaks.

*Liquid Product Transfer:* The transport of liquid products from field processing facilities to refineries or distributors produces emissions from the loading and unloading of tankers, storage losses, equipment leaks and process vents. The transport systems included are liquefied petroleum gas (LPG) (by both surface transport and high-vapour-pressure pipeline systems), pentane-plus systems (by both surface transport and low-vapour-pressure pipeline systems) and crude-oil pipeline systems.

*Accidents and Equipment Failures:* Fugitive emissions can result from human error or extraordinary equipment failures in all segments of the conventional upstream oil and gas (UOG) industry. The major sources are emissions from pipeline ruptures, well blowouts and spills. Emissions from the disposal and land treatment of spills are not included owing to insufficient data.

*Surface Casing Vent Blows and Gas Migration:* At some wells, fluids will flow into the surface casing from the surrounding formation. Depending on the well, the fluids will be collected, sealed in the casing, flared or vented. The vented emissions are estimated in this section. At some wells, particularly in the Lloydminster (Alberta) region, gas may migrate outside of the well, either from a leak in the production string or from a gas-bearing zone that was penetrated but not produced. The emissions from the gas flowing to the surface through the surrounding strata have been estimated.

## Oil Sands / Bitumen

This component includes emissions from oil sand open pit mining operations and heavy oil/bitumen upgrading to produce synthetic crude oil and other derived products for sale. Fugitive emissions are primarily from hydrogen production, flue gas desulphurization (FGD), venting and flaring activities, storage and handling losses, fugitive equipment leaks, and CH<sub>4</sub> from the open mine surfaces and from methanogenic bacteria in the mine tailings settling ponds.

Emissions related to methanogenic bacteria in the tailings ponds continue to be studied by the operators. It is believed that with the planned implementation of new bitumen recovery techniques, the lighter hydrocarbons in the waste streams of the current processes will be reduced, and the emissions will be correspondingly lowered.

## Downstream Oil and Gas

Downstream oil and gas includes all fugitive emissions from the production of refined petroleum products and the distribution of natural gas to end consumers. The emissions have been divided into two major groups:

*Petroleum Refining:* There are three main sources of fugitive emissions from refineries: process, unintentional fugitive and flaring. Process emissions result from the production of hydrogen as well as from process vents. Unintentional fugitive emissions are the result of equipment leaks, wastewater treatment, cooling towers, storage tanks and loading operations. Flaring emissions result from the combustion of hazardous waste gas streams (such as acid gas) and fuel gas (or natural gas). GHG emissions from the combustion of fuel for energy purposes are reported under the Energy Industries subsector.

*Natural Gas Distribution:* The natural gas distribution system receives high-pressure gas from the gate of the transmission system and distributes this through local pipelines to the end user. The major emission sources are fugitive emissions from main and service pipelines and meter/regulator stations, which account for about 42% and 33% of emissions, respectively.

### 3.3.2.2. Methodological Issues

#### Upstream Oil and Gas

Fugitive emission estimates from the UOG industry are based on the Canadian Association of Petroleum Produc-



ers' (CAPP) study of the industry: *A National Inventory of Greenhouse Gas (GHG), Criteria Air Contaminant (CAC) and Hydrogen Sulphide (H<sub>2</sub>S) Emissions by the Upstream Oil and Gas Industry* (CAPP 2005a). The complete methodology is presented in volumes 1, 3 and 5 of the report.

For the year 2000, emissions were identified at the facility level for over 5000 facilities. These estimates were then extrapolated to approximately 370 000 primary sources from flaring, venting, equipment leaks, formation CO<sub>2</sub> venting, storage losses, loading/unloading losses and accidental releases. Natural gas systems, gas production and gas processing are considered to be part of the upstream petroleum industry, and the emissions for these sections were included.

A multitude of data were collected and used in the study. These included activity data from the facilities, such as process and equipment data. Emission factors were obtained from a variety of sources: published reports, such as the U.S. EPA (1995a, 1995b); equipment manufacturers' data; observed industry values; measured vent rates; simulation programs; and other industry studies. A list of data and emission factors can be found in Volume 5 of the UOG report (CAPP 2005a).

The 1990–1999 and 2001–2009 fugitive emissions were estimated using annual industry activity data from conventional UOG production and the 2000 emission results. The 1990–1999 estimates and method are presented in Volume 1 of the UOG report. A consistent UOG estimation model (hereafter referred to as the UOG model) for 2001 and onwards was developed by Clearstone Engineering Ltd. for use in estimating annual national- and provincial-level GHG estimates. The emissions for both time spans were estimated using emission data from the year 2000 together with the annual production and activity data for the relevant years. A detailed description of the methodology can be found in the UOG report (CAPP 2005a) and the UOG model report (CAPP 2005b). A summary of the estimation method of the UOG model is represented in Annex 3.

### Natural Gas Transmission

Fugitive emissions from natural gas transmission for 1990–1996 are from the conventional UOG industry study, *CH<sub>4</sub> and VOC Emissions from the Canadian Upstream Oil and Gas Industry* (CAPP 1999). This study is considered to follow a rigorous IPCC Tier 3 approach in estimating GHG emissions. Fugitive emission estimates for 1997 onward were estimated based on length of natural gas pipeline and

leakage rates, as developed based on the results from the original study. The methodology can be found in Annex 3.

### Oil Sands / Bitumen

Fugitive GHG emissions from oil sands mining, bitumen extraction, heavy oil/bitumen upgraders and integrated cogeneration facilities are from the bitumen study, *An Inventory of GHGs, CACs, and H<sub>2</sub>S Emissions by the Canadian Bitumen Industry: 1990 to 2003* (CAPP 2006). The bitumen study is a compilation of GHG emissions from the following companies: Suncor Energy Inc., Syncrude Canada Ltd., Shell Canada Ltd. and Husky Energy Inc. Methods used to estimate fugitive emissions from in situ bitumen extraction are from CAPP's UOG study (CAPP 2005a) (see Section 3.3.2.1).

Facilities' inventories were reviewed by Clearstone Engineering Ltd. to ensure that each facility's estimates were complete, accurate and transparent. Issues were corrected by facilities, and the final bitumen inventory was compiled by Clearstone Engineering Ltd. In general, the IPCC Tier 3 approach was used by each operator to develop a bottom-up approach in estimating GHG emissions. Where gaps existed, estimates were prepared by Clearstone Engineering Ltd. and provided to each operator for review. QA/QC and an uncertainty analysis following the IPCC Good Practice Guidance (IPCC 2000) were also included in the study.

A bitumen estimation model (hereafter referred to as the bitumen model) was developed to allow annual updating of fugitive emissions from oil sands mining and bitumen/heavy oil upgrading activities from 2004 onwards. The bitumen model was developed based on relevant parameters and results from the original bitumen study along with annual activity data. The activity data required by the model are published in the following reports: *Mineable Alberta Oil Sands Annual Statistics ST43* from the Energy Resources Conservation Board (ERCB 2010a) and the National Energy Board's (NEB 1998–2009) online statistics: *Estimated Production of Canadian Crude Oil and Equivalent*. These data are updated annually and used to estimate GHG emissions. Refer to both the bitumen study (CAPP 2006) and the bitumen model (Environment Canada 2007) for a detailed description of the methodology. A summary of the estimation method of the bitumen model is also presented in Annex 3.

Emissions for new oil sands facilities such as the CNRL Horizon Mine and Upgrader, Nexen Long Lake Upgrader and Petro-Canada Fort Hills Mine have been estimated using



activity data from ERCB (2010a) and emission factors from similar facilities.

## Downstream Oil and Gas Production

Fugitive emissions from refineries are based on the Canadian Petroleum Products Institute (CPPI) study, *Economic and Environmental Impacts of Removing Sulphur from Canadian Gasoline and Distillate Production* (CPPI 2004). Refer to the CPPI report for full details on the study. Historical fuel, energy and emission data were gathered from the Canadian Industrial Energy End-Use Data Analysis Centre (CIEEDAC) and directly from refineries for the years 1990 and 1994–2002. Fugitive, venting and flaring emissions for the years 1991–1993 were interpolated, and emissions for 2003–2009 were extrapolated, using data in the CPPI report and the petroleum refinery energy consumption and production data from the RESD published by Statistics Canada (#57-003). A detailed description of the methodology used to estimate emissions from 1991 to 1993 and from 2003 onward can be found in Annex 3.

## Natural Gas Distribution

The emission estimates were derived from a study prepared for the Canadian Gas Association (CGA 1997). The study estimated the emissions from the Canadian gas pipeline industry for the years 1990 and 1995.

Emissions in the study were calculated based upon emission factors from the U.S. EPA, other published sources and engineering estimates.

The activity data in the study were obtained from published sources and from specialized surveys of gas distribution system companies. The surveys obtained information on schedules of equipment, operation parameters of equipment, pipeline lengths used in the Canadian distribution system, etc.

In the year 2000, the Gas Research Institute (GRI) reviewed

and revised the 1997 CGA study, with more accurate and better substantiated data for station vents (GRI 2000).

General emission factors were developed for the distribution system based on the study data (CGA 1997; GRI 2000) and gas distribution pipeline distances by province provided by Statistics Canada.<sup>4</sup> Statistics Canada was only able to provide pipeline distance data up to 2008. The 2009 data were projected based on average change in distance in previous years. More details on the methodology used to estimate fugitive emission from natural gas distribution system are presented in Annex 3.

The original study method is a rigorous IPCC Tier 3 approach.

### 3.3.2.3. Uncertainties and Time-Series Consistency

#### Upstream Oil and Gas

The UOG fugitive emissions for 2000 are taken directly from CAPP's UOG study (CAPP 2005a). The emissions from 1990–1999 and from 2001–2009 have been estimated using the 2000 data, along with other factors discussed above. The uncertainty for the overall 2000 emissions is  $\pm 1.5\%$ . The uncertainties for the 2000 emissions for the oil and natural gas industries are listed in Table 3–7 and Table 3–8, respectively. The detailed uncertainties for each gas can be found in the UOG report (CAPP 2005a).

The uncertainties were determined using the Tier 1 uncertainty approach presented in the IPCC Good Practice Guidance (IPCC 2000). According to the IPCC (2000), there are three sources of uncertainties: definitions, natural variability of the process that produces the emissions, and the

4 Statistics Canada. 2010. Personal communication (email from Bristow C, Manufacturing and Energy Division, Statistics Canada to Smyth S, Pollutant Inventories and Reporting Division (PIRD), Environment Canada dated 7 Dec 2010).

Table 3–7 Uncertainty in Oil Production Industry Fugitive Emissions

GHG Source Category	Uncertainty (%)		
	Oil Exploration	Oil Production	Oil Transportation
Flaring	$\pm 4.2$	$\pm 2.3$	$\pm 24.0$
Fugitive	-8.9 to +8.3	$\pm 7.4$	-20.9 to +21.0
Venting	-38.4 to +30.4	-3.7 to +3.4	—
Total	-2.3 to +2.1	$\pm 3.1$	-16.7 to +16.8

Table 3–8 Uncertainty in Natural Gas Production Industry Fugitive Emissions

GHG Source Category	Uncertainty (%)
	Gas Production/Processing
Flaring	-2.6 to +2.2
Fugitive	-0.6 to +1.1
Other	±1.7
Venting	-4.0 to +3.5
Total	±0.7

Source: CAPP (2005b).

assessment of the process or quantity. Only the last two sources of uncertainty were considered in the analysis: it was assumed that the uncertainties from the definitions were negligible, as they were adequately controlled through QA/QC procedures. The uncertainty would be greater for those years that were estimated using the UOG model as compared with the uncertainty of the facility-based emission estimates for the 2000 data year.

## Downstream Oil and Gas

The emission data used in the inventory for fugitive emissions from refineries for 1990 and for 1994–2002 are taken directly from the CPPI (2004) study. There is greater uncertainty for the 1991–1993 and the 2003–2009 periods due to the available level of disaggregation of the activity data. Tier 1 and Tier 2 uncertainty analyses were performed, for comparison purposes, of the emission factors and activity data, for an overall CO<sub>2</sub> uncertainty in the 2002 data (CPPI 2004).

The results of these analyses are as follows: For the Tier 1 analysis, the overall uncertainty was ±8.3%. The Tier 2 analysis determined that the overall uncertainty was ±14%. The difference between the Tier 1 and Tier 2 uncertainties may be due to the high level of variability in some of the emission factors. The uncertainty results can be found in Table 3–9.

## Oil Sands / Bitumen

Only facility-level uncertainty estimates are currently available. Clearstone Engineering Ltd. conducted an IPCC Good Practice Guidance Tier 1 uncertainty assessment for each facility, and full details of the assessment can be found in the bitumen study (CAPP 2006) and the bitumen model (Environment Canada 2007). Development of an overall uncertainty range for this industry will be part of the uncertainty analysis improvement plan.

### 3.3.2.4. QA/QC and Verification

To ensure that the results were correct in the UOG study (CAPP 2005a), Clearstone Engineering Ltd. performed the following QA/QC procedures. First, all results were reviewed internally by senior personnel to ensure that there were no errors, omissions or double counting. The report was also reviewed by individual companies for comment. A second level of review was performed by the project steering committee and nominated experts. Furthermore, where possible, results were compared with previous baseline data and other corporate, industrial and national inventories. Any anomalies were verified through examination of activity levels, changes in regulations, and voluntary industry initiatives.

Tier 1 QC checks consistent with IPCC Good Practice Guidance (IPCC 2000) were performed on the CO<sub>2</sub> and CH<sub>4</sub>

Table 3–9 Uncertainty in Oil Refining Fugitive Emissions

	Uncertainty (%)			
	Overall	Excluding Refinery Fuel Gas	Excluding Flare Gas	Excluding Refinery Fuel and Flare Gas
Tier 1	± 8.3	± 4.3	± 8.3	± 8.3
Tier 2	± 14	± 5	± 14	± 14

estimates for the following key subcategories:

- Oil and Natural Gas Industries
- Oil and Natural Gas Venting and Flaring

No significant mathematical errors were found during the QC checks. The data, methodologies and changes related to the QC activities are documented and archived in both paper and electronic form.

### 3.3.2.5. Recalculations

*Petroleum Refining:* Fugitive emissions from refineries for the years 1991–1993, 1999 and 2003–2008 were recalculated using an updated refinery model, while 2008 also used updated activity data from Statistics Canada. These recalculations affected three CRF categories: 1.B.2.A.4 Refining/Storage, 1.B.2.c.i Venting – Oil, and 1.B.2.c.i Flaring – Oil. In previous years, fugitive emissions were calculated using total refinery energy use; however, a methodology employing total refinery production proved a better predictor of these emissions.

*Upstream Oil and Gas:* Updated activity data from various sources, including Statistics Canada and several provincial and industry reports, were utilized in the Upstream Oil and Gas (UOG) extrapolation model. Revisions to crude bitumen production volumes for 2007 and 2008 resulted in changes in emissions for CRF categories 1.B.2.a.iii Oil – Transport, and 1.B.2.c.i Venting – Oil. Changes in natural gas production volumes for 2007 and 2008 resulted in changes to emission estimates for CRF categories 1.B.2.b.ii Gas – Production/Processing, and 1.B.2.c.ii Venting – Natural Gas. Also, changes in the total number of wells capable of producing in Saskatchewan from 2001 to 2008 resulted in changes to emissions in CRF category 1.B.2.b.v Natural Gas – Other Leakage – at industrial plants and power stations. In addition, emissions from glycol dehydrators in Saskatchewan in the year 2000 in the heavy oil production industry were mistakenly omitted from previous submissions. Their inclusion resulted in a small increase in emissions in CRF category 1.B.2.c.i Venting – Oil. Various other changes to activity data resulted in changes to CRF category 1.B.2.c.i Flaring – Oil.

Finally, 1990 to 1999 formation CO<sub>2</sub> emissions were revised based on a new method that affected emissions in CRF category 1.B.2.c.ii Venting – Gas. In the 2009 submission, the method for estimating formation CO<sub>2</sub> emissions from 2001 onwards was changed from using natural gas production to using natural gas shrinkage as a proration factor. The

use of natural gas shrinkage, which contains the volume of formation CO<sub>2</sub> removed and vented, is a better measure of changes in formation CO<sub>2</sub> emissions than production. This same method was applied to 1990 to 1999 in order to provide a consistent method for the full time series. Note that 2000 data are taken directly from the UOG study (CAPP 2005a), and 1990–1999 and 2001–2009 data are estimated based on the 2000 data and various proration factors.

All of these updates were implemented in order to improve accuracy and ensure consistency with the most up-to-date data available.

*Natural Gas Transmission:* Revised natural gas transmission pipeline lengths from Statistics Canada<sup>5</sup> included a natural gas transmission pipeline in the Northwest Territories in 2007 and 2008. Further investigation revealed that the Ikhlil Pipeline had been serving Inuvik, Northwest Territories, with natural gas since 1999 (Quenneville 2009). The 2007 distance was assumed to be constant for all years back to 1999. This resulted in a slight change to the national estimates for Natural Gas Transmission emissions.

*Natural Gas Distribution:* Based on revised emissions data (GRI 2000) for the natural gas distribution industry, a new emission factor was developed. The Gas Research Institute reviewed and revised the Canadian Gas Association study (CGA 1997) with more accurate and better substantiated data for station vents (GRI 2000). This new emission factor resulted in a decrease in emissions of approximately 52% for the entire time series from 1990 to 2009. In addition, natural gas distribution pipeline lengths were revised for Nova Scotia from 2000 to 2008,<sup>6</sup> New Brunswick from 2000 to 2008,<sup>7</sup> Northwest Territories from 1999 to 2008 and British Columbia from 2000 to 2008, based on new data from Statistics Canada.<sup>8</sup> These changes in provincial activity data resulted in minor changes to the national estimates.

5 Statistics Canada. 2010. Personal communication (email from Bristow C, Manufacturing and Energy Division, Statistics Canada to Smyth S, Pollutant Inventories and Reporting Division (PIRD), Environment Canada dated 7 Dec 2010).

6 Heritage Gas. 2010. Personal communication (email from Bracken J, President, Heritage Gas to Smyth S, Pollutant Inventories and Reporting Division (PIRD), Environment Canada dated 7 Dec 2010).

7 Enbridge Gas New Brunswick. 2010. Personal communication (email from Nicholson L, Communications Coordinator, Enbridge Gas New Brunswick to Smyth S, Pollutant Inventories and Reporting Division (PIRD), Environment Canada dated 7 Dec 2010).

8 Statistics Canada. 2010. Personal communication (email from Bristow C, Manufacturing and Energy Division, Statistics Canada to Smyth S, Pollutant Inventories and Reporting Division (PIRD), Environment Canada dated 7 Dec 2010).

**Bitumen/Oil Sands:** Updated activity data from the ERCB (2010a) used in the bitumen model, along with corrections to transcription errors, resulted in changes in emissions for the years 2006 to 2008 in CRF categories 1.B.2.a.ii. Oil – Production, 1.B.2.c.i Venting – Oil, and 1.B.2.c.i Flaring – Oil. In addition, a small oil sands facility that was erroneously omitted in previous submissions was added in this submission, resulting in a slight increase in emissions in the aforementioned CRF categories. These updates were implemented in order to improve accuracy and ensure consistency with the most up-to-date data available.

### 3.3.2.6. Planned Improvements

**Upstream Oil and Gas:** A multi-year study is being commissioned to update the UOG study (CAPP 2005a) and is expected to be completed by the end of 2012. This study will update emissions based on the most recent data available (i.e. 2010 data year) as well as incorporate new and emerging oil and gas sources.

**Bitumen/Oil Sands:** In the long term, a comprehensive study to update the bitumen study (CAPP 2006) is planned with the goal of improving emission estimates from oil sands mining and extraction, in-situ production and upgrading in Canada. The new study will also develop a robust method for updating emission estimates in the rapidly expanding oil sands industry, as prioritized in recent expert review team (ERT) reviews.

**Natural Gas Transmission and Distribution:** Improvements to the natural gas transmission and distribution emissions model will be investigated with a focus on developing a method that will better reflect the improvements in efficiency made by the industry. The current models are not capable of capturing equipment changes or technology improvements, as emissions are estimated based on pipeline lengths and static emission factors.

## 3.4. Memo Items (CRF Category 1.C)

### 3.4.1. International Bunker Fuels (CRF Category 1.C.1)

According to the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997), emissions resulting from fuels sold for international marine and aviation transportation should not be included in national inventory totals, but should

be reported separately as emissions from international bunkers. Historically, in the Canadian inventory, any fuel reported by Statistics Canada as having been sold to foreign-registered marine or aviation carriers was excluded from national inventory emission totals. However, this assumption is no longer necessary due to the implementation of a true origin-destination aviation model.

For marine fuels, it has not been clear whether all of the fuel sold to foreign-registered carriers in Canada is used for international transport. More importantly, it has become apparent that not all of the fuels sold to domestically registered carriers are consumed within the country. The UNFCCC and the IPCC are currently developing clearer guidelines for bunkers, and modified statistical procedures may be required to track marine bunker fuels more accurately.

#### 3.4.1.1. Aviation (CRF Category 1.C.1.A)

Emissions (Table 3–10) have been calculated using the same methods listed in the Civil Aviation (Domestic Aviation) section (see Section 3.2.3.2). Fuel-use data are reported in the RESD (Statistics Canada #57-003) and are identified as being sold to foreign airlines; however, Statistics Canada is most confident in their published total fuel use value for aviation, with rapidly increasing uncertainty associated with further disaggregation into various categories such as foreign airlines. Therefore, the total fuel reported in the RESD is adhered to, with AGEM using flight-by-flight aircraft movements to determine whether or not a flight stage is domestic or international. This method greatly improves the allocation between domestic and international flights, as the highest resolution data available are being used to apply the mandated definitions.

#### 3.4.1.2. Marine (CRF Category 1.C.1.B)

Emissions (Table 3–11) have been calculated using the same methods listed in the Navigation (Domestic Marine) section (see Section 3.2.3.2). Fuel-use data are reported as foreign marine in the RESD (Statistics Canada #57-003).

### 3.4.2. CO<sub>2</sub> Emissions from Biomass

As per the UNFCCC reporting guidelines, CO<sub>2</sub> emissions from the combustion of biomass used to produce energy are not included in the Energy Sector totals but are re-

Table 3–10 GHG Emissions from Domestic and International Aviation

	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Aviation Bunkers (International)	6 000	8 900	9 300	10 000	9 600	9 900	9 700	9 500
Civil Aviation (Domestic)	7 200	7 500	7 700	7 700	7 800	7 800	7 800	7 200
<b>Total</b>	<b>13 200</b>	<b>16 400</b>	<b>17 000</b>	<b>17 700</b>	<b>17 500</b>	<b>17 800</b>	<b>17 400</b>	<b>16 600</b>

Note: Totals may not add up due to rounding.

Table 3–11 GHG Emissions from Domestic and International Navigation

	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Marine Bunkers (International)	3 100	3 600	2 000	2 000	1 800	2 100	1 800	1 600
Navigation (Domestic)	5 000	5 100	6 600	6 400	5 800	6 100	5 900	5 100
<b>Total</b>	<b>8 200</b>	<b>8 700</b>	<b>8 700</b>	<b>8 400</b>	<b>7 600</b>	<b>8 300</b>	<b>7 600</b>	<b>6 700</b>

Note: Totals may not add up due to rounding.

ported separately as memo items. They are accounted for in the Land Use, Land-Use Change and Forestry (LULUCF) Sector and are recorded as a loss of biomass (forest) stocks. CH<sub>4</sub> and N<sub>2</sub>O emissions from the combustion of biomass fuels for energy are reported in the fuel combustion section in the appropriate categories.

Biomass emissions have been grouped into three main sources: residential firewood, industrial wood wastes, and fuel ethanol/biodiesel used in transportation.

### 3.4.2.1. Residential Firewood

Firewood is used as a primary or supplementary heating source for many Canadian homes. Combustion of firewood results in CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O emissions which are considered technology-dependent. The main types of residential wood combustion devices considered are stoves, fireplaces, furnaces and other equipment (i.e. pellet stoves).

### 3.4.2.2. Industrial Wood Wastes

Biomass combusted in the industrial sector consists of industrial fuelwood and spent pulping liquor. This combustion of biomass is reported in the RESD and is attributed

solely to the pulp, paper and print industry as it is the primary consumer. Data are not available at a high enough level of aggregation to determine usage by other industries.

### 3.4.2.3. Fuel Ethanol

Amounts of fuel ethanol used in transportation (TFIS Inc. 2008) are presented in Table 3–12. Ethanol properties were developed according to chemistry and resulted in a higher heating value (HHV) (gross calorific value, or GCV) of 24.12 TJ/ML, 52.14% carbon content and 789.2 kg/m<sup>3</sup> density. The 2007 ethanol fuel consumption in transportation was kept constant for years 2008 and 2009 since no data were available during the development of this submission. An updated biofuel study should be completed in 2011.

Fuel ethanol was introduced and modelled as if it were mixed into the total gasoline for the region(s). Total fuel ethanol available per province was allocated to each mode (on-road, by vehicle technology classes, and off-road as a whole) as per the percentage of total gasoline calculated traditionally with MGEM. In lieu of reviewed emission factors for CH<sub>4</sub> and N<sub>2</sub>O for ethanol, the representative gaso-

Table 3–12 Ethanol Used for Transport in Canada

Year	1990	2000	2004	2005	2006	2007	2008	2009
Ethanol Consumed (ML)	7	355	477	462	482	1110	1110	1110



Table 3–13 Biodiesel Used for Transport in Canada

Year	1990	2000	2004	2005	2006	2007	2008	2009
Biodiesel Consumed (ML)	0	0	2	2	29	43	43	43

line emission factor was applied as per mode and technology class. CO<sub>2</sub> emission factors used are those based upon true chemical characteristics mentioned previously and a 99% oxidation rate.

In previous inventory years, it was assumed that the RESD gasoline fuel use totals did not include ethanol. However, based on recent feedback from Statistics Canada, it was confirmed that ethanol is included in RESD gasoline fuel consumption data. MGEM now allocates ethanol based on this information, such that the total sum of gasoline and ethanol for each mode is equivalent to the total amount of gasoline reported in the RESD.

#### 3.4.2.4. Fuel Biodiesel

The amounts of biodiesel fuel used in transportation (TFIS Inc. 2008) are presented in Table 3–13. The properties used for biodiesel were extracted from a biodiesel study conducted between 2004 and 2005 (BioMer 2005). The higher heating value (HHV) (gross calorific value, or GCV) used is 35.18 TJ/ML, with a 76.5% carbon content and 882 kg/m<sup>3</sup> density. The 2007 biodiesel fuel consumption in transportation was kept constant for years 2008 and 2009 since no data were available during the development of this submission. An updated biofuel study should be completed in 2011.

Biodiesel was introduced and modelled as if it were mixed into the total petrodiesel for the region(s). Total fuel available per province was allocated to each mode (on-road, by vehicle technology classes, and off-road, railways and domestic marine as a whole) as per the percentage of total petrodiesel calculated traditionally with MGEM. In lieu of reviewed emission factors for CH<sub>4</sub> and N<sub>2</sub>O for biodiesel, the representative petrodiesel emission factor was applied as per mode and technology class. CO<sub>2</sub> emission factors used are those based upon true chemical characteristics mentioned previously and a 99% oxidation rate.

## 3.5. Other Issues

### 3.5.1. Comparison of Sectoral and Reference Approaches

Refer to Annex 4: Comparison of Sectoral and Reference Approaches, which presents a full discussion of this topic.

### 3.5.2. Feedstocks and Non-energy Use of Fuels

Emissions from fuel use in the Energy Sector are those related to the combustion of fuels for generating heat or work. In addition to being combusted for energy production, fossil fuels are also consumed for non-energy purposes. Non-energy uses of fossil fuels include application as waxes, solvents, lubricants and feedstocks (including the manufacturing of fertilizers, rubber, plastics and synthetic fibres). Emissions from the non-energy use of fossil fuels have been included in the Industrial Processes Sector, whereas emissions from the use of fossil fuels associated with flaring activities by the oil and gas industry are included in the Fugitive subsector.

Refer to the Industrial Processes chapter (Chapter 4) for a discussion of the use of feedstocks and the non-energy use of fossil fuels and the methodological issues associated with calculating emissions from this source.

### 3.5.3. Carbon Capture and Storage – Enhanced Oil Recovery

In Canada, anthropogenic CO<sub>2</sub> is used as a flooding agent in enhanced oil recovery (EOR) operations to increase crude oil production volume at two depleting oil reservoirs. CO<sub>2</sub> flooding started in 2000 at the Weyburn site and in 2005 at the Apache Midale site in order to extend the life of these mature reservoirs by another 30 years. CO<sub>2</sub> is purchased from the Dakota Gasification Company located in North Dakota (US), transported via pipeline to the field



and this fresh supply is then combined with recycled CO<sub>2</sub> for re-injection back into the reservoir. Currently about 2.8 Mt per year of CO<sub>2</sub> is injected at the Weyburn-Midale operations (Mourits 2010). The Weyburn site has, from 2000 to 2009, injected about 15 Mt of fresh CO<sub>2</sub> as purchased from the Dakota gasification plant with a current injection rate of 6500 t of CO<sub>2</sub> per day. From 2005 to 2009, the Midale site has injected about 1.91 Mt of fresh CO<sub>2</sub> with an injection rate of 1165 t of CO<sub>2</sub> per day.

CO<sub>2</sub> is used as flooding agent in EOR since it acts as a solvent and it helps increase the reservoir pressure, resulting in the release of trapped hydrocarbons to production wells. The flooding process of CO<sub>2</sub> at high pressure also results in CO<sub>2</sub> being trapped in the voids which were previously occupied by hydrocarbon molecules and this process is commonly known as geological storage of CO<sub>2</sub>.

In addition to being a CO<sub>2</sub> EOR operation, Weyburn is also the site of a full-scale geological CO<sub>2</sub> storage research program led by the International Energy Agency's Greenhouse Gas (IEA GHG) Research and Development Programme with the support of various industries, research organizations and governments. Modelling and simulation results from the first phase (from 2000 to 2004) of the IEA GHG's CO<sub>2</sub> monitoring and storage project, as managed by the Petroleum Technology Research Centre (PTRC), indicates that over 98% of CO<sub>2</sub> will remain trapped in the Weyburn reservoir after 5000 years and only 0.14% will be released to the atmosphere (Mourits 2008). Additional details on the findings of the first phase of the research project are available on the PTRC website ([www.ptrc.ca](http://www.ptrc.ca)—see PTRC 2004).

The final phase (from 2005 to 2011) of the IEA Weyburn-Midale research project as outlined on the PTRC website will focus on technical and non-technical components such as site characterization, selection, well bore integrity, monitoring and verification, risk assessment, regulatory issues, public communication and outreach, and business environment policy in order to develop a best practice manual for future projects on the geological storage of CO<sub>2</sub>.

### 3.5.4. Country-specific Issues: Emissions Associated with the Export of Fossil Fuels

Canada exports a great deal of its produced fossil fuel resources, mostly to the United States. In 2009, Canada exported approximately 63% (energy equivalent) of its gross natural gas and crude oil production. These exports have shown a large growth trend since 1990 and are responsible for considerable quantities of greenhouse gas emissions. The emissions associated with the import/export of crude oil and natural gas are estimated using the updated “Fossil Fuel Import and Export” models (Smyth 2010; Environment Canada 2008). The methodology used in the updated models is based on McCann (1997) which, prior to the 2009 submission, had been used to estimate these emissions. The new methodology employs pre-existing models currently used in assembling the national emission estimates, as well as annually updated activity data from a variety of sources. The emissions/sectors included within the two main fuel stream estimates are as follows:

- *Natural Gas:* This category accounts for GHG emissions specific to the production, gathering, processing and transmission of natural gas. It includes emissions from gas conservation systems at oil batteries (i.e. dehydrators, compressors and related piping) and excludes emissions that may be attributed to the handling, processing (e.g. stabilization, treating and/or fractionation) or storage of NGLs at gas facilities. Basically, only those sources that exist for the primary purpose of producing natural gas for sale are considered. Gas distribution systems and end-use emissions are specifically excluded, since they pertain to domestic gas consumption rather than gas imports and exports.
- *Crude Oil:* Similarly, this category considers emissions related to the production, treatment, storage and transport of crude oils. Emissions from venting and flaring of associated or solution gas at these facilities are allocated to this category. Any gas equipment that is dedicated to servicing on-site fuel needs is part of the oil system. Gas conservation systems that collect emissions in a gas-gathering system are allocated to the natural gas system.

It must be noted that the absolute emission estimates provided here have a high level of uncertainty—as great as 40% or more. On the other hand, the trend estimates are more accurate and can be considered to be representative.

The results demonstrate that between 1990 and 2009, oil exports grew by 180% to 4306 petajoules (PJ)<sup>9</sup> (almost three times the rate of growth of oil production) (Table 3–14), while exports of natural gas increased 137% to 3639 PJ (approximately two and a half times the rate of growth of natural gas production) (Table 3–15).<sup>10</sup> Furthermore, the sum total of oil and gas energy exports increased by 158% over the same period (Table 3–16). It is important to note that natural gas exports had not experienced substantial changes from 2000 to 2008, but from 2008 to 2009, exports dropped by 8%. This is due to declining U.S. demand along with decreasing production and diminishing reserves in Canada's largest natural gas reservoir (the Western Sedimentary Basin) (Nyboer and Tu 2006).

Whereas natural gas production is declining in Canada, production and exports from Canada's oil sands have been increasing. From 1990 to 2002, exports of conventional crude oil in energy terms increased by 111%, but from

2002 to 2009 exports decreased by 12% (Table 3–17). However, exports of unconventional crude oil<sup>11</sup> from Canada's oil sands have been consistently increasing (Table 3–18). From 1990 to 2009, exports of unconventional crude oil have increased by 430%.

Increased Canadian fossil fuel exports have been offset, in part, by increased fossil fuel imports. Indeed, 51% more energy from crude oil was imported in 2009 than in 1990. The balance between changes in exports, imports and production reflects an increase of 17% in domestic consumption of crude oil and natural gas between 1990 and 2009 (Table 3–16).

Activities associated with the oil and gas industry result in considerable GHG emissions. Between 1990 and 2009, increases in oil and gas production for export (principally to the United States) contributed substantially to emissions growth. Total emissions associated with the production, processing and transmission of all oil and gas destined for export were about 92 Mt in 2009, up 178% from 1990.

9 A petajoule (PJ) is a measure of the energy content of fuels.

10 The source for all export and energy production data is Statistics Canada's Report on Energy Supply–Demand in Canada (RES-D, #57-003). The 1990–2009 GHG emissions associated with net exports are from Environment Canada (2008).

11 Unconventional crude oil includes crude bitumen from mining and *in-situ* sources as well as synthetic crude oil.

Table 3–14 Crude Oil: Production, Export and GHG Emission Trends, Select Years

Crude Oil Trends	1990	2000	2004	2005	2006	2007	2008	2009
Domestic Production (PJ)	3 785	5 410	5 938	5 834	6 147	6 394	6 339	6 318
<i>Growth Since 1990</i>	NA	43%	57%	54%	62%	69%	67%	67%
Energy Imported (PJ)	1 198	1 985	2 092	2 071	1 898	1 919	1 896	1 804
<i>Growth Since 1990</i>	NA	66%	75%	73%	58%	60%	58%	51%
Energy Exported (PJ)	1 538	3 423	3 832	3 830	4 212	4 203	4 238	4 306
<i>Growth Since 1990</i>	NA	123%	149%	149%	174%	173%	176%	180%
Apparent Domestic Consumption (PJ)	3 445	3 971	4 197	4 075	3 833	4 110	3 997	3 816
<i>Growth Since 1990</i>	NA	15%	22%	18%	11%	19%	16%	11%
Emissions Associated with Gross Exports (Mt CO <sub>2</sub> eq.)	21.0	49.5	53.1	53.1	56.6	57.9	60.6	62.2
<i>Growth Since 1990</i>	NA	136%	153%	153%	170%	176%	189%	196%
Emissions Associated with Gross Imports (Mt CO <sub>2</sub> eq.)	8.6	14.0	15.4	16.5	16.3	16.6	18.2	16.0
<i>Growth Since 1990</i>	NA	63%	80%	92%	90%	93%	112%	87%
Emissions Associated with Net Exports <sup>1</sup> (Mt CO <sub>2</sub> eq.)	12.4	35.5	37.6	36.6	40.3	41.3	42.4	46.1
<i>Growth Since 1990</i>	NA	187%	204%	196%	225%	233%	243%	273%

Notes:

1. See box on Net Export Emissions at end of Chapter 3

\* NA = Not applicable.

Table 3–15 Natural Gas: Production, Import, Export and GHG Emission Trends, Select Years

<b>Natural Gas Trends</b>	<b>1990</b>	<b>2000</b>	<b>2004</b>	<b>2005</b>	<b>2006</b>	<b>2007</b>	<b>2008</b>	<b>2009</b>
Domestic Production (PJ)	4 184	7 250	7 118	7 183	7 196	6 994	6 652	6 200
<i>Growth Since 1990</i>	NA	73%	70%	72%	72%	67%	59%	48%
Energy Imported (PJ)	24.2	252	416	364	369	504	596	789
<i>Growth Since 1990</i>	NA	939%	1619%	1403%	1423%	1980%	2361%	3159%
Energy Exported (PJ)	1 537	4 103	4 035	4 061	3 901	4 117	3 936	3 639
<i>Growth Since 1990</i>	NA	167%	162%	164%	154%	168%	156%	137%
Apparent Domestic Consumption (PJ)	2 671	3 398	3 500	3 486	3 663	3 380	3 313	3 351
<i>Growth Since 1990</i>	NA	27%	31%	31%	37%	27%	24%	25%
Emissions Associated with Gross Exports (Mt CO <sub>2</sub> eq.)	12.3	35.7	34.8	35.6	34.0	36.2	34.8	30.2
<i>Growth Since 1990</i>	NA	190%	183%	190%	177%	195%	183%	146%
Emissions Associated with Gross Imports (Mt CO <sub>2</sub> eq.)	0.3	3.4	5.1	4.5	3.9	5.1	5.7	6.8
<i>Growth Since 1990</i>	NA	1019%	1554%	1366%	1178%	1554%	1766%	2120%
Emissions Associated with Net Exports <sup>1</sup> (Mt CO <sub>2</sub> eq.)	12.0	32.2	29.7	31.1	30.1	31.1	29.0	23.4
<i>Growth Since 1990</i>	NA	169%	148%	160%	151%	160%	143%	95%

Notes:

1. See box on Net Export Emissions at end of Chapter 3

\* NA = Not applicable.

Table 3–16 Combined Crude Oil and Natural Gas: Production, Export, and GHG Emission Trends, Select Years

<b>Crude Oil &amp; Natural Gas Trends</b>	<b>1990</b>	<b>2000</b>	<b>2004</b>	<b>2005</b>	<b>2006</b>	<b>2007</b>	<b>2008</b>	<b>2009</b>
Domestic Production (PJ)	7 969	12 660	13 056	13 017	13 343	13 388	12 991	12 518
<i>Growth Since 1990</i>	NA	59%	64%	63%	67%	68%	63%	57%
Energy Imported (PJ)	1 222	2 236	2 508	2 435	2 267	2 423	2 492	2 593
<i>Growth Since 1990</i>	NA	83%	105%	99%	86%	98%	104%	112%
Energy Exported (PJ)	3 075	7 527	7 867	7 891	8 113	8 321	8 174	7 945
<i>Growth Since 1990</i>	NA	145%	156%	157%	164%	171%	166%	158%
Apparent Domestic Consumption (PJ)	6 116	7 369	7 697	7 562	7 496	7 490	7 309	7 166
<i>Growth Since 1990</i>	NA	20%	26%	24%	23%	22%	20%	17%
Emissions Associated with Gross Exports (Mt CO <sub>2</sub> eq.)	33.3	85.2	87.8	88.8	90.6	94.1	95.4	92.4
<i>Growth Since 1990</i>	NA	156%	164%	167%	172%	183%	187%	178%
Emissions Associated with Gross Imports (Mt CO <sub>2</sub> eq.)	8.9	17.5	20.5	21.0	20.2	21.7	23.9	22.9
<i>Growth Since 1990</i>	NA	96%	131%	136%	127%	144%	169%	157%
Emissions Associated with Net Exports <sup>1</sup> (Mt CO <sub>2</sub> eq.)	24.4	67.7	67.3	67.8	70.4	72.4	71.5	69.5
<i>Growth Since 1990</i>	NA	178%	176%	178%	189%	197%	193%	185%

Notes:

1. See box on Net Export Emissions at end of Chapter 3

\* NA = Not applicable.

Table 3–17 Conventional Crude Oil: Production, Export, and GHG Emission Trends, Select Years

<b>Conventional Crude Oil Trends</b>	<b>1990</b>	<b>2000</b>	<b>2004</b>	<b>2005</b>	<b>2006</b>	<b>2007</b>	<b>2008</b>	<b>2009</b>
Domestic Production (PJ)	2 973	3 657	3 567	3 459	3 434	3 542	3 447	3 115
<i>Growth Since 1990</i>	NA	23%	20%	16%	16%	19%	16%	5%
Energy Imported (PJ)	1 198	1 985	2 092	2 071	1 898	1 919	1 896	1 804
<i>Growth Since 1990</i>	NA	66%	75%	73%	58%	60%	58%	51%
Energy Exported (PJ)	1 117	2 353	2 378	2 301	2 399	2 245	2 180	2 073
<i>Growth Since 1990</i>	NA	111%	113%	106%	115%	101%	95%	86%
Apparent Domestic Consumption (PJ)	3 054	3 289	3 281	3 229	2 933	3 217	3 163	2 846
<i>Growth Since 1990</i>	NA	8%	7%	6%	-4%	5%	4%	-7%
Emissions Associated with Gross Exports (Mt CO <sub>2</sub> eq.)	11.4	29.2	29.6	29.8	29.8	26.3	27.7	25.3
<i>Growth Since 1990</i>	NA	155%	159%	160%	160%	129%	142%	121%
Emissions Associated with Net Exports <sup>1</sup> (Mt CO <sub>2</sub> eq.)	2.9	15.2	14.2	13.3	13.5	9.6	9.5	9.3
<i>Growth Since 1990</i>	NA	432%	396%	365%	371%	237%	232%	225%

Notes:

1. See box on Net Export Emissions at end of Chapter 3

\* NA = Not applicable.

Table 3-18 Unconventional Crude Oil: Production, Export, and GHG Emission Trends, Select Years

Unconventional Crude Oil Trends	1990	2000	2004	2005	2006	2007	2008	2009
Domestic Production (PJ)	813	1 753	2 371	2 375	2 712	2 852	2 892	3 203
<i>Growth Since 1990</i>	N/A	116%	192%	192%	234%	251%	256%	294%
Energy Exported (PJ)	421	1 071	1 454	1 529	1 813	1 959	2 058	2 233
<i>Growth Since 1990</i>	N/A	154%	245%	263%	330%	365%	388%	430%
Apparent Domestic Consumption (PJ)	391	682	916	847	900	893	834	970
<i>Growth Since 1990</i>	N/A	74%	134%	116%	130%	128%	113%	148%
Emissions Associated with Gross Exports (Mt CO <sub>2</sub> eq.)	9.5	20.3	23.5	23.3	26.8	31.6	32.9	36.9
<i>Growth Since 1990</i>	N/A	113%	146%	145%	181%	232%	246%	287%

Notes:

\* NA = Not applicable.

### Net Export Emissions

Since Canada both exports and imports significant quantities of fuel, determination of emissions associated with *net* (rather than gross) exports provides a clearer picture of emissions arising from domestic energy demand. Net export emissions are the Canadian emissions associated with extracting, processing and transporting exported fuels minus the foreign emissions associated with the same activities for imported fuels. The emissions associated with net exports approximate the quantity of GHGs that would be ascribed to Canada if it was responsible only for those above and beyond its own demand. Net exports rose from about 24 Mt in 1990 to 70 Mt in 2009 (a 185% increase; Table 3-16).\*

\*Note that the long-term trends for net export emissions are more accurate than net export emissions calculated for any given year.

# Chapter 4

## Industrial Processes (CRF Sector 2)

### 4.1. Overview

Greenhouse gas (GHG) emissions are produced from a variety of activities that are not related to energy. The main emission sources are industrial processes that chemically or physically transform materials. During these processes, many different greenhouse gases, including CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, and PFCs, can be released (IPCC/OECD/IEA 1997). Certain halocarbons (HFCs and PFCs) and SF<sub>6</sub> are also consumed in industrial processes or used as alternatives to ozone-depleting substances (ODS) in various applications; these emissions are also included in the Industrial Processes Sector.

GHG emissions from fuel combustion supplying energy to industrial activities are generally assigned to the Energy Sector. In some cases, such as when natural gas is used to produce ammonia, it is difficult to differentiate between emissions associated with energy and those produced by industrial process use of fuel. In such cases, and where predominance is with the industrial process use of fuel, the emissions are allocated to the Industrial Processes Sector. Emissions associated with the use of natural gas as feedstock in the upstream and downstream oil industries, to produce hydrogen, are assigned to the Energy Sector.

The processes addressed in the Industrial Processes Sector include production and use of mineral products, metal production, chemical (including petrochemicals) production, consumption of SF<sub>6</sub>, halocarbon production and use as alternatives to ozone-depleting substances, and other and undifferentiated production using fossil fuels for non-energy purposes.

CO<sub>2</sub> emissions resulting from the use of fossil fuels as feedstock for the production of chemical products other than ammonia, nitric acid, adipic acid and fuels used as reduc-

tants in the mining and processing of non-ferrous metals are reported in the Other and Undifferentiated Production subsector (Section 4.16).

Indirect GHG (such as CO, non-methane volatile organic compounds [NMVOC] and SO<sub>2</sub>) emissions from industrial process activities including asphalt roofing, road paving with asphalt, pulp and paper production, and production of food and drink have not been estimated. These emissions, along with the indirect GHG emissions associated with energy activities, are reported under Annex 10 of this National Inventory Report.

As shown in Table 4-1, GHG emissions from the Industrial Processes Sector contributed 46.3 Mt to the 2009 national GHG inventory, compared with 56.8 Mt in 1990. These emissions represented 6.7% of the total Canadian GHG emissions in 2009. The contributing factors of the long-term and short-term trends in this sector are discussed in detail in Chapter 2, and highlights of these are provided below.

Between 1990 and 2009, N<sub>2</sub>O emissions from the adipic acid industry had been considerably reduced (-94% or -10 Mt CO<sub>2</sub> eq) due to the idling of operation of Canada's only plant (in Maitland, Ontario) as of spring 2009. For the years post-1997 for which there was production, the use of an emission abatement system since 1997 also helped decrease emissions. Aluminium producers have also decreased their perfluorocarbon (PFC) emissions (-67% or -4.4 Mt CO<sub>2</sub> eq) by means of emission control technologies, while almost doubling their production volume. In addition, the industry has tried to gradually reduce its use of the old Soderberg production technology. For the category of Magnesium Production, the replacement of SF<sub>6</sub> with alternatives could explain the drop in emissions for one of the major plants between 1990 and 2005. For more recent years (i.e. 2006 onwards), the emission decrease was mainly due to a decrease in production and, hence, a decrease in SF<sub>6</sub> use, in anticipation of plant closures (one at the beginning of 2007 and another in August 2008). Also worth mentioning is that the 1990 estimate for the Iron and Steel industry was lower than it would normally be due to a strike (hence distorting to some extent the emission reductions achieved by the industry at regular production levels in the years after). The global economic downturn in 2009 has had a considerable impact on demand for steel products, such as in car manufacturing, construction and exports (in particular), which in turn affected the production in the Iron and Steel industry. The

Table 4–1 GHG Emissions from the Industrial Processes Sector, Selected Years

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Industrial Processes TOTAL	56 800	53 500	58 100	57 200	56 500	55 500	54 500	46 300
a. Mineral Products	8 300	9 600	9 500	9 500	9 600	9 300	8 600	6 800
<i>Cement Production</i>	5 400	6 700	7 100	7 200	7 300	7 300	6 600	5 100
<i>Lime Production</i>	1 800	1 900	1 800	1 700	1 600	1 600	1 500	1 200
<i>Limestone and Dolomite Use</i>	730	620	240	250	310	160	290	280
<i>Soda Ash Use</i>	210	220	150	160	170	170	150	100
<i>Magnesite Use</i>	147	181	184	175	149	68	57	69
b. Chemical Industry	17 000	9 000	11 000	10 000	9 100	8 900	10 000	8 100
<i>Ammonia Production</i>	5 000	6 800	6 800	6 300	6 600	6 200	6 700	6 200
<i>Nitric Acid Production</i>	1 010	1 230	1 230	1 250	1 230	1 130	1 230	1 150
<i>Adipic Acid Production</i>	11 000	900	3 100	2 600	1 200	1 500	2 400	660
<i>Petrochemical</i>	110	100	100	80	80	80	70	60
c. Metal Production	22 600	22 500	19 900	19 600	20 100	18 800	18 500	15 000
<i>Iron and Steel Production</i>	10 200	11 500	10 400	10 100	11 000	11 000	10 600	7 650
<i>Aluminium Production</i>	9 300	8 200	7 300	8 200	7 700	7 300	7 400	7 200
<i>Magnesium Production</i>	2 870	2 310	2 000	1 090	1 200	325	181	0
<i>Magnesium Casting</i>	236	471	191	201	190	198	280	193
d. Production and Consumption of Halocarbons (HFCs & PFCs)	770	3 000	4 700	5 200	5 000	5 400	5 500	6 800
e. SF <sub>6</sub> Use in Electric Utilities and Semiconductors	220	220	230	180	190	240	210	180
f. Other and Undifferentiated Production	8 000	9 200	13 000	12 000	13 000	13 000	11 000	9 400

Note: Totals may not add up due to rounding. Also, because of number rounding, some slight emission decreases or increases discussed in the paragraphs above may not be reflected in this table.

production of pig iron decreased by 39% and the production of steel decreased by 38% between 2008 and 2009. The Canadian chemical manufacturing sector also saw declines in its demand (albeit not as severe as for other manufacturing sectors). This is reflected in lower emission values for Ammonia Production, Adipic Acid (a closing plant), and the petrochemicals industry (included in the Other and Undifferentiated Production category).

The emission decreases mentioned above were partly offset by significant increases in emissions from other sources between 1990 and 2009. For instance, emissions from consumption of halocarbons have grown by ten times, or 5.0 Mt CO<sub>2</sub> eq, (since 1995) because of the progressive replacement of ODS, such as chlorofluorocarbons (CFCs) and hydrochlorofluorocarbons (HCFCs), by hydrofluorocarbons (HFCs).

A Tier 1 uncertainty assessment has been done for each of the categories under the Industrial Processes Sector. Results of the assessment are provided in the uncertainty section of each category.

To ensure that the inventory was correctly prepared, the key, new, and updated categories of this sector have all undergone Tier 1 level quality control checks.

To keep up with the principle of continuous improvement and to address a comment made by the Inventory Expert Review Team, a study covering emission categories (e.g. carbide production) previously reported as “NE” (not estimated) in the CRF reporter was carried out. Results of this study are used in this submission and lead to the addition of the category of Petrochemical Production under the Chemical Industry subsector. Other improvements for existing categories are also underway.

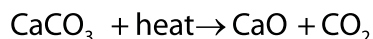
## 4.2. Cement Production (CRF Category 2.A.1)

### 4.2.1. Source Category Description

CO<sub>2</sub> is generated during the production of clinker, an intermediate product from which cement is made. Calcium



carbonate ( $\text{CaCO}_3$ ) from limestone, chalk, or other calcium-rich materials and other raw ingredients, such as silicates, are heated in a high-temperature kiln, forming lime ( $\text{CaO}$ ) and  $\text{CO}_2$ . This process is called calcination or calcining. It occurs in the lower-temperature section of the kiln (800–900°C) and can be represented as follows:



The lime is then combined with silica-containing materials in the higher-temperature section of the kiln (1350–1450°C) to produce clinker (greyish-black pellets about the size of 12 mm diameter marbles). The clinker is removed from the kiln, cooled, and pulverized, and gypsum is added to produce Portland cement. According to Statistics Canada's publications (catalogue #44-001 and CANSIM table 303-0060), more than 90% of the cement produced in Canada is of the Portland cement type. Portland cement contains 95–97% clinker by weight. The lime content of clinker ranges between 60% and 67% (IPCC 2006). Other specialty cements are lower in lime, but are typically used in small quantities.

$\text{CO}_2$  emissions from cement production are essentially directly proportional to lime content. The emissions resulting from the combustion of fossil fuels to generate the heat to drive the reaction in the kiln fall under the Energy Sector and are not considered here.

#### 4.2.2. Methodological Issues

To estimate  $\text{CO}_2$  emissions from cement production at the national level, the equation recommended in the IPCC Good Practice Guidance (IPCC 2000), as shown below, was used:

Equation 4–1:

$$\text{CO}_2 \text{ emissions} = \text{EF}_{\text{clinker}} \times \text{Clinker Production} \times \text{CKD Correction Factor}$$

where:

$\text{EF}_{\text{clinker}}$	=	emission factor based on clinker production, $\text{kt CO}_2/\text{kt clinker}$
Clinker Production	=	clinker production data, $\text{kt}$
CKD Correction Factor	=	factor that corrects for the loss of cement kiln dust (CKD), fraction

The IPCC default  $\text{EF}_{\text{clinker}}$  of 0.5071  $\text{kt CO}_2/\text{kt clinker}$  produced was applied. This factor was developed based on an average  $\text{CaO}$  content of 64.6% and the molecular weight ratio of  $\text{CO}_2$  to  $\text{CaO}$  in the raw material, which is 0.785 (IPCC/OECD/IEA 1997). The IPCC Good Practice Guidance (IPCC 2000) suggests 1.02 (i.e. adding 2% to the  $\text{CO}_2$  calculated for clinker) as the default CKD correction factor.

Clinker production data for 1990–1996 were obtained from *A Review of Energy Consumption and Related Data: Canadian Cement Manufacturing Industry, 1990 to 2008* (CIEEDAC 2010). Clinker production data for 1997–2004 were obtained from Statistics Canada (#44-001) and for 2005–2009 from CANSIM table 303-0060 (Statistics Canada 2005-2009). Applying Equation 4–1 above to the clinker production data is considered a Tier 2 type approach.

To estimate  $\text{CO}_2$  provincial/territorial emissions, data on clinker capacity of cement plants across Canada were used. The source of 1990–2006 data was the *Canadian Minerals Yearbook* (NRCan 1990–2006, and the author of the cement section of the *Canadian Minerals Yearbook* provided the 2007–2009 data<sup>1</sup>). These data were used to derive the percentage of total national clinker capacity attributed to each province/territory.  $\text{CO}_2$  emissions on a provincial/territorial level were estimated by multiplying the percentage attributed to each province/territory by the national emission estimate.

#### 4.2.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty has been developed based on the IPCC (2006) default uncertainty values provided for various parameters in the equation for  $\text{CO}_2$  emissions. Also considered was the error associated with the non-response rate of the Statistics Canada survey for clinker production data. The Tier 1 uncertainty associated with the  $\text{CO}_2$  estimate for cement production was  $\pm 28\%$ . The main contributor to the uncertainty was the use of the IPCC default correction factor related to cement kiln dust. The uncertainty value is applicable to all years of the time series. Equation 3.1 of the IPCC Good Practice Guidance (IPCC 2000) has been consistently applied over the time series. The activity data sources are described in Section 4.2.2.

1 Panagapko D. 2008–2010. Personal communications (emails from Panagapko D. to A. Shen, Greenhouse Gas Division, on December 12, 2008, to A. Au, Greenhouse Gas Division, on November 13, 2009, and to S. Chakrovorty, Greenhouse Gas Division, on June 22, 2010). Natural Resources Canada.

### 4.2.4. Category-Specific QA/QC and Verification

This key category in the Industrial Processes Sector has undergone Tier 1 quality control (QC) checks as elaborated in the quality assurance / quality control (QA/QC) plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

To improve the quality of this category with country-specific values of CaO content in clinker and CKD correction factor, a quality assurance (QA) package was sent to the Cement Association of Canada (CAC) in mid-2009. However, at the request of CAC, the project was deferred to a later date. Results of this QA project will be applied in the inventory as they become available.

### 4.2.5. Category-Specific Recalculations

Although estimates of emissions by province were updated, there have been no recalculations at the national level for this category because the updates done were only reallocations of emissions by province for the years 2006 to 2008.

### 4.2.6. Category-Specific Planned Improvements

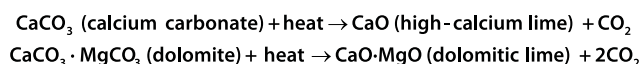
As mentioned above, to improve on methodology, emission factors and uncertainty of this category, a QA package was sent to the CAC. Results of this project will be applied in the inventory as they become available.

## 4.3. Lime Production (CRF Category 2.A.2)

### 4.3.1. Source Category Description

The production of lime involves a series of steps, including quarrying, crushing and sizing, and calcining (heat processing) of the raw materials, followed by transfer, storage and handling of the products (IPCC/OECD/IEA 1997). As mentioned in the source category description of cement production (Section 4.2.1), emissions of CO<sub>2</sub> occur at the

calcination stage, in which lime is formed via the thermal decomposition of carbonates at high temperatures in a rotary kiln. There are three main types of lime: high-calcium lime, dolomitic lime and hydraulic lime. It is important to distinguish between these in the emission estimation because the first two types have different stoichiometric ratios and the third has a substantially lower CaO content (IPCC 2000). High-calcium quicklime (CaO) and dolomitic quicklime (CaO·MgO) are obtained by calcining quarried calcium carbonate (CaCO<sub>3</sub>) and dolomite (CaCO<sub>3</sub>·MgCO<sub>3</sub>), respectively, as shown in the reactions below:



Both high-calcium and dolomitic limes can be slaked (i.e. treated with water under controlled conditions) and converted to hydrated limes in the form of Ca(OH)<sub>2</sub> and Ca(OH)<sub>2</sub>·Mg(OH)<sub>2</sub>, respectively. Also, according to the IPCC Good Practice Guidance (IPCC 2000), when there is no information on hydraulic lime, as in the case of Canada, the proportion of hydraulic lime should be assumed to be zero.

Emissions from the regeneration of lime from spent pulping liquors at pulp mills are not accounted for in the Industrial Processes Sector. Since this CO<sub>2</sub> is biogenic in origin, it is recorded as a change in forest stock in the Land Use, Land-Use Change and Forestry (LULUCF) Sector. The CO<sub>2</sub> associated with the use of natural limestone for producing lime in the pulp and paper industry is accountable and is included in the category Limestone and Dolomite Use (Section 4.4).

### 4.3.2. Methodological Issues

The methodology used to estimate the CO<sub>2</sub> emissions from lime production is of the Tier 2 type, as country-specific emission factors were applied to national activity data. The country-specific emission factors for high-calcium lime and dolomitic lime were developed based on the information on Canadian lime compositions collected from the Canadian Lime Institute.<sup>2</sup>

Data on total national lime production, hydrated lime production and lime plant calcining capacities were obtained from the *Canadian Minerals Yearbook* (NRCan

<sup>2</sup> Kenefick W. Personal communication (email from Wayne Kenefick to Amy Shen, Greenhouse Gas Division, dated October 7, 2008). Canadian Lime Institute.

1990–2006) or from the author of the lime section of the *Canadian Minerals Yearbook* (for 2007–2009). For any given year, the most recent lime production numbers provided are preliminary and are subject to revision in subsequent publications. As per the IPCC Good Practice Guidance (IPCC 2000), the total national lime production data were corrected by multiplying by a factor of  $1 - (x * y)$ , where  $x$  is the proportion of hydrated lime production to total lime production and  $y$  is the water content in the hydrated lime. Canadian hydrated lime has a  $y$  value (i.e. water content) of 28.25%<sup>3</sup>. Furthermore, the corrected lime production data were divided into high-calcium lime and dolomitic lime production based on the data on calcining capacities of lime production facilities across Canada. National CO<sub>2</sub> emissions were then calculated by applying the Canadian emission factors (provided in Annex 8) to the estimated yearly national lime production data, by lime type.

To estimate CO<sub>2</sub> emissions at the provincial level, the national emissions were allocated by province, according to the calcining capacity of each province.

#### 4.3.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the Lime Production category. It took into account the uncertainties associated with the production data, emission factors, correction factor for hydrated lime and the percentage split between the two types of lime. The uncertainty associated with the category as a whole was evaluated at  $\pm 8.0\%$ , with lime production data and the percentage split being the largest contributors. The uncertainty value is applicable to all years of the time series.

The same emission factors were consistently applied over the time series. The activity data source is provided in Section 4.3.2.

The data source and estimation technique used are consistent over the time series.

#### 4.3.4. Category-Specific QA/QC and Verification

This category in the Industrial Processes Sector has undergone Tier 1 QC checks as elaborated in the QA/QC plan

(see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No anomalies were observed.

#### 4.3.5. Category-Specific Recalculations

The whole time series (1990–2008) was revised due to corrections made to the emission factors that were applied for high-calcium lime and dolomitic lime. The emission factors applied in the previous submission were rounded values, whereas the ones applied in this submission are unrounded. As a result, the 1990–2008 estimates (compared to the ones shown in the previous submission) have slightly changed by around  $-0.04\%$  (or  $-0.68$  to  $-0.54$  kt CO<sub>2</sub> eq).

#### 4.3.6. Category-Specific Planned Improvements

There are currently no improvements planned for this category.

### 4.4. Limestone and Dolomite Use (CRF Category 2.A.3)

#### 4.4.1. Source Category Description

Limestone is a basic raw material used in a number of industries. In addition to its consumption in the production of cement and lime for resale, limestone is used as a raw material in glass factories. As well, significant amounts of limestone are used as flux stone in iron and steel furnaces and in non-ferrous smelters. Dolomite may also be used in iron and steel furnaces. The proportion of limestone to dolomite used in the iron and steel industry varies depending on the character of iron ore and how the resulting slag is used. Moreover, limestone is used in other areas such as pulp and paper mills (used for makeup lime), flue gas desulphurization (FGD) in fossil fuel-burning power plants and other high-sulphur fuel industrial combustion, and wastewater treatment.

Since limestone at high temperatures is calcined to lime in these industries, CO<sub>2</sub> is produced by the same reaction

<sup>3</sup> Kenefick W. Personal communication (email from Wayne Kenefick to Amy Shen, Greenhouse Gas Division, dated October 22, 2008). Canadian Lime Institute.

described in Section 4.2.1 on cement production.

To avoid double counting, the category of Limestone and Dolomite Use does not include emissions from limestone used for cement and lime production. Emissions from limestone used to produce cement and lime are accounted for under the Cement Production and Lime Production categories, respectively.

#### 4.4.2. Methodological Issues

CO<sub>2</sub> emissions from limestone and dolomite were calculated separately using two different emission factors. The emission estimation method used is considered to be of the Tier 2 type.

Based on the process stoichiometry, it was determined that 440 g of CO<sub>2</sub> could be emitted per kilogram of pure limestone used. However, since there was no pure limestone used in the Canadian industry, a purity fraction of 95% was applied to come up with the overall emission factor of 418 g CO<sub>2</sub>/kg of limestone used (AMEC 2006). The purity fraction of 95% came from a report prepared by the Ministry of Northern Development and Mines (1989) for the Ontario Ministry of Natural Resources.

Dolomite consists of both limestone (CaCO<sub>3</sub>) and magnesite (MgCO<sub>3</sub>). A major Canadian producer of dolomite reported the composition of its dolomite to range from 56% to 58% CaCO<sub>3</sub> and from 38% to 41% MgCO<sub>3</sub>. An overall emission factor of 468 g CO<sub>2</sub>/kg of dolomite used was derived based on the emission factors for pure limestone (440 kg CO<sub>2</sub>/tonne) and magnesite (522 kg CO<sub>2</sub>/tonne), and the assumption that dolomite is composed of 58% CaCO<sub>3</sub> and 41% MgCO<sub>3</sub> (AMEC 2006).

Data on raw stone use in iron and steel furnaces, non-ferrous smelters, glass factories, pulp and paper mills, and other chemical uses were obtained from the *Canadian Minerals Yearbook* (NRCan 1990–2006) or from the author of the stone section of the Yearbook (for 2007–2008). To estimate the stone used by a given sector for 2009, the percentage of change (between 2008 and 2009) in the gross output dollar value<sup>4</sup> for that sector was applied to the 2008 use value. Moreover, data for stone use as flux in iron and steel furnaces for all years were disaggregated into limestone and dolomite based on a 70/30 split (AMEC 2006). National CO<sub>2</sub> emissions were estimated by multiply-

ing the quantities of limestone and dolomite consumed by the corresponding emission factors.

Provincial emission estimates were obtained by apportioning the national emissions according to the sum of the provincial gross output values for the major sectors in which limestone and dolomite were used (i.e. pulp and paper, iron and steel, non-ferrous metal, glass and chemical sectors).

#### 4.4.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Limestone and Dolomite Use. It took into account the uncertainties associated with the use of data by use type and emission factors. The uncertainty associated with the category as a whole for the time series ranged from ±17% to ±19%, with data on the use of limestone and dolomite in the chemical sector and as flux in iron and steel furnaces being the largest contributors.

The same emission factors were consistently applied over the time series. The activity data source is provided in Section 4.4.2.

#### 4.4.4. Category-Specific QA/QC and Verification

This key category in the Industrial Processes Sector has undergone Tier 1 QC checks as elaborated in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No anomalies were observed.

#### 4.4.5. Category-Specific Recalculations

The whole time series (1990–2008) was revised due to corrections made to the applied emission factors. The emission factors used in the previous submission were rounded values, whereas the ones applied in this submission are unrounded. For 1998, 1999, 2003, 2004 and 2008 in particular, the activity data have also been revised (as new data from the author of the *Canadian Minerals Yearbook* have become available and data anomalies have been corrected), resulting in additional changes in these estimates. The 1990–2007 estimates have slightly changed by around

4 Sonnen C. Personal communication (email from Carl Sonnen to Alice Au, Greenhouse Gas Division, dated December 15, 2010). Informetrica Ltd.



-2.75% to 4.90% (or -6.93 to 23.5 kt CO<sub>2</sub> eq).and the 2008 estimate by 75.2% (or 0.123 Mt CO<sub>2</sub> eq).

Moreover, the basis used to allocate emissions by province has been changed. For this submission, gross output values, instead of gross domestic product values, have been used.

#### 4.4.6. Category-Specific Planned Improvements

The emissive portion of the subcategory Other Chemical Uses published in the *Canadian Minerals Yearbook* (NRCan 1990–2006) or obtained from the author of the stone section of the Yearbook, was estimated based on U.S. activity data. There is a plan to improve or refine this estimate by using Canadian activity data.

### 4.5. Soda Ash Production and Use (CRF Category 2.A.4)

#### 4.5.1. Source Category Description

Soda ash (sodium carbonate, Na<sub>2</sub>CO<sub>3</sub>) is a white crystalline solid that is used as a raw material in a large number of industries, including glass manufacture, chemical production, soap and detergents, pulp and paper manufacture, flue gas desulphurization, and wastewater treatment (AMEC 2006). Based on the information on soda ash use by sector in AMEC (2006) and the *Non-Metallic Mineral Products Industries* (Statistics Canada #44-250) publication, it appears that soda ash in Canada is used mainly in the glass products manufacturing industry. CO<sub>2</sub> is emitted as the soda ash decomposes at high temperatures in a glass manufacturing furnace.

CO<sub>2</sub> is also released during the Solvay process in which soda ash is produced. However, as the CO<sub>2</sub> is a necessary component in the carbonation stage of the production process, it is usually recovered and recycled for use.

#### 4.5.2. Methodological Issues

Based on the carbon mass balance, there is one mole CO<sub>2</sub> emitted for each mole of soda ash used. The emission factor (EF) for the mass of CO<sub>2</sub> emitted is estimated based on

the stoichiometry of the chemical process as follows:

Equation 4–2:

$$EF = \frac{(1000\text{g/kg}) \times (44.01\text{g CO}_2/\text{mol})}{(105.99\text{g Na}_2\text{CO}_3/\text{mol})} = 415\text{g CO}_2/\text{kg Na}_2\text{CO}_3$$

National CO<sub>2</sub> emissions were calculated by applying the emission factor of 415 g CO<sub>2</sub>/kg to the national soda ash consumption data, and by assuming that the soda ash used in Canada has a purity of 100%. Quantities of soda ash used were estimated based on soda ash production, import, and export data. Canada stopped its soda ash production in 2001. Production before 2002 was assumed to be equal to the capacity of the only soda ash plant in Canada. Import and export data were obtained from Global Trade Information Services (GTIS 1995–2006, 2007–2009). It should be noted that, since GTIS did not report trade data before 1995, it was assumed that the trade data for the years 1990–1994 were the average of the 1995–2000 trade data. The total quantities of soda ash used were distributed by application type, based on the U.S. pattern of soda ash consumption. According to the U.S. Geological Survey, soda ash can be used in the following sectors: glass, chemical, soaps and detergents, pulp and paper, flue gas desulphurization and others. In the emission estimation, uses of soda ash in soaps and detergents manufacturing, and in water treatment were excluded because these were considered as non-emissive (AMEC 2006).

This method is considered to be Tier 1 type, as it is based on the use of national consumption data and an emission factor derived from the stoichiometry of the process. Methodological issues for calculating CO<sub>2</sub> emissions from soda ash use are not addressed specifically in the IPCC Good Practice Guidance (IPCC 2000).

Provincial emission estimates were obtained by apportioning the national emissions according to provincial gross output values of the sectors in which soda ash was used (i.e. glass, pulp and paper and inorganic chemical sectors).

There is currently no soda ash production in Canada. The only soda ash producing plant, which produced soda ash using the Solvay process, closed in 2001. Although most CO<sub>2</sub> emitted from this facility was recovered for reuse (as mentioned in Section 4.5.1), some CO<sub>2</sub> may have been released from vents on absorbers, scrubbers and distillation units. However, the amount of net CO<sub>2</sub> emissions from

soda ash production in Canada is assumed to be negligible (AMEC 2006).

### 4.5.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Soda Ash Use. It took into account the uncertainties associated with the production data (for years before 2001), import and export data. The uncertainty associated with the category as a whole for the time series ranged from  $\pm 10.2\%$  to  $\pm 13.8\%$ .

The same emission factor was consistently applied over the time series. The activity data source is provided in Section 4.5.2.

### 4.5.4. Category-Specific QA/QC and Verification

The Tier 1 QC checklist was not completed for the category of Soda Ash Use, as it was not a key category. However, several checks that were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000) were done. No anomalies were observed.

### 4.5.5. Category-Specific Recalculations

The 1990–2005 estimates were revised due to corrections made to the applied emission factors. The emission factors used in the previous submission were rounded values, whereas the ones applied in this submission are unrounded. For 2008, the activity data have been revised, as new data on the U.S. pattern of soda ash consumption have become available. The 1990–2005 estimates have slightly changed by around  $-0.3\%$  to  $0.16\%$  (or  $-0.622$  to  $0.342$  kt CO<sub>2</sub> eq) and the 2008 estimate by  $1.4\%$  (or  $2.01$  kt CO<sub>2</sub> eq).

The basis used to allocate emissions by province has also been changed. For this submission, gross output values, instead of gross domestic product values, have been used.

### 4.5.6. Category-Specific Planned Improvements

There are currently no improvements planned specifically for estimating CO<sub>2</sub> emissions from soda ash production and use.

## 4.6. Magnesite Use (CRF Category 2.A.7.2)

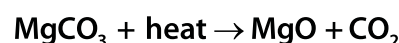
### 4.6.1. Source Category Description

Magnesite, or magnesium carbonate (MgCO<sub>3</sub>), is a silver-white solid that is used as a raw material in the making of a variety of products, including magnesium metal (Mg) and magnesia (MgO).

CO<sub>2</sub> is emitted when magnesite is used during the leaching step of the magnesium production process, as shown below:



Magnesite can also be processed to become lighter-fired caustic magnesia and sintered magnesia, which are then used in refractory manufacturing (AMEC 2006). CO<sub>2</sub> is generated along with magnesia when magnesium carbonate decomposes at high temperatures:



Three facilities in Canada have reported use of magnesite in their processes at different moments during the period 1990–2008. One of them was closed in 1991 and another one in 2007.

### 4.6.2. Methodological Issues

In the *5th Strategic Diversification Newsletter* (SIDEX 2004), one of the facilities (users of magnesite) reported that the purity fraction of the magnesite it used was 97%, and this magnesite was mined by the facility's parent company. Therefore, it was assumed that all three facilities used magnesite with a purity fraction of 97%. Taking the purity of magnesite into account, an overall emission factor of 506 g CO<sub>2</sub>/kg was derived and used in estimating CO<sub>2</sub> emissions from magnesite use.

For the plant that had operated between 1990 and 1991, as no magnesite use data were available, the amount used was backcalculated from the amount of magnesium produced. The amount produced was assumed to be half of 1990 capacity reported in the *Minerals and Metals Foundation Paper*, 1999 (AMEC 2006).

For the other two plants, the 1990–2005 facility-specific



magnesite use data came from British Columbia's Ministry of Energy, Mines and Petroleum Resources (2006) and Environment Canada, Quebec Region, Environmental Protection Branch (2006 email from J. Banville).<sup>5</sup> For 2006 and 2007, activity data were not available; hence, to estimate the use of magnesite for these two plants, some assumptions were made.

For the plant that was closed in 2007, the ratio of magnesite use to magnesium production was first calculated for each year of the 1990–2005 period (2007 email from J. Banville).<sup>6</sup> The average of the calculated (magnesite use / magnesium production) ratios was then taken. This average was multiplied by the plant's 2006 and 2007 magnesium production to yield the 2006 and 2007 magnesite use, respectively.

For the other plant (the only one still in operation in 2009), the 2006–2009 magnesite use data came from British Columbia's Ministry of Energy, Mines and Petroleum Resources.<sup>7</sup>

Finally, multiplying the consumption data (either actual or estimated, depending on the years) by the above-mentioned emission factor gave the national and provincial emission estimates for this subsector.

This method is considered to be of the Tier 1 type, as it is based on the use of national consumption data and an emission factor derived from the stoichiometry of the process.

### 4.6.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Magnesite Use. It took into account the uncertainties associated with the activity data (for years before 2001) and emission factor. The uncertainty associated with the category as a whole for the time series ranging from  $\pm 5.0\%$  to  $\pm 9.0\%$ , with data on the use of magnesite being the largest contributor.

5 Banville J. 2006. Personal communication (email from Banville J to Renata Zaremba, Greenhouse Gas Division, dated March 3, 2006). Environment Canada, Environmental Protection Branch, Quebec Region.

6 Banville J. 2007. Personal communication (email from Banville J. to Maryse Pagé, Greenhouse Gas Division, dated October 4, 2007). Environment Canada, Environmental Protection Branch, Quebec Region.

7 Meredith-Jones S. 2010. Personal communication (email from Meredith-Jones S. to Shanta Chakrovorty, Greenhouse Gas Division, dated June 21, 2010). B.C. Ministry of Energy, Mines and Petroleum Resources. (Also, see link: [http://minfile.gov.bc.ca/report.aspx?f=PDF&r=Production\\_Detail.rpt&minfilno=082JNW001](http://minfile.gov.bc.ca/report.aspx?f=PDF&r=Production_Detail.rpt&minfilno=082JNW001)).

The same emission factor was consistently applied over the time series. The activity data source is provided in Section 4.6.2.

### 4.6.4. Category-Specific QA/QC and Verification

The Tier 1 QC checklist was not completed for the category of Magnesite Use, as it was not a key category. However, several checks that were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000) were done. No anomalies were observed.

### 4.6.5. Category-Specific Recalculations

The 1990–2008 estimates were revised due to a correction made to the applied emission factor. The emission factor used in the previous submission was a rounded value, whereas the one applied in this submission is unrounded. Furthermore, for 2006, the activity data have been revised, as new data on magnesite consumption have become available from the B.C. Ministry of Energy, Mines and Petroleum Resources. The 1990–2005 and 2007–2008 estimates have slightly changed by around  $-0.005\%$  (or around  $-9.48$  to  $-2.85$  tonnes  $\text{CO}_2$  eq) and the 2006 estimate by  $-16.96\%$  (or  $-30.39$  kt  $\text{CO}_2$  eq).

### 4.6.6. Category-Specific Planned Improvements

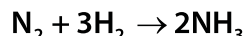
There are currently no improvements planned specifically for estimating  $\text{CO}_2$  emissions from magnesite use.

## 4.7. Ammonia Production (CRF Category 2.B.1)

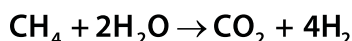
### 4.7.1. Source Category Description

Ammonia ( $\text{NH}_3$ ) is gaseous at standard temperature and pressure. It is toxic and corrosive, and has a pungent odour. Commercially used ammonia is referred to as “anhydrous ammonia,” which must be stored under pressure or at low temperature to remain a liquid. It is used mainly in the production of fertilizers, explosives and polymers.

To produce anhydrous ammonia, nitrogen (N<sub>2</sub>) and hydrogen (H<sub>2</sub>) react together in the Haber-Bosch process. The reaction (as shown below) occurs at high temperature in the presence of a catalyst:

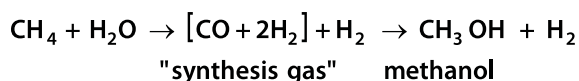


The nitrogen required is obtained from air. The typical source of hydrogen for ammonia plants is the catalytic steam reforming of CH<sub>4</sub> (and minor amounts of other hydrocarbons) contained in natural gas. CO<sub>2</sub> is also generated, as a by-product gas, during the steam methane reforming (SMR) process:



CO<sub>2</sub> is then removed from the process gas by absorption, usually with a solution of monoethanolamine or potassium carbonate (K<sub>2</sub>CO<sub>3</sub>). The primary release of CO<sub>2</sub> occurs during the regeneration (for reuse) of the CO<sub>2</sub>-rich absorption solution by steam stripping or boiling. The stripping gas, which contains CO<sub>2</sub> and other impurities, is then vented to the atmosphere. Alternatively, it can be directed to a neighbouring urea plant, where the CO<sub>2</sub> is recovered and utilized as a feedstock gas. Since the carbon will only be stored for a short period, no account should be taken for intermediate binding of CO<sub>2</sub> in downstream manufacturing processes and products (IPCC/OECD/IEA 1997).

For most Canadian ammonia production facilities, SMR plants are essential units for the operations, because they can generate hydrogen in sufficient quantities to support large-scale ammonia production. However, some plants may use by-product hydrogen to feed into the Haber-Bosch reaction, thereby eliminating release of CO<sub>2</sub> from the ammonia production process. In other words, the hydrogen needed for producing ammonia can also be obtained in ways that do not involve an on-site SMR operation. For instance, at methanol plants, a synthesis gas (or "syn gas") consisting of one part CO and two parts hydrogen is prepared by using a variation of the SMR reaction. The reaction (as depicted below) produces an excess of hydrogen that is more than what is required for methanol production:



This excess of hydrogen is often purged from the methanol plant and used at neighbouring ammonia plants. Also, ethylene plants generate hydrogen as a co-product from

cracking furnaces in making ethylene and other chemicals (e.g. propylene, butadiene). This hydrogen stream can be used at the nearby ammonia plants as well (Cheminfo Services 2006).

## 4.7.2. Methodological Issues

Emissions from ammonia production were estimated by multiplying CO<sub>2</sub>-related ammonia production by an emission factor of 1.56 t CO<sub>2</sub>/t NH<sub>3</sub> produced. The emission factor was developed based on typical energy and material requirements for ammonia production in Canada (Jaques 1992). Data on CO<sub>2</sub>-related production used in the calculation were either directly collected from ammonia plants or estimated. For plants that had used SMR (i.e. releasing CO<sub>2</sub>), but for which data were not available, estimation of production was done based on the reported production and capacity data of other plants and national ammonia production statistics. Data on production and production capacity for 1990–2004, where available, were gathered in a study conducted by Cheminfo Services (2006); those for 2005–2009 were collected by Environment Canada's GHG Division through a voluntary data submission process. Data on national ammonia production were found in Statistics Canada's publication #46-002 for 1990-2007 and in Statistics Canada's CANSIM Table 303-0053 for 2008-2009.

The estimation technique (emissions = production of ammonia × emission factor) is one of the default methods suggested in the revised 1996 IPCC guidelines (IPCC/OECD/IEA 1997). However, it should be noted that the emission factor of 1.56 t CO<sub>2</sub>/t NH<sub>3</sub> produced is a national average value and that plant-specific production data were used whenever possible. Methodological issues for calculating CO<sub>2</sub> emissions from ammonia production are not addressed specifically in the IPCC Good Practice Guidance (IPCC 2000).

Finally, the quantity of natural gas used to produce hydrogen for ammonia production was also recorded by Statistics Canada with all other non-energy uses of natural gas. Therefore, to avoid double counting, the CO<sub>2</sub> emissions from ammonia production were subtracted from the total non-energy fossil fuel use CO<sub>2</sub> emissions reported under the Other and Undifferentiated Production category at the national level. At the provincial and territorial levels, CO<sub>2</sub> emissions from ammonia production are reported together with emissions from Other and Undifferentiated Production under the Other and Undifferentiated category.

Further details with respect to the calculation method used are provided in Annex 3.

### 4.7.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Ammonia Production. It took into account the uncertainties associated with the national and facility-specific ammonia production data and the emission factor. The uncertainty associated with the category as a whole was evaluated at  $\pm 7\%$  and is applicable to all years of the time series.

The same emission factor was consistently applied over the time series. The activity data sources are described in Section 4.7.2

### 4.7.4. Category-Specific QA/QC and Verification

Ammonia production was a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

### 4.7.5. Category-Specific Recalculations

There has been no recalculation for this category.

### 4.7.6. Category-Specific Planned Improvements

Efforts will be made to determine the amounts of natural gas used as feedstock and as fuel during ammonia production over the years and to update the emission factor to better reflect the process-related emissions.

## 4.8. Nitric Acid Production (CRF Category 2.B.2)

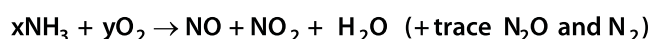
### 4.8.1. Source Category Description

Nitric acid ( $\text{HNO}_3$ ) is a highly corrosive and toxic inorganic

compound that is used mainly as a raw material in the manufacture of synthetic commercial fertilizer. It can also be used in the production of adipic acid and explosives, in metal etching, and in the processing of ferrous metals (IPCC/OECD/IEA 1997).

The production of nitric acid is a two-stage process involving catalytic oxidation of ammonia ( $\text{NH}_3$ ) to nitrogen dioxide ( $\text{NO}_2$ ) and then subsequent formation of nitric acid by addition of water ( $\text{H}_2\text{O}$ ) to  $\text{NO}_2$ . As shown below, the first stage is the reaction of ammonia gas with oxygen ( $\text{O}_2$ ) (from air) at high temperatures:

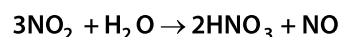
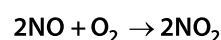
Stage 1 (Reaction):



The hot gases pass through many sheets of wire gauze catalyst. These are usually made from platinum, palladium, gold, or rhodium alloy wire knitted into a fine mesh surface. The reaction products are a mixture of nitric oxide ( $\text{NO}$ ),  $\text{NO}_2$ , and water vapour, with trace amounts of  $\text{N}_2\text{O}$  and nitrogen ( $\text{N}_2$ ) (Cheminfo Services 2006). An excess of oxygen may drive the  $\text{NO}$  to be converted to  $\text{NO}_2$ . Nitrogen oxidation steps under reducing conditions are sources of  $\text{N}_2\text{O}$ . More specifically,  $\text{NO}$ , an intermediate in the production of nitric acid, can readily decompose to  $\text{N}_2\text{O}$  and  $\text{NO}_2$  at high pressures and at a temperature range of 30–50°C (IPCC/OECD/IEA 1997).

During the second stage of the production process, water is added at the top of an absorber tower to hydrate the  $\text{NO}_2$  and to scrub the gases. As shown below, hydration of cooled  $\text{NO}_2$  with water forms a 60–65% solution of nitric acid, leaving the bottom of the tower. To complete the conversion of  $\text{NO}$  to  $\text{NO}_2$ , excess air (oxygen) is introduced at the bottom tray of the absorber tower. The  $\text{NO}_2$  formed is also absorbed.

Stage 2 (Absorption):



Since the hydration reaction is exothermic, the absorber towers require cooling, and some of them have a cooling circuit on each tower tray. The typical conversion yield to nitric acid is 93% if a fresh reaction catalyst is used. As the catalyst ages and degrades, conversion can fall to about 90%. The tail gases that leave the absorber tower consist mostly of nitrogen, a small concentration of oxygen, and trace quantities of  $\text{N}_2\text{O}$ ,  $\text{NO}$ ,  $\text{NO}_2$ , and other nitrogen oxides

(NO<sub>x</sub>). The concentration of N<sub>2</sub>O in the exhaust gases depends on the type of plant and its emission controls (Cheminfo Services 2006).

There exist two basic types of nitric acid production technology: high pressure and dual pressure. Both technologies can be found in Canadian nitric acid plants. The high-pressure design, commonly used in North America, applies a single pressure throughout the reaction and absorption stages. High-pressure process plants can function with a non-selective catalytic reduction (NSCR) or selective catalytic reduction (SCR) system. The emission abatement systems are classified as “non-selective” when natural gas is used as reductant to reduce all NO<sub>x</sub>. In contrast, a “selective” catalytic reduction (SCR) uses ammonia, which selectively reacts only with NO and NO<sub>2</sub> gases, and not with N<sub>2</sub>O (hence a higher N<sub>2</sub>O emission factor). Most Canadian plants operate with a high-pressure design and have NSCR abatement technology installed (Cheminfo Services 2006).

The second type of nitric acid production technology design, dual pressure, was developed in Europe. This older technology uses low pressure for the reaction stage and higher pressure for the absorption stage. To increase the efficiency of the absorption stage, dual-pressure plants can “extend” the absorption tower by adding more trays. This is referred to in Table 4- 2 as “absorption Type 1.” Alternatively, plants can have in place a second tower to allow “double absorption.” This is referred to in Table 4- 2 as “absorption Type 2” (Cheminfo Services 2006).

## 4.8.2. Methodological Issues

Data supporting the estimation of N<sub>2</sub>O emissions from nitric acid production for 1990–2004 were gathered through a study conducted for Environment Canada (Cheminfo Services 2006), and those for 2005–2009 were obtained by the Department’s GHG Division from industry through

a voluntary data submission process. The collected data were used in the country-specific hybrid emission estimation methodology, which could be described as three categories:

1. Plant-specific production data and plant-specific emission factors (i.e. Tier 3 type method) when these were available from companies; or
2. Plant-specific production data and production technology-specific emission factors that are national average values (i.e. Tier 2 type method) when plant-specific emission factors were not available; or
3. Estimated production data and national average technology-specific emission factors (i.e. Tier 1 type method) when limited or no plant-specific data were available.

In all three scenarios, the equation applied was as follows:

Equation 4–3:

$$\text{N}_2\text{O Emission (t)} = \text{Production - Based Emission Factor (kg N}_2\text{O / t HNO}_3\text{)} \times \text{Production (kt HNO}_3\text{)}$$

To estimate emissions in scenarios 2 and 3, the types of production process and emission control technology of a plant were first determined. The reported or estimated production was then multiplied by the corresponding emission factor. The industry-typical emission factors used had been obtained from the Canadian Fertilizer Institute in the early 1990s (G. Collis 1992).<sup>8</sup> These were confirmed again, as being applicable, by industry representatives during the recent (Cheminfo Services 2006) study. In addition, another industry-typical emission factor is provided in the IPCC Good Practice Guidance (IPCC 2000) and was confirmed through the same study (Cheminfo Services 2006).

<sup>8</sup> Collis G. 1992. Personal communication (letter from Collis G. to Director, Greenhouse Gas Division, dated March 23, 1992). Canadian Fertilizer Institute.

Table 4–2 Nitric Acid Industry–Typical Emission Factors

Type of Production Process Technology	Type of Emission Control Technology	Emission Factor (kg N <sub>2</sub> O/t HNO <sub>3</sub> )	Data Source
Dual Pressure	Extended Absorption “Type 1”	9.4	1992 letter from G. Collis <sup>1</sup>
Dual Pressure	Extended Absorption “Type 2”	12	1992 letter from G. Collis <sup>1</sup>
High Pressure	NSCR	0.66	1992 letter from G. Collis <sup>1</sup>
High Pressure	SCR	8.5	IPCC (2000)

<sup>1</sup> Collis G. 1992. Personal communication (letter from Collis G. to Director, Greenhouse Gas Division, dated March 23, 1992). Canadian Fertilizer Institute.

Table 4–2 summarizes the industry-typical emission factors by process and control types.

For plants that did not have production data available, production was estimated based on the national production data from Statistics Canada (# 46-002), and the reported production data and the production capacity data of other plants. More specifically, the sum of all production reported by companies was deducted from the national total production (Statistics Canada # 46-002) to give the national unreported nitric acid production. The latter was then allocated based on the capacities of the non-reporting plants to give the estimated production values for the non-reporting plants. The estimated production was multiplied by what was believed to be the most appropriate industry-typical emission factor to estimate emissions coming from plants for which no or few data were available. For 1990–2004, the raw activity data and plant-specific emission factors (when available) used to develop emission estimates were collected through the 2006 Cheminfo study (Cheminfo Services 2006). For 2005–2009, the data used were reported by companies to Environment Canada's Greenhouse Gas Division on a voluntary basis.

Finally, the estimates of N<sub>2</sub>O emissions (by plant) were summed either all together to yield the national emission estimate or by province to give the provincial emission estimate.

#### 4.8.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Nitric Acid Production. It took into account the uncertainties associated with the national and facility-specific nitric acid production data and the emission factors. The uncertainty associated with the category as a whole was evaluated at  $\pm 10\%$ , with the emission factors being the largest contributors. The uncertainty value is applicable to all years of the time series.

The same emission factors were consistently applied over the time series. The activity data source is provided in Section 4.8.2.

#### 4.8.4. Category-Specific QA/QC and Verification

Nitric acid production was a category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

#### 4.8.5. Category-Specific Recalculations

For the years 1992 and 1997, activity data transcription errors were detected and corrected. These corrections have resulted in recalculations. The 1992 estimate was increased by 2.68% (or 0.0284 Mt CO<sub>2</sub> eq); the 1997 estimate was decreased by 0.051% (or – 0.543 kt CO<sub>2</sub> eq).

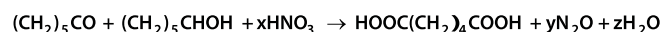
#### 4.8.6. Category-Specific Planned Improvements

There are currently no improvements planned for this category.

### 4.9. Adipic Acid Production (CRF Category 2.B.3)

#### 4.9.1. Source Category Description

Adipic acid (HOOC(CH<sub>2</sub>)<sub>4</sub>COOH) is a dicarboxylic acid used primarily in the production of Nylon 66, resins, and plasticizers. It is produced via a two-stage oxidation process. The first step involves the oxidation of cyclohexane to form a cyclohexanone ((CH<sub>2</sub>)<sub>5</sub>CO) / cyclohexanol ((CH<sub>2</sub>)<sub>5</sub>CHOH) mixture. The mixture is then oxidized by a 50–60% nitric acid solution in the presence of a catalyst (e.g. vanadium or copper) to form adipic acid. N<sub>2</sub>O is generated as a by-product in the second oxidation reaction, as shown below:



Emissions of N<sub>2</sub>O from this manufacturing process depend on both the amount generated and the amount that can potentially be destroyed in any subsequent abatement process. When emission abatement equipment is not installed at a facility, the N<sub>2</sub>O generated is generally vented



to the atmosphere in a waste gas stream. Adipic acid production also results in emissions of non-methane volatile organic compounds (NMVOC), CO, and NO<sub>x</sub> (IPCC/OECD/IEA 1997). Emissions of these indirect GHGs are not covered in this section. Annex 10 provides details on indirect GHG emissions.

Invista Canada, formerly Dupont Canada, located in Maitland, Ontario, had operated the only adipic acid production facility in Canada. It had significantly reduced its N<sub>2</sub>O emissions since 1997, when a catalytic N<sub>2</sub>O abatement system with an emission monitoring system was started up. The plant has, however, become indefinitely idled as of spring 2009.

### 4.9.2. Methodological Issues

Emission estimates for adipic acid production have always been provided by Invista. For the 1990–1996 period, when no emission controls were in place, the reported emission estimates were calculated by simply multiplying the annual adipic acid production by the IPCC default generation factor of 0.3 kg N<sub>2</sub>O/kg adipic acid.

As mentioned above, in 1997, Invista installed an N<sub>2</sub>O abatement system with a continuous emission monitor on the controlled off-gas stream at the abatement system outlet. Since then, the emission estimation method applied by Invista has become the following:

Equation 4–4:

$$\text{TotalEmissions(t)} = \text{N}_2\text{O Emissions(t) with abator} + \text{N}_2\text{O Emissions(t) without abator}$$

The first term accounts for emissions that occur when the abator is operating, and the second for emissions that occur when the abator is not operating because of maintenance or technical problems.

N<sub>2</sub>O Emissions with Abator:

Equation 4–5:

$$\text{N}_2\text{O Emissions(t) with Abator} = \left( \text{Production(t)} \times \left( \frac{0.3 \text{ t N}_2\text{O}}{\text{t adipic acid}} \right) \times (1 - \text{Destruction Efficiency}) \times (\text{Abatement Utilization Ratio}) \right)$$

where:

Destruction Efficiency is determined based on the difference between the amount of N<sub>2</sub>O entering the abatement unit and that leaving the unit. It is a monthly average calculated using values recorded by analyzers, which are located at the inlet and outlet of the abator. The targeted instantaneous destruction efficiency is 97%.

N<sub>2</sub>O Emissions(t) with Abator=

$$\left( \text{Production(t)} \times \left( \frac{0.3 \text{ t N}_2\text{O}}{\text{t adipic acid}} \right) \times (1 - \text{Destruction Efficiency}) \times (\text{Abatement Utilization Ratio}) \right)$$

Abatement Utilization Ratio is the number of hours during which N<sub>2</sub>O goes through the abator divided by the total operating time.

N<sub>2</sub>O Emissions without Abator:

Equation 4–6:

N<sub>2</sub>O Emissions(t) without Abator=

$$\left( \text{Production(t)} \times \left( \frac{0.3 \text{ t N}_2\text{O}}{\text{t adipic acid}} \right) \times (1 - \text{Abatement Utilization Ratio}) \right)$$

It is important to note that the in-line continuous emission monitor has never been used to directly monitor net N<sub>2</sub>O emissions. This is because the analyzer is limited to accurately measuring relatively low concentrations of N<sub>2</sub>O only when the reactor is online and abating N<sub>2</sub>O gas. The analyzer is not capable of measuring the full range of N<sub>2</sub>O concentrations that could potentially exist in the stack. The N<sub>2</sub>O concentration can vary from a low nominal level of 0.3% when the stream leaves the abator to a high nominal level of 35–39% N<sub>2</sub>O in the unabated stream. When the abatement reactor is bypassed, there is no N<sub>2</sub>O abatement occurring, and the analyzer will not record N<sub>2</sub>O stack emissions (Cheminfo Services 2006).

The calculation techniques used to estimate emissions for the periods 1990–1997 and 1998–2009 are basically the same as the default methods presented in the IPCC Good Practice Guidance (IPCC 2000) and the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997).

### 4.9.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Adipic Acid Production. It took into account the uncertainties associated with the adipic acid production data, the emission factor, the destruction efficiency and the abatement utilization factor. The uncertainty associated with the category as a whole was evaluated at ±11%, with the emission factor being the largest contributor. The uncertainty value is applicable to all years of the time series.

As explained in Section 4.9.2, two methods were applied in the time series: one for the period of time during which



the plant operated *with* the emission abatement system and another for the period of time during which the plant operated *without* the emission abatement system.

#### 4.9.4. Category-Specific QA/QC and Verification

Adipic acid production was a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

#### 4.9.5. Category-Specific Recalculations

There were no recalculations of  $N_2O$  emissions related to adipic acid production.

#### 4.9.6. Category-Specific Planned Improvements

There are currently no improvements planned specifically for this category.

### 4.10. Petrochemical Production – Carbide Production (CRF Category 2.B.4)

#### 4.10.1. Source Category Description

Two kinds of carbide are considered in this section: silicon carbide (SiC) and calcium carbide ( $CaC_2$ ). SiC and  $CaC_2$  are no longer produced in Canada, since the last of two SiC plants closed in 2002 and the only  $CaC_2$  plant closed in 1992. Silicon carbide (SiC) is widely used as an abrasive and is also increasingly being used in electronics. It is produced by reducing silicon quartz ( $SiO_2$ ) using carbon as a reducing agent (reductant). Petroleum coke is usually used as a carbon source because of its high carbon content.

Calcium carbide ( $CaC_2$ ) is produced mainly as a precursor for the production of acetylene ( $CH\equiv CH$ ). It is produced by reducing lime (CaO) using carbon as the reductant.

Petroleum coke is commonly used as the carbon source. Approximately 67% of the carbon from the petroleum coke used is bound in the product.

The use of coke as a reductant has the potential to release small amounts of methane ( $CH_4$ ) gas in the high operating temperatures (1600–2500°C) of the electric resistance reduction furnaces used for carbide production.  $CH_4$  can be released directly from the decomposition of coke (which still contains trace levels of methane absorbed in its structure) and can also be generated from the thermal decomposition (in the presence of hydrogen) of trace volatile compounds still contained in coke. Most  $CH_4$  is likely to be released in the initial stages of carbide reduction when the coke is at high temperatures, particularly from the top layers of coke exposed directly to the atmosphere. When coke is manufactured from coal in coke ovens, most of the volatile matter in the coal is driven off as raw coke oven gas and recovered as liquids and fuel gas. Once the tars, liquid oils and ammonia are removed from raw coke oven gas, the remaining coke oven gas typically contains 60% hydrogen and 25% methane. This is an indication that: i)  $CH_4$  is present in volatile organic matter contained in coal; and ii) significant hydrogen is present to contribute to methane formation from the thermal decomposition of heavier volatile compounds. The coking process removes the vast majority of volatile matter from coal, but the large solid masses of coke still contain trace amounts of volatile matter (Cheminfo Services 2010).

#### 4.10.2. Methodological Issues

To estimate  $CH_4$  emissions from carbide production at national and provincial/territorial levels, a Tier 1 method (i.e. with the application of Tier 1 IPCC default emission factors) was applied. Since no survey of active facilities was possible, research was conducted by Cheminfo Services, on behalf of the GHG Division, to identify and establish the production capacities of the three carbide production facilities. A time series of process  $CH_4$  emissions was estimated for two silicon carbide facilities and one calcium carbide facility from 1990 to 2001 based on assumed capacity utilization and  $CH_4$  emission factors. Only (SiC and  $CaC_2$ ) production capacity data during the period 1990–2001 were identified during the study. As such, the following equation was used to estimate total  $CH_4$  emissions from carbide production:

Equation 4–7:

$$\sum_y [(SiC \text{ capacity} \times \text{capacity utilization} \times \text{Emission Factor}_{SiC}) + (CaC_2 \text{ capacity} \times \text{capacity utilization} \times \text{Emission Factor}_{CaC_2})]$$

where:

y	=	companies
SiC or CaC <sub>2</sub> capacity	=	data collected from the industry, kt
Capacity utilization	=	based Cheminfo Services's knowledge of the industry, %
Emission Factor <sub>SiC</sub>	=	11.6 kg CH <sub>4</sub> /t SiC (IPCC 2006)
Emission Factor <sub>CaC<sub>2</sub></sub>	=	4.8 kg CH <sub>4</sub> /t CaC <sub>2</sub> , derived from CH <sub>4</sub> emission factor for silicon carbide and the ratio of IPCC default Calcium Carbide CO <sub>2</sub> emission factor to IPCC default Silicon Carbide CO <sub>2</sub> emission factor (i.e. 11.6 (kg CH <sub>4</sub> /t SiC) * (1.09 tCO <sub>2</sub> /tCaC <sub>2</sub> / 2.62 tCO <sub>2</sub> /tSiC))

### 4.10.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Carbide Production (Cheminfo Services 2010), following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below  $\pm 30\%$ .

Since there is no longer carbide production in Canada, a set of default uncertainties (based on Cheminfo Services' knowledge of the industry) was used in the Cheminfo Services' Tier 1 uncertainty assessment. Regarding the carbide capacity data, an uncertainty of  $\pm 5\%$  was applied when survey uncertainties were not provided. The uncertainty associated with the category as a whole for the time series ranged from  $\pm 0\%$  to  $\pm 27\%$  (Cheminfo Services 2010).

### 4.10.4. Category-Specific QA/QC and Verification

The category of Carbide Production has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 General Inventory Level QC Procedure outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of impor-

tance were detected from the Tier 1 QC process.

### 4.10.5. Category-Specific Recalculations

CH<sub>4</sub> from Carbide Production is a new category for the 2011 submission.

### 4.10.6. Category-Specific Planned Improvements

There are currently no improvements planned specifically for this category, as Canadian carbide production stopped in 2002.

## 4.11. Petrochemical Production – Carbon Black Production (CRF Category 2.B.5.1)

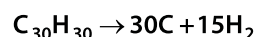
### 4.11.1. Source Category Description

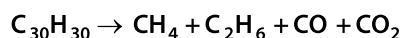
Carbon black is a solid product consisting of fine particles of carbon with small amounts of inorganic impurities. Carbon black is used in rubber compounding (e.g. tires), pigments, printing inks, and in many other additives. There are four facilities that have produced carbon black in Canada since 1990. Three facilities are currently operating.

Carbon black is made from the pyrolysis or thermal cracking of various hydrocarbon feedstocks. Cracking reactions, which separate the hydrogen from the carbon, occur between 1200°C and 1600°C. There are two main carbon black processes used in Canada: 1) pyrolysis of liquid hydrocarbon feedstock; and 2) pyrolysis of natural feedstock, from which CH<sub>4</sub> can be emitted.

During pyrolysis of liquid hydrocarbon feedstock in the furnace black process, hydrogen atoms are separated from the carbon to yield the carbon black product particles (Cheminfo Services 2010).

Furnace Black Pyrolysis:



**Side Reactions:**

The vent gas from the furnace black process contains different products as hydrogen, methane, acetylene, carbon monoxide, carbon dioxide, sulphur compounds, carbon-sulphur compounds and water (IPCC 2006). A portion of the tail gas is burned for energy recovery for the process.

During pyrolysis of natural feedstock, purchased natural gas is injected into one of a pair of reactors that is preheated to a temperature of 1300°C, and decomposed to solid carbon particles, hydrogen and residual hydrocarbons.

**Thermal Black Pyrolysis:**

The mixture of solid carbon particles and hydrogen-rich by-product gas is cooled with water injection and the carbon is separated from the hydrogen by-product gas in a baghouse. The hydrogen by-product gas is used as fuel to preheat the second reactor of the unit up to the reaction temperature of 1300°C. As the process of producing the carbon black consumes energy, the first reactor cools to a point where the reaction becomes inefficient. This is when the production mode is switched, reactors trade roles and the second one, now hot, becomes the producing reactor while the first one is reheated (Cheminfo Services 2010).

### 4.11.2. Methodological Issues

To estimate CH<sub>4</sub> emissions from carbon black production, a consulting study has been performed by Cheminfo Services (2010). A survey was sent to the three operating carbon black facilities requesting 1990–2009 data on carbon black capacity and production, and on process GHG emissions. All three facilities reported 1990–2009 data for carbon black capacity, but not all facilities reported process CH<sub>4</sub> emissions. From the received responses, two facility-specific Tier 3 emission factors were derived as weighted averages of the reported 2007–2009 data. One sector-wide process CH<sub>4</sub> emission factor was also calculated as a weighted average based on the same set of data reported by the two facilities. It was applied when facility-specific emission factor could not be used. When process emissions were reported directly by a facility, the reported data were used in the inventory. When reported emission data were not available, estimates were calculated based on the unreported carbon black production (allocated to each non-reporting facility by its share of capacity) and the

Tier 3 sector average emission factor (either facility-specific or sector-wide). The unreported carbon black production was calculated from total national carbon black production less the sum of all reported carbon black production. National carbon black production data were taken from Camford's CPI Product Profile for 1990–1995 and company-reported production for 2007–2009. Interpolations were made for years in between (i.e. 1996–2006) based on a sector average growth rate for 1990–1994. The total sector production for each year of 1996–2006 was calculated by multiplying the sector average growth rate by the total sector production of the preceding year (starting from 1995).

### 4.11.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed by Cheminfo Services for the category of Carbon Black Production following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below ±30%.

In the Cheminfo Services (2010) study, respondents were asked to provide their best estimate of the uncertainty of each variable reported. Very few survey respondents provided uncertainty estimates for their data. As such, a set of default uncertainties (based on Cheminfo Services' knowledge of the industry) was used in the analysis.

The following default uncertainties were applied when survey uncertainties were not provided:

- capacity data: ± 5%;
- reported production data: ±2%;
- capacity share fractions used for allocation of national production data: ±10%;
- reported process CH<sub>4</sub> emissions: ±20%; and
- reported process N<sub>2</sub>O emissions: ±30%.

The Tier 1 uncertainty associated with the CH<sub>4</sub> emission estimates ranged from ±9% to ±11%.

### 4.11.4. Category-Specific QA/QC and Verification

This key category in the Industrial Processes Sector has undergone Tier 1 QC checks as elaborated in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 General Inventory Level QC Procedures

outlined in the IPCC Good Practice Guidance (IPCC 2000). No anomalies were observed.

### 4.11.5. Category-Specific Recalculations

CH<sub>4</sub> from Carbon Black Production is a new category for the 2011 submission.

### 4.11.6. Category-Specific Planned Improvements

Efforts will be made to improve the accuracy of the CH<sub>4</sub> estimates for the 1996–2006 period by using facility-specific carbon black data that are to be acquired by Environment Canada from Statistics Canada.

## 4.12. Petrochemical Production – Ethylene Production (CRF Category 2.B.5.2)

### 4.12.1. Source Category Description

There have been five ethylene facilities operated by four companies in Canada since 1990, but one of these, Pétromont Varennes, was shut down in 2008. Ethylene is the key building block in the basic organic chemicals industry since it is a precursor to several high-volume industrial chemicals and resins. The most important ethylene derivatives include polyethylene, ethylene oxide/ethylene glycol, ethylene dichloride/vinyl chloride, styrene, linear alpha olefins, vinyl acetates and alcohols.

Ethylene is produced by thermally cracking natural gas liquids (NGLs) such as ethane (C<sub>2</sub>H<sub>6</sub>), propane (C<sub>3</sub>H<sub>8</sub>), or butanes (C<sub>4</sub>H<sub>10</sub>) or petroleum-based liquid feedstocks, such as naphthas or gas oils. The production of ethylene is a two-stage process in which the first stage is to thermally crack the bonds in the raw materials and the second stage is to separate the products of the thermal cracking reaction through distillation. The cracking reaction occurs in specially designed high-temperature, tubular furnaces. Each of the feedstocks requires different amounts of energy per unit of ethylene produced (or amount of feedstock used). Therefore, each feedstock has different GHG emission

intensities.

Process CH<sub>4</sub> emissions from ethylene production come mainly from combustion of process off-gases, flaring of process materials containing methane, and fugitive emissions of volatile hydrocarbon streams that contain methane. Process N<sub>2</sub>O emissions come mainly from the combustion of fuel gas derived from the feedstock (Cheminfo Services 2010).

### 4.12.2. Methodological Issues

A consulting study has been performed by Cheminfo Services Inc to estimate CH<sub>4</sub> and N<sub>2</sub>O emissions from ethylene production. Cheminfo Services, on behalf of the GHG Division, sent a questionnaire to the four companies that have had ethylene production operation in Canada, requesting 1990–2009 data on ethylene capacity and production, and on process CH<sub>4</sub> and N<sub>2</sub>O emissions. Responses were received from two companies for three of the four operating plants, representing 90% of Canadian ethylene capacity in 2009. Sector-wide CH<sub>4</sub> and N<sub>2</sub>O GHG emission factors were estimated as weighted averages based on the reported process emissions and production data from three facilities for 2007–2009. When possible, for 1990–2009, weighted average facility-specific process GHG emission factors were developed and applied to estimated facility ethylene production, because there was a significant difference between the calculated emission factors for each facility.

When process GHGs were reported directly by a facility, the reported data were used in the inventory. When reported emission data were not available, emissions were estimated based on the unreported ethylene production (allocated to each non-reporting facility by share of capacity) and the corresponding emission factors. The unreported production was calculated by subtracting the sum of reported production by the total national production. National ethylene production data were taken from Camford's CPI Product Profile for 1990–1995 and company-reported production for 2007–2009. It should be noted that the emission factors applied should be kept confidential, as they were derived from business-sensitive data.

Equation 4–8:

$$\text{CH}_4 \text{ or N}_2\text{O emissions (t)} = \sum^y \text{Reported emissions} + (\text{allocated unreported production} \times \text{Emission Factor}_{\text{CH}_4/\text{N}_2\text{O}})$$

where:

y	=	companies
allocated unreported production (kt)	=	remaining unreported ethylene production x ethylene capacity of a specific company / total unreported ethylene capacity
remaining unreported ethylene production (kt)	=	total production – total reported sample

### 4.12.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed by Cheminfo Services (2010) for the category of Ethylene Production following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below  $\pm 30\%$ .

In the Cheminfo Services (2010) study, respondents were asked to provide their best estimate of the uncertainty of each variable reported. Very few survey respondents provided any uncertainty estimates for their data. As such, a set of default uncertainties (based on Cheminfo Services' knowledge of the industry) was used in the analysis. The following default uncertainties were applied when survey uncertainties were not provided:

- capacity data:  $\pm 5\%$ ;
- reported production data:  $\pm 2\%$ ;
- capacity share fractions used for allocation of national production data:  $\pm 10\%$ ;
- reported process  $\text{CH}_4$  emissions:  $\pm 20\%$ ; and
- reported process  $\text{N}_2\text{O}$  emissions:  $\pm 30\%$ .

The uncertainties for the time series ranged from  $\pm 8\%$  to  $\pm 12\%$  for  $\text{CH}_4$  emission estimates and from  $\pm 12\%$  to  $\pm 21\%$  for  $\text{N}_2\text{O}$  emission estimates.

### 4.12.4. Category-Specific QA/QC and Verification

This category has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 general inventory level QC procedure outlined in the IPCC Good Practice Guidance (IPCC 2000). A transcription error (for the 1992  $\text{N}_2\text{O}$  estimate) was detected from the Tier 1 QC process. Due to a file linkage problem, the 1992  $\text{N}_2\text{O}$  estimate for ethylene production has not been correctly transcribed to the national summary table shown in Annex 12, resulting in an unusually large emission value for Petrochemical Production as a whole. The correct 1992  $\text{N}_2\text{O}$  emission value for ethylene production should be 0.027 kt  $\text{N}_2\text{O}$  (or 8.3 kt  $\text{CO}_2$  eq). The correct 1992  $\text{N}_2\text{O}$  emission estimate for Petrochemical Production should be 0.027 kt  $\text{N}_2\text{O}$  (or 8.3 kt  $\text{CO}_2$  eq) as well. As such, the total 1992 emission estimate for Petrochemical Production should be 98 kt  $\text{CO}_2$  eq, not 7433 kt  $\text{CO}_2$  eq as shown. To maintain consistency between the NIR and the CRF tables, the same 1992 estimate is inputted to the CRF Reporter.

### 4.12.5. Category-Specific Recalculations

Ethylene Production is a new category for the 2011 submission.

### 4.12.6. Category-Specific Planned Improvements

The transcription error noted for the 1992  $\text{N}_2\text{O}$  estimate will be corrected in the next inventory submission.

## 4.13. Petrochemical Production – Ethylene Dichloride (EDC) Production (CRF Category 2.B.5.3)

### 4.13.1. Source Category Description

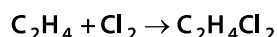
Ethylene dichloride (EDC,  $\text{C}_2\text{H}_4\text{Cl}_2$ ) is the old name for 1,2-dichloroethane, a large-volume chlorinated hydrocar-



bon intermediate derived from ethylene that is used in the manufacture of vinyl chloride monomer (VCM,  $C_2H_3Cl$ ,  $CH_2=CHCl$ ), the precursor of polyvinyl chloride resins. Three EDC production facilities had operated in Canada during different periods of time between 1990 and 2009, but they are all closed now. The last one was shut down in 2006.

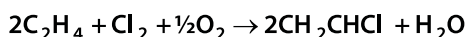
Two processes had been used for the production of EDC in Canada. One is the direct chlorination of ethylene in a vapour or liquid phase reaction using ethylene dibromide as catalyst.

Direct Chlorination:

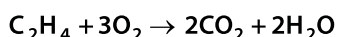


The second process is called oxychlorination. In this process the ethylene feed is chlorinated in the liquid phase. EDC vapours produced are condensed, degassed and a portion is withdrawn from the unit. The balance of the feed is passed onto the vinyl chloride section of a production plant where hydrochloric acid and oxygen in a fluidized catalyst bed reactor produce crude EDC. The crude EDC from both streams is combined and cracked in a reactor to form vinyl chloride with the by-product hydrochloric acid being recycled back into the oxychlorination unit.

Balanced EDC/VCM Reaction:



Feedstock Oxidation Reaction:



Regarding emissions, the process off-gas that contains the chlorinated hydrocarbons is combusted within the plant prior to release, so any carbon in this off-gas is converted to  $CO_2$ . The process  $CO_2$  emissions from EDC production come from the side reaction of feedstock oxidation. The process  $CH_4$  emissions would most likely come from light hydrocarbons from distillation operations that are not captured by a flare gas recovery system. These emissions are vented to the atmosphere (Cheminfo Services 2010).

### 4.13.2. Methodological Issues

Cheminfo Services Inc has been retained by the GHG Division to estimate 1990–2009  $CH_4$  emissions from EDC production. Since all EDC plants are now closed and no survey response can be provided, a Tier 1 calculation ap-

proach (i.e. annual production \* Tier 1 IPCC default emission factor) was taken to develop 1990–2006 process  $CH_4$  emission estimates. The annual EDC production data come from the Canadian C2+ Petrochemical Report, which was obtained via the Cheminfo Services (2010) study. The default process  $CH_4$  emission factor for EDC as applied comes from Table 2-10 of the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997), under the name “dichloroethylene.” The Canadian C2+ Petrochemical Report was prepared and published by an independent consultant who supplies market intelligence to the Canadian chemical industry. It provides balances of ethylene and its derivatives using total production, dispositions, and Canadian trade statistics. For the purpose of emission estimation at the provincial level, the annual EDC production was allocated by Cheminfo Services to each plant based on the capacity share (calculated from production capacity data reported by companies during the Cheminfo Services [2010] study).

### 4.13.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed by Cheminfo Services (2010) for the category of EDC production following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below  $\pm 30\%$ .

As no plant-specific uncertainty estimates could be collected by Cheminfo Services (2010), a set of default uncertainties (based on Cheminfo Services' knowledge of the industry) was used in the analysis. The uncertainty associated with the category as a whole for the time series is estimated at  $\pm 21\%$  (Cheminfo Services 2010).

### 4.13.4. Category-Specific QA/QC and Verification

This category has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 general inventory level QC procedure outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.



### 4.13.5. Category-Specific Recalculations

CH<sub>4</sub> from EDC Production is a new category for the 2011 submission.

### 4.13.6. Category-Specific Planned Improvements

There are currently no improvements specifically planned for this category, as there is no more EDC production in Canada.

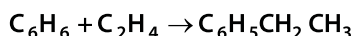
## 4.14. Petrochemical Production – Styrene Production (CRF Category 2.B.5.4)

### 4.14.1. Source Category Description

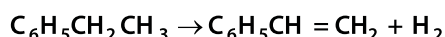
Styrene is a cyclic hydrocarbon liquid that is the precursor monomer for polystyrene and several copolymers, such as ABS resins, styrene-butadiene (SBR) rubber, styrene-butadiene latex, and styrene-acrylonitrile resin (SAN). These materials are used in rubber, plastic, insulation, fiberglass, pipes, automobile and boat parts, food containers, and carpet backing. There have been three styrene facilities that have produced styrene in Canada since 1990, but one facility closed in 1998.

Styrene is an organic chemical intermediate that is manufactured in a two-stage process in the same industrial facility. The first stage involves the alkylation of benzene (C<sub>6</sub>H<sub>6</sub>) with ethylene (C<sub>2</sub>H<sub>4</sub>) to produce ethylbenzene, a chemical intermediate. The second stage is the catalytic dehydrogenation of the ethylbenzene to produce styrene, as shown below:

Ethylbenzene Reaction:



Styrene Reaction:



The styrene is produced in multiple sequential dehydrogenation reactors operated under vacuum to increase

conversion and selectivity towards styrene. Steam is used in the dehydrogenation reaction to provide the reaction energy, dilute the reactants and remove any coke build-up from the catalyst. The by-products from the reaction include process off-gas, benzene (recycled) and toluene, steam condensate, and tars, which must be separated from the product stream.

Process CO<sub>2</sub> emissions can come from the combustion of the process off-gas (fuel gas) as fuel or from flaring of over-pressured process streams. Methane (CH<sub>4</sub>) could be present along with the process reactants ethylene and benzene and would be emitted if there was any venting of these process or recycle streams. Fugitive emissions from these streams would also contain methane (Cheminfo Services 2010).

### 4.14.2. Methodological Issues

Cheminfo Services Inc has been retained by the GHG Division to estimate 1990–2009 CH<sub>4</sub> emissions from styrene production. A survey was sent to the two operating facilities, but none of them provided a response with respect to emission estimates. Since survey responses were not available, a Tier 1 calculation approach (i.e. annual production \* Tier 1 IPCC default emission factor) was taken to develop process CH<sub>4</sub> emissions estimates. Annual styrene production data come from the Canadian C2+ Petrochemical Report, which was obtained via the Cheminfo (2010) study. This report was prepared and published by an independent consultant to supply market intelligence to the Canadian Chemical Industry. It provides balances of ethylene and its derivatives using total production, dispositions, and Canadian trade statistics. For the purpose of emission estimation at provincial level, the annual styrene production was allocated by Cheminfo Services to each plant based on capacity share (calculated from production capacity data reported by companies during the Cheminfo [2010] study). The default process CH<sub>4</sub> emission factor for styrene (4 kg/t) comes from Table 2-10 of the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997). As the 2006 IPCC Guidelines do not cover styrene production under its petrochemicals section, a more recent emission factor could not be found.

### 4.14.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed by Cheminfo Services (2010) for the category of Styrene

Production, following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below  $\pm 30\%$ .

As no plant-specific uncertainty estimates could be collected by Cheminfo Services, a set of default uncertainties (based on Cheminfo Services' knowledge of the industry) was used in the analysis. The Tier 1 uncertainty associated with the category as a whole for the time series was estimated at  $\pm 30\%$  (Cheminfo Services 2010).

#### 4.14.4. Category-Specific QA/QC and Verification

The category of Styrene Production has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 general inventory level QC procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

#### 4.14.5. Category-Specific Recalculations

CH<sub>4</sub> from Styrene Production is a new category for the 2011 submission.

#### 4.14.6. Category-Specific Planned Improvements

There are currently no improvements specifically planned for this category.

### 4.15. Petrochemical Production – Methanol Production (CRF Category 2.B.5.5)

#### 4.15.1. Source Category Description

Methanol (CH<sub>3</sub>OH) is a flammable, highly volatile liquid alcohol at room temperature. It is primarily used as a chemical building block to manufacture formaldehyde, the key precursor of industrial thermoset resins and diverse products. There were three methanol production facilities

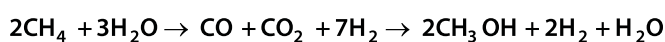
operating in Canada during the 1990-2006 period. One was closed as of 2001, one as of 2005 and the other as of 2006. Methanol was last produced in Canada in 2006.

Methanol is produced by one of these two processes:

1) conventional reforming; and 2) combined (conventional and partial oxidation).

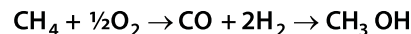
In conventional reforming, methanol is produced in a two-stage process by reacting a synthesis gas (syn gas) containing hydrogen (H<sub>2</sub>), carbon monoxide (CO), and carbon dioxide (CO<sub>2</sub>) over a copper/zinc oxide/alumina catalyst, then separating the product from water and other by-products.

Overall:

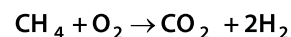


In combined reforming, a catalytic partial oxidation step is added to the conventional reforming process to achieve a better ratio of CO to H<sub>2</sub> in the synthesis gas. The catalytic partial oxidation reaction produces less hydrogen than the conventional reforming reactions step and eliminates by-product hydrogen.

Partial Oxidation Reaction:



Feedstock Oxidation Side Reaction:



Process GHG (CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O) emissions come mainly from process off-gas that is separated from methanol and combusted on-site for energy recovery. The process off-gas contains excess CO, CO<sub>2</sub> and light hydrocarbons. Additional CH<sub>4</sub> emissions can occur in venting of process gases containing CH<sub>4</sub> from the methanol distillation train and methanol storage tanks and fugitive emissions from equipment leaks (Cheminfo Services 2010).

#### 4.15.2. Methodological Issues

To estimate CH<sub>4</sub> emissions from methanol production, a consulting study has been performed by Cheminfo Services. A survey was sent to former employees of the methanol facilities who still have access to the facilities' records, requesting 1990–2009 data on methanol production capacity, production and process GHG emissions. A sector-wide process emission factor was developed as a weight-average using the reported data on CH<sub>4</sub> emissions.

It was calculated by dividing the process CH<sub>4</sub> emission data set by the production data set obtained for the period 2004–2006. The sector-wide process emission factor was used, when necessary, to estimate CH<sub>4</sub> emissions for the 1990–2006 period (there has been no methanol production after 2006).

When CH<sub>4</sub> emission data directly reported by former employees for methanol production facilities were available, they were used in this submission. In the case where there were no reported data, emissions were estimated by multiplying the “unreported” methanol production by the sector-average emission factor. The “unreported” methanol production of a facility was calculated by multiplying its production capacity share (%) by the difference between total national methanol production and the sum of all reported methanol production. National methanol production values were taken from Camford’s CPI Product Profile for 1990–1999 and estimated based on assumed capacity utilization for 2000–2006 (Cheminfo Services 2010).

### 4.15.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed by Cheminfo Services (2010) for the category of Methanol Production, following the 2006 IPCC Guidelines - Propagation of error method for combining uncertainties. The Guidelines state that this method is applicable when the contributing variable uncertainties are below  $\pm 30\%$ .

As no plant-specific uncertainty estimates could be collected by Cheminfo Services (Cheminfo Services 2010), a set of default uncertainties (based on Cheminfo Services’ knowledge of the industry) was used in the analysis:

- national methanol production: 5%;
- reported methanol production: 2%;
- facility methanol capacities: 5%;
- facility fraction of total sector unreported production: 10%;
- reported process CH<sub>4</sub> emissions: 20%;
- reported process N<sub>2</sub>O emissions: 30%.

The uncertainty associated with the category as a whole for the time series ranged from 0% (for the years with no production) to  $\pm 20\%$  for CH<sub>4</sub> emissions and ranged from 0% (for the years with no production) to  $\pm 30\%$  for N<sub>2</sub>O emissions.

### 4.15.4. Category-Specific QA/QC and Verification

This category has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 general inventory level QC procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

### 4.15.5. Category-Specific Recalculations

CH<sub>4</sub> from Methanol Production is a new category for the 2011 submission.

### 4.15.6. Category-Specific Planned Improvements

There are currently no improvements planned for this category

## 4.16. Iron and Steel Production (CRF Category 2.C.1)

### 4.16.1. Source Category Description

Crude (pig) iron is produced in a blast furnace through the reduction of iron oxide (ore), with the carbon in coke or other carbonaceous material as the reducing agent. In most iron furnaces, the process is aided by the use of limestone fluxes (IPCC 2000). Steel can be made in electric arc furnaces (EAFs), basic oxygen furnaces or cupola furnaces. Low-carbon steel is produced in basic oxygen furnaces (BOFs), where a mixture of pig iron and iron scrap is remelted in the presence of pure oxygen, which oxidizes the dissolved carbon to CO or CO<sub>2</sub>. Carbon and alloy steels are produced in EAFs, refractory-lined pots that utilize electric heating through graphite electrodes, which are consumed in the process (IPCC/OECD/IEA 1997).

In the production of pig iron, carbon plays the dual role of fuel and reductant. Emissions from the combustion of fuels such as coke oven gas are not reported in this category, but rather under the appropriate industrial category in the Energy Sector. CO<sub>2</sub> emissions from carbon oxidation, which

occurs when iron ore is reduced to pig iron, are included in this category. Also accounted for in this category are emissions during steel production, which occur to a much lesser extent. These come from the oxidation of carbon in crude iron and electrode consumption. Additional CO<sub>2</sub> given off by limestone flux in the blast furnace is covered under the Limestone and Dolomite Use category (Section 4.4.1).

## 4.16.2. Methodological Issues

A new set of emission parameters—reflecting Canada-specific circumstances, e.g., emission factor for coke, carbon content of pig iron, and carbon content of pig iron entering the steelmaking process—has been used to develop the 1990–2009 estimates. The slightly modified methodology used for the Iron and Steel Production category in this submission follows the IPCC Tier 2, as described in the IPCC Good Practice Guidance (IPCC 2000), with the addition of a term to account for emissions arising from the ladle metallurgy stage—further processing, in electric arc furnaces, of steel output from basic oxygen furnaces. The fate of carbon is tracked throughout the production process, with emissions from iron production and steel production being calculated separately. The following equation was used to estimate emissions from pig iron production:

Equation 4–9:

$$\text{Emissions}_{\text{pig iron}} = (\text{Emission Factor}_{\text{reductant}} \times \text{mass of reductant}) + (\text{mass of carbon in the ore} - \text{mass of carbon in pig iron}) \times \left(\frac{44}{12}\right)$$

where:

Emissions <sub>pig iron</sub>	=	emissions from pig iron production, kt
Emission factor <sub>reductant</sub>	=	year-specific emission factors (t CO <sub>2</sub> / t coke used) obtained from Cheminfo Services (2010) study
mass of reductant	=	mass of metallurgical coke used in the process, kt
mass of carbon in the ore	=	zero; according to IPCC (2000), kt
mass of carbon in pig iron	=	total pig iron production, kt × carbon content in pig iron
44/12	=	ratio of the molecular weight of CO <sub>2</sub> to the molecular weight of carbon

Reducing agents used to produce crude iron from iron ore can be coke, coal, charcoal, heavy fuel oil or petroleum coke. However, for the purposes of this category's emission estimates, it was assumed that the reductant used in the Canadian industry is 100% metallurgical coke. The carbon content in ore is almost zero (IPCC 2000). The GHG emissions associated with the use of reductants other than metallurgical coke are estimated in the Other and Undifferentiated Production source category and under the appropriate industrial category in the Energy Sector.

The data source for the use of metallurgical coke was the *Report on Energy Supply-Demand in Canada* (RES-D – Statistics Canada #57-003). Data on total pig iron production in Canada came from Statistics Canada (for 1990–2003: #41-001; for 2004–2007: #41-019; and for 2008–2009: confidential company-specific [or so-called Micro Data] data). The emission factors for coke use are year-specific; they come from the Cheminfo Services 2010 study (Cheminfo Services 2010). During the study, Cheminfo Services surveyed four integrated steel mills in Canada for their coke consumption and their emission estimates for the years 1990–2009. The emission factors were calculated as ratios of CO<sub>2</sub> emissions to coke consumption. They were then applied to the set of coke use data provided by Statistics Canada. With respect to the carbon content in pig iron, the Canadian Steel Producers Association (CSPA)<sup>9</sup> provided an industry-average content value, which has to be kept confidential.

Emissions from steel production were estimated using the following equation:

Equation 4–10:

$$\begin{aligned} & (\text{Carbon\% in iron charged to steel furnaces} \times \text{charge}) \\ & + (\text{Carbon\% in scrap steel charge to steel furnace} \times \text{scrap steel}) \\ & - (\text{Carbon\% in steel prod BOF} \times \text{steel production in BOF}) \\ & - (\text{Carbon\% in steel prod EAF} \times \text{steel production in EAF}) \times 44/12 \\ & + (\text{Emission Factor for EAF} \times \text{Steel produced from EAF}) \\ & + (\text{Emission Factor for BOF Ladle Metallurgy} \times \text{Steel produced from BOF}) \end{aligned}$$

According to Equation 4–10, part of the amount of CO<sub>2</sub> emitted from the steel production process is estimated based on the difference between the amount of carbon in the iron and in scrap steel used to make steel and the amount of carbon in the steel produced in BOFs and EAFs. It should be noted that the amount of pig iron fed to steel furnaces (used in Equation 4–10) is not equal to

9 Chan K. 2009. Personal communication (email from Chan K. to Maryse Pagé, Greenhouse Gas Division, dated July 21, 2009). Canadian Steel Producers Association.

the amount of total pig iron production (used in Equation 4–9). As part of the steel production process, there are also emissions coming from consumption of electrodes in EAFs and in the secondary ladle metallurgy. These are accounted for in the last two terms of the equation.

Data on the total pig iron charged to steel furnaces, on total steel production, and on the amount of steel produced in EAFs were obtained from Statistics Canada (for 1990–2003: #41-001, for 2004–2007: #41-019 and for 2008–2009: confidential Micro Data). The values of the carbon contents and emission factors mentioned in Equation 4–10: were all provided by the CSPA.<sup>10</sup>

The total emission from the category of Iron and Steel Production is the sum of Equation 4–9 and Equation 4–10 above.

Data on metallurgical coke use at provincial/territorial levels from the RESD (Statistics Canada #57-003) were used to derive the percentage of total reductant consumption attributed to each province and territory. CO<sub>2</sub> emissions at provincial/territorial levels were then estimated by multiplying the percentage derived by the national emission estimate.

It should be noted that RESD data (Statistics Canada #57-003) published for any given year are preliminary and subject to revision in subsequent publications.

The method described above does not account for additional CO<sub>2</sub> given off by the use of limestone as flux in blast furnaces, since the limestone consumption-related emissions are included in the subsector of Limestone and Dolomite Use.

The use of petroleum coke in EAF electrodes is reported by Statistics Canada with all other non-energy uses of petroleum coke. To avoid double counting, the CO<sub>2</sub> emissions from the consumption of electrodes in the steel production process in EAFs are therefore subtracted from the total non-energy emissions. It is assumed that there are no imported electrodes used for steel production in EAFs in Canada. If electrodes are imported, then the portion of CO<sub>2</sub> generated by the imported electrodes will need to be subtracted from the emissions from electrode consumption before being subtracted from the total non-energy emissions.

<sup>10</sup> Chan K. 2009. Personal communication (email from Chan K. to Maryse Pagé, Greenhouse Gas Division dated July 21, 2009). Canadian Steel Producers Association.

### 4.16.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Iron and Steel Production. It took into account the uncertainties associated with all the parameters used in the equations stated above, such as data on metallurgical coke use, emission factor of coke, data on pig iron and steel production and carbon contents of pig iron and steel. The assessment also considered the error associated with the non-response rate of the Statistics Canada surveys. The uncertainty associated with the category as a whole for the time series is around  $\pm 5.4\%$ .

The data sources and methodology used are consistent over the time series.

### 4.16.4. Category-Specific QA/QC and Verification

Iron and Steel Production is a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed are consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

### 4.16.5. Category-Specific Recalculations

As a result of the methodology update explained in Section 4.16.2 the whole time series of 1990–2008 was recalculated. The 1990–2008 estimates have increased by 43–47% (or 3.2–3.8 Mt CO<sub>2</sub> eq).

### 4.16.6. Category-Specific Planned Improvements

Efforts will be made to rectify the assumption made regarding the agents used to reduce iron ore. In the current estimation methodology, it is assumed that the reductant used by industry is 100% metallurgical coke. However, it is known that there are other reductants, such as natural gas, heavy fuel oil and coal, that have been used by the iron and steel industry over the years. Hence, part of the CO<sub>2</sub> emissions coming from non-energy use of fossil fuels, currently reported under the category of Other and Undifferentiated Production and also those reported under the Energy Sector, may be reallocated to the category of Iron



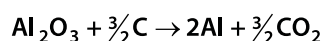
and Steel Production.

## 4.17. Aluminium Production (CRF Category 2.C.3)

### 4.17.1. Source Category Description

Primary aluminium is produced in two steps. In the first stage, bauxite ore is ground, purified and calcined to produce alumina ( $\text{Al}_2\text{O}_3$ ). The latter is then, in the second stage, electrically reduced to aluminium in large pots with carbon-based anodes. The pot itself (a shallow steel container) forms the cathode, whereas the anode consists of one or more carbon blocks suspended within it. Inside the pot, alumina is dissolved in a cryolite ( $\text{Na}_3\text{AlF}_6$ ) bath. Passing a current through the resistance of the cell causes the heating effect, which maintains the contents in a liquid state. The aluminium forms at the cathode and gathers on the bottom of the pot.

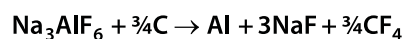
As the anode is consumed,  $\text{CO}_2$  is formed in the following reaction, provided that enough alumina is present at the anode surface:



Although most of the  $\text{CO}_2$  forms from the electrolysis reaction of the carbon anode with alumina as shown above, other sources, namely the baking of prebaked anodes, can contribute to some (usually less than 10%) of the total non-energy-related  $\text{CO}_2$  emissions. Emissions from the combustion of fossil fuels used in the production of baked anodes are covered in the Energy Sector, but emissions arising specifically from the combustion of volatile matter released during the baking operation and from the combustion of baking furnace packing material are accounted for under the Industrial Processes Sector (IPCC 2006).

In addition to  $\text{CO}_2$  emissions, primary aluminium smelting is a major source of carbon tetrafluoride ( $\text{CF}_4$ ) and carbon hexafluoride ( $\text{C}_2\text{F}_6$ ). When alumina levels are too low, these PFCs are formed during an occurrence known as the “anode effect” or “anode event.” In theory, when an anode event occurs, the cell resistance increases very suddenly (within a 50th of a second). As a result, the voltage rises and the temperature goes up, forcing the molten fluorine salts in the cell to chemically combine with the carbon anode (Laval University 1994). During the anode event, the

competing reactions shown below occur to produce  $\text{CF}_4$  and  $\text{C}_2\text{F}_6$ .



PFC emissions can be controlled by computerized alumina feeders. Sensors measure the alumina concentration and automatically feed more to the pot when levels become low. In this way, anode events can be controlled. The computers can be programmed to detect the onset of anode events as well, providing additional warning for the system to take counteractive measures. “Point” feeders, as opposed to “centre-break” types, also tend to reduce emissions (Øye and Huglen 1990).

Besides  $\text{CO}_2$ ,  $\text{CF}_4$  and  $\text{C}_2\text{F}_6$ , a small amount of  $\text{SF}_6$  is also emitted from its use as cover gas at some aluminium plants that produce high magnesium-aluminium alloys (2007 email from P. Chaput).<sup>11</sup>

Aluminium plants are characterized by the type of anode technology employed. In general, older plants with Søderberg technology have higher emissions than newer plants, which usually use pre-baked anodes. The trend in the Canadian aluminium industry has been towards modernizing facilities, since production efficiency has improved. In some cases, this has meant taking old lines out of production as new ones are installed to meet increasing demand.

Finally, even though aluminium production consumes extremely large quantities of electrical energy, currently estimated to be 13.5 kWh/kg of aluminium (AIA 1993), GHG emissions associated with its electricity consumption are not necessarily high in a Canadian context. All of Canada’s primary aluminium smelters are located in Quebec and British Columbia, where almost all (95%) of the electricity generated is produced by hydraulic generators; these are believed to emit a negligible amount of GHGs compared with conventional fossil fuel-based electricity generators.

### 4.17.2. Methodological Issues

Process-related emission estimates for aluminium production were directly obtained from companies via the Aluminum Association of Canada (AAC). In addition to the smelter-specific emission estimates, information on the methodologies used by the aluminium producers to calcu-

<sup>11</sup> Chaput P. 2007. Personal communication (email from Chaput P. to A. Au, Greenhouse Gas Division, dated Oct 12, 2007). Aluminum Association of Canada.



late CO<sub>2</sub>, PFC and SF<sub>6</sub> emissions and plant-specific production data for the time series were obtained from companies via the AAC. The estimation techniques applied may be Tier 3, Tier 2 or Tier 1 type, as described below, depending on data availability; a Tier 3 type technique has mostly been applied for estimating emissions for recent years.

Calculating CO<sub>2</sub> emissions:

Typically, the equations used by smelters to estimate CO<sub>2</sub> emissions from the reaction of the carbon anode with alumina are (AAC 2002b) as follows:

Equation 4–11:

$$\text{CO}_2 \text{ Emissions(t)} = \left[ \frac{\text{CC} \times \text{MP} \times (100 - \%S_a - \%Ash_a - \%Imp_a)}{100} \right] \times \left[ \frac{44}{12} \right]$$

where:

CC	=	baked anode consumption per tonne of aluminium (t C/t Al)
MP	=	total aluminium production (t)
Sa	=	sulphur content in baked anodes (wt%)
Asha	=	ash content in baked anodes (wt%)
Impa	=	fluorine and other impurities (wt%)*
44/12	=	ratio of the molecular weight of CO <sub>2</sub> to the molecular weight of carbon

\*The weight percentage of fluorine and other impurities may not be a parameter considered by all the smelters.

Equation 4–12: for Söderberg anode consumption

$$\text{CO}_2 \text{ Emissions(t)} = \left[ \left( \text{PC} \times \text{MP} \right) - \left( \frac{\text{BSM} \times \text{MP}}{1000} \right) - \left( \frac{\%BC}{100} \times \text{PC} \times \text{MP} \times \frac{(\%S_p + \%Ash_p + \%H_2)}{100} \right) \right] \times \left[ \frac{44}{12} \right] - \left( \frac{(100 - \%BC)}{100} \times \text{PC} \times \text{MP} \times \frac{(\%S_c + \%Ash_c)}{100} \right)$$

where:

PC	=	paste consumption (t paste/t Al)
MP	=	total aluminium production (t)
BSM	=	emissions of benzene-soluble matter (kg/t Al)
BC	=	average binder content in paste (wt%)
Sp	=	sulphur content in pitch (wt%)
Ashp	=	ash content in pitch (wt%)
H <sub>2</sub>	=	hydrogen content in pitch (wt%)
Sc	=	sulphur content in calcinated coke (wt%)
Ash <sub>c</sub>	=	ash content in calcinated coke (wt%)
44/12	=	ratio of the molecular weight of CO <sub>2</sub> to the molecular weight of carbon

The use of the above equations with actual plant-specific data is considered a Tier 3 type methodology. A Tier 2 type technique involves applying some measured data in combination with industry-typical values to these equations. Shown in Table 4–3 are Tier 2 industry-typical values that can be applied by facilities.

Table 4–3 Default Tier 2 Parameter Values for the Estimation of CO<sub>2</sub> Emissions from Anode Consumption

Parameter	Industry-Typical Value	Source
<b>For Prebaked Anode Consumption:</b>		
sulphur content in baked anodes (wt %) - S <sub>a</sub>	2%	IAI 2006
ash content in baked anodes (wt %) - Ash <sub>a</sub>	0.4%	IAI 2006
fluorine and other impurities (wt %) - Imp <sub>a</sub>	0.4%	AAC 2002b
<b>For Söderberg Anode Consumption:</b>		
emissions of benzene-soluble matter (kg/t Al) - BSM	For HSS: 4.0 kg/ t Al For VSS: 0.5 kg/ t Al *	IAI 2006
average binder content in paste (wt %) - BC	Dry Paste: 24% Wet Paste: 27%	IAI 2006
sulphur content in pitch (wt %) - S <sub>p</sub>	0.6 %	IAI 2006
ash content in pitch (wt %) - Ash <sub>p</sub>	0.2%	IAI 2006
hydrogen content in pitch (wt %) - H <sub>2</sub>	3.3%	IAI 2006
sulphur content in calcinated coke (wt %) - S <sub>c</sub>	1.9%	IAI 2006
ash content in calcinated coke (wt %) - Ash <sub>c</sub>	0.2%	IAI 2006

\* HSS = Horizontal Stud Söderberg; VSS = Vertical Stud Söderberg.

When no process data other than aluminium production are available, emission factors (EFs) for a Tier 1 method (as shown below) can be used. These factors depart slightly from the IPCC default factors, because the IPCC Tier 1 default factors reflect 1990 emissions and would produce considerable errors if applied to current production. The factors below reflect the considerable progress that has been made over the period from 1990 to 2001 (AAC 2002b; IAI 2006):

Søderberg: EF = 1.7 t CO<sub>2</sub>/t Al produced; and

Pre-baked: EF = 1.6 t CO<sub>2</sub>/t Al produced.

To calculate CO<sub>2</sub> emissions resulting from anode baking (i.e. pitch volatiles combustion and combustion of baking furnace packing material), the following equations are used (AAC 2002a):

Equation 4–13: for Pitch Volatiles Combustion

$$\text{CO}_2 \text{ Emissions(t)} = (\text{GAW} - \text{BAP} - \text{HW} - \text{RT}) \times \left( \frac{44}{12} \right)$$

where:

GAW	=	green anode weight (t)
BAP	=	baked anode production (t)
HW	=	weight of hydrogen from pitch (t) = %H <sub>2</sub> / 100 * PC / 100 * GAW
H <sub>2</sub>	=	hydrogen content in pitch (wt%)
PC	=	average pitch content in green anode (wt%)
RT	=	waste tar collected (t)
44/12	=	ratio of the molecular weight of CO <sub>2</sub> to the molecular weight of carbon

Equation 4–14: for Packing Coke

$$\text{CO}_2 \text{ Emissions(t)} = \left[ \text{PCC} \times \text{BAP} \times \frac{(100 S_{\text{pc}} \text{ Ash}_{\text{pc}})}{100} \right] \times \left[ \frac{44}{12} \right]$$

where:

PCC	=	packing coke consumed (t coke/ t of baked anode)
BAP	=	baked anode production (t)
S <sub>pc</sub>	=	sulphur content in packing coke (wt%)
Ash <sub>pc</sub>	=	ash content in packing coke (wt%)
44/12	=	ratio of the molecular weight of CO <sub>2</sub> to the molecular weight of carbon

As in the case of anode consumption, the use of Equation 4–13 and Equation 4–14 with actual plant-specific data is considered a Tier 3 type methodology. A Tier 2 type technique involves applying some measured data in combination with industry-typical values to these equations. Shown in Table 4–4 are Tier 2 industry-typical values that can be applied by facilities to estimate CO<sub>2</sub> emissions arising from anode baking.

According to a communication with representatives of the Canadian aluminium industry (2007 emails from P. Chaput and C. Dubois),<sup>12</sup> one of the three aluminium companies has only relied on plant-specific values to develop its process CO<sub>2</sub> estimates (Tier 3), whereas the other two companies have used default parameter values shown in Table 4–3 and Table 4–4 above for estimates of some earlier years of the time series.

12 Chaput P. and Dubois C. 2007. Personal communications (emails from Chaput P. and Dubois C. to A. Au, Greenhouse Gas Division, dated October 12, 2007). Aluminum Association of Canada and Alcoa.

Table 4–4 Default Tier 2 Parameter Values for the Estimation of CO<sub>2</sub> Emissions from Anode Baking

Parameter	Industry-Typical Value	Source
<b>For Pitch Volatiles Combustion:</b>		
hydrogen content in pitch (wt%) - %H <sub>2</sub>	0.5%	IAI 2006
waste tar collected (t) – RT	For Riedhammer furnaces only; all others are insignificant	IAI 2006
<b>For Packing Coke:</b>		
packing coke consumed (t coke/ t of baked anode) – PCC	0.015 t/ t	IAI 2006
sulphur content in packing coke (wt%) - S <sub>pc</sub>	2%	IAI 2006
ash content in packing coke (wt%) - Ash <sub>pc</sub>	2.5%	IAI 2006

It should be noted that the use of petroleum coke in anodes for the production of aluminium was also reported by Statistics Canada with all other non-energy uses of petroleum coke. To avoid double counting, the CO<sub>2</sub> emissions from the consumption of anodes in the aluminium smelting process were therefore subtracted from the total non-energy emissions associated with the consumption of petroleum coke.

#### Calculating PFC Emissions:

CF<sub>4</sub> and C<sub>2</sub>F<sub>6</sub> emitted during anode effects can be calculated by smelters using either the Slope Method or the Pechiney Overvoltage Method, depending on the smelter technology (AAC 2002a):

Equation 4–15: for Slope Method:

$$\text{PFC Emissions (t CO}_2\text{ eq)} = \frac{\text{slope} \times \text{AEF} \times \text{AED} \times \text{MP} \times \text{GWP}}{1000}$$

where:

Slope	=	slope (for CF <sub>4</sub> or C <sub>2</sub> F <sub>6</sub> ) of the emission relationship ([kg PFC/t Al]/[AE-minutes/cell-day])
AEF	=	number of anode effects per pot per day (AE/cell-day)
AED	=	anode effect duration (minutes)
MP	=	total aluminium production (t)
GWP	=	global warming potential for CF <sub>4</sub> or C <sub>2</sub> F <sub>6</sub>

Equation 4–16: for Pechiney Overvoltage Method:

$$\text{PFC Emissions (t CO}_2\text{ eq)} = \frac{\text{overvoltage coefficient} \times \text{AEO} / \text{CE} \times \text{GWP} \times \text{MP}}{1000}$$

where:

overvoltage coefficient	=	([kg PFC/t Al]/[mV/cell-day])
AEO	=	anode effect overvoltage (mV/cell-day)
CE	=	aluminium production process current efficiency expressed as a fraction
GWP	=	global warming potential for CF <sub>4</sub> or C <sub>2</sub> F <sub>6</sub>
MP	=	total aluminium production (t)

The use of the above equations with actual process data to estimate PFC emissions is considered a Tier 3 type methodology. The estimation technique is considered as Tier 2 type when the default coefficients shown in Table 4–5 (IAI 2006) are used together with smelter-specific operating parameters. In a Tier 2 approach, one would first estimate emissions of CF<sub>4</sub> using the slope or overvoltage coefficients as per Equation 4–15 and Equation 4–16. Emissions of C<sub>2</sub>F<sub>6</sub> are then calculated by multiplying the CF<sub>4</sub> estimates by the CF<sub>4</sub>/C<sub>2</sub>F<sub>6</sub> weight fraction.

If only production statistics are available (i.e. no data on anode effect frequency, anode effect duration, or anode effect overvoltage), the Tier 1 emission factors shown in Table 4–6 can be used by smelters (IAI 2006).

Table 4–5 Tier 2 Default Slope and Overvoltage Coefficients (IAI 2006)

Type of Cell	Slope Coefficients for CF <sub>4</sub> ([kg PFC/t Al]/ [AE-minutes/cell-day])	Overvoltage Coefficients for CF <sub>4</sub> ([kg PFC/t Al]/ [mV/cell-day])	Weight Fraction CF <sub>4</sub> / C <sub>2</sub> F <sub>6</sub>
Centre Worked Pre-Baked	0.143	1.16	0.121
Side Worked Pre-Baked	0.272	3.65	0.252
Vertical Stud Søderberg	0.092	NA	0.053
Horizontal Stud Søderberg	0.099	NA	0.085

Note: NA = not applicable

Table 4–6 PFC Emission Factors

Type of Cell	Emission Factors (kg PFC/t Al)	
	CF <sub>4</sub>	C <sub>2</sub> F <sub>6</sub>
Centre Worked Pre-Baked	0.4	0.04
Side Worked Pre-Baked	1.6	0.4
Vertical Stud Søderberg	0.8	0.04
Horizontal Stud Søderberg	0.4	0.03

Based on recent information provided by the Canadian aluminium industry (2007 email from P. Chaput),<sup>13</sup> one of the three aluminium companies has solely relied on plant-specific values to develop its process PFC estimates (Tier 3), whereas the other two companies have used default parameter values shown in Table 4–5 and Table 4–6 above for estimates of some earlier years of the time series.

Calculating SF<sub>6</sub> emissions:

According to the methodology documents supplied by the AAC, SF<sub>6</sub> emissions are equal to consumption in the aluminium industry. This method is consistent with the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997).

### 4.17.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the subsector of Aluminium Production (i.e. for the CO<sub>2</sub>, PFC and SF<sub>6</sub> emission estimates). It took into account the uncertainties associated with all the parameters used in the equations stated above. *The Aluminium Sector Greenhouse Gas Protocol* published by the International Aluminium Institute (IAI) (2006) was the main source of the parameters' uncertainty values. The uncertainties for the CO<sub>2</sub>, PFC and SF<sub>6</sub> estimates were ±7%, ±9% and ±3%, respectively. For the CO<sub>2</sub> and PFC estimates, it should be noted that the uncertainty assessment was done for only one year of the time series (2006 for CO<sub>2</sub> and 2007 for PFC). It is expected that emission estimates of more recent years would have similar uncertainties, while older estimates would have slightly higher uncertainties. For the SF<sub>6</sub> estimate, it was assumed that the uncertainty would be the same as that of the Magnesium Casting category, since the method used to develop SF<sub>6</sub> emission estimates is the same for both Aluminium Production and Magnesium Casting.

The AAC has consistently been used as the data source of estimates shown in this inventory over the time series. The methodology applied by smelters may be of the Tier 3, Tier 2, or Tier 1 type, depending on data availability. However, for recent years, a Tier 3 type technique has been applied by all smelters for estimating emissions.

### 4.17.4. Category-Specific QA/QC and Verification

CO<sub>2</sub> and PFC emissions from Aluminium Production were key categories that have undergone Tier 1 QC checks as elaborated in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

### 4.17.5. Category-Specific Recalculations

There was no recalculation for the category of Aluminium Production.

### 4.17.6. Category-Specific Planned Improvements

Efforts will also be made to acquire more information on the QC procedures followed by member companies during the emission estimation.

## 4.18. Magnesium Metal Production and Casting (CRF Categories 2.C.5.1 & 2.C.4.2)

### 4.18.1. Source Category Description

SF<sub>6</sub> is emitted during magnesium production and casting, where it is used as a cover gas to prevent oxidation of the molten metals. Although emitted in relatively small quantities, SF<sub>6</sub> is an extremely potent GHG, with a 100-year GWP of 23 900. SF<sub>6</sub> is not manufactured in Canada. All SF<sub>6</sub> is imported.

During the 1990–2006 period, there were two major magnesium producers in Canada: Norsk Hydro and Timminco Metals. Norsk Hydro was shut down in the first quarter of 2007. Another magnesium producer, Métallurgie Magnola, existed between 2000 and 2003, but was shut down in April 2003. Between 1990 and 2004, Norsk Hydro had invested in research and development projects having as objectives finding a substitute for SF<sub>6</sub> and eventually elimi-

<sup>13</sup> Chaput P. 2007. Personal communication (email from Chaput P. to A. Au, Greenhouse Gas Division, dated October 12, 2007). Aluminum Association of Canada.

nating the use of SF<sub>6</sub> as cover gas at its plant (2004 email from J. Laperrière).<sup>14</sup> This research, as well as the use of substitute gas mixtures, produced significant reductions in SF<sub>6</sub> emissions in the mid-1990s to late 1990s. For the years 2005–2007, Norsk Hydro's SF<sub>6</sub> emissions were significantly reduced as a result of gradual production reduction and the plant's closure in 2007. Timminco was also closed in August 2008.

There were in total 11 magnesium casting facilities in operation during the 1990–2004 period (Cheminfo Services 2005b). Only a few of them had used SF<sub>6</sub> every year during the entire period. Some casters started using SF<sub>6</sub> towards the mid- or late 1990s, whereas others replaced it with an alternative gas, such as SO<sub>2</sub>. Two facilities have ceased their casting operations over the last few years. During the 2005–2008 period, only seven facilities were in operation and had used SF<sub>6</sub>. Two companies shut down their magnesium casting operations in different times of 2009 (one in June and one in December).

#### 4.18.2. Methodological Issues

For SF<sub>6</sub> emissions from magnesium production, data for 1999–2007 were directly reported by the companies (Norsk Hydro, Timminco Metals and Métallurgie Magnola Inc.) through a mandatory emissions reporting program known as the National Pollutant Release Inventory (NPRI). Emission estimates used in this report were obtained from the NPRI's online database ([http://www.ec.gc.ca/pdb/querysite/query\\_e.cfm](http://www.ec.gc.ca/pdb/querysite/query_e.cfm)). For previous years (i.e. 1990–1998), the data were provided voluntarily by the producers to the GHG Division over the telephone. Since there were no reported 2008 data for Timminco, its 2008 SF<sub>6</sub> value was estimated based on its 2007 data and number of months of operation in 2008 (i.e. 7 months). For 2009 onwards, since there have been no magnesium production plants operating in Canada, there has been no need to perform any data collection.

Representatives from both Norsk Hydro and Timminco were contacted in 2006, so that the methodology they had applied to estimate SF<sub>6</sub> emissions could be understood. Both companies reported that they used the IPCC default method (Emissions of SF<sub>6</sub> = Consumption of SF<sub>6</sub>), as recommended in the IPCC Good Practice Guidance (IPCC 2000). However, they have used different ways for estimating their SF<sub>6</sub> consumption. Norsk Hydro confirmed

the use of the weight difference method (2006 email from J. Laperrière).<sup>15</sup> This method is based on measuring the weight of gas cylinders used at the facility at the time when these were purchased and when these were returned to suppliers at the end of the usage. The accounting method was reported as being used by Timminco for estimating its SF<sub>6</sub> use (2006 email from R. Katan).<sup>16</sup> In this method, accounting of delivered purchases and inventory changes of SF<sub>6</sub> used are recorded. The purchases must be the actual volumes received in the calendar period; therefore, beginning-of-year and end-of-year inventories are taken into account.

The technique applied to estimate emissions from magnesium production is considered to be a Tier 3 type method, as it is based on the reporting of facility-specific emission data.

For calculating SF<sub>6</sub> emissions from casters, the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) provide one general equation that assumes that all SF<sub>6</sub> used as a cover gas is emitted to the atmosphere. To estimate SF<sub>6</sub> emissions for 1990–2004 using this equation, attempts had been made, through a study (Cheminfo Services 2005b) in 2005, to collect data on SF<sub>6</sub> consumption from casting facilities. A couple of facilities indicated that they did not hold any historical records of their past SF<sub>6</sub> consumption. Therefore, to estimate SF<sub>6</sub> use for the entire time series, results of a previous study (Cheminfo Services 2002) were used in combination with the data received from the Cheminfo Services (2005b) study and some assumptions. For casters that had SF<sub>6</sub> data for only a year, it was assumed that their SF<sub>6</sub> use stayed constant, during the other operating years, at the level of the year for which the actual SF<sub>6</sub> data were obtained. For casters that had data for more than one year, linear interpolation between two data points was applied to estimate SF<sub>6</sub> consumption for the other years.

For 2005–2007, consumption data were provided by all seven operating casting facilities in a voluntary data submission process. They were used for the calculation of emissions. For 2008, data were made available by six out of the seven casting facilities in the voluntary data submission process. For the remaining facility, it was assumed that its 2008 SF<sub>6</sub> use stayed at the 2007 level. For 2009, communication was established with all seven companies. Two of

14 Laperrière J. 2004. Personal communication (email from Laperrière J. to A. Au, Greenhouse Gas Division, dated October 27, 2004). Norsk Hydro.

15 Laperrière J. 2006. Personal communication (email from Laperrière J. to A. Au, Greenhouse Gas Division, dated October 4, 2006). Norsk Hydro.

16 Katan R. 2006. Personal communication (emails from Katan R. to A. Au, Greenhouse Gas Division, dated March 16–22, 2006). Timminco.



the companies, for which magnesium casting operations were shut down in 2009, were not able to report their 2009 SF<sub>6</sub> use data, but provided reasonable assumptions that could be used to estimate the 2009 SF<sub>6</sub> use. SF<sub>6</sub> use data for 2009 were provided by the other five facilities.

The technique applied to estimate emissions from magnesium casting for 1990–2004 and 2008–2009 is considered to be a modified Tier 3 type method, as it is based on the reporting of facility-specific emission data and some assumptions. For 2005–2007, the method used is considered as a Tier 3 type.

### 4.18.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of Magnesium Casting. It took into account the uncertainty associated with the SF<sub>6</sub> data reported by each facility. The uncertainty for the category as a whole was estimated at  $\pm 4.0\%$ . It should be noted that the uncertainty assessment was done for only one year of the time series (2007). As such, it is expected that emission estimates of more recent years (2005 onwards) would have a similar uncertainty value, while older estimates would have a slightly higher uncertainty.

As the last magnesium production facility was closed in August 2008, it became difficult to gather the data needed for the Tier 1 uncertainty assessment of the Magnesium Production category. Hence, based on the fact that the same emission estimation method (i.e. emissions = consumption of SF<sub>6</sub>) was applied to both categories of Magnesium Casting and Magnesium Production, it was assumed that the Magnesium Production category would have the same uncertainty ( $\pm 4.0\%$ ) as the Magnesium Casting category.

The data source remains consistent over the time series. The methodology, which equates consumption of SF<sub>6</sub> as a cover gas by magnesium casters to emissions of SF<sub>6</sub>, is applied over the time series with some assumptions for some historical years, as discussed in the methodology section.

### 4.18.4. Category-Specific QA/QC and Verification

Magnesium Production and Magnesium Casting have both undergone Tier 1 QC checks as elaborated in the QA/QC plan (see Annex 6). The checks performed were consistent

with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected from the Tier 1 QC process.

### 4.18.5. Category-Specific Recalculations

There was no recalculation for the category of Magnesium Production. The 2008 estimate for Magnesium Casting has been recalculated, as SF<sub>6</sub> use data from one of the plants have become available and have replaced the previous estimated data. As a result of the activity data update, the 2008 estimate has slightly increased by 0.41% (or 1.1 kt CO<sub>2</sub> eq).

### 4.18.6. Category-Specific Planned Improvements

There is currently no improvement planned specifically for the estimation of SF<sub>6</sub> emissions from magnesium production and casting in Canada.

## 4.19. Production of Halocarbons (CRF Category 2.E)

### 4.19.1. Source Category Description

Chlorodifluoromethane (HCFC-22 or CHClF<sub>2</sub>) is produced when reacting chloroform (CHCl<sub>3</sub>) with hydrogen fluoride (HF), in the presence of antimony pentachloride (SbCl<sub>5</sub>) as a catalyst. During the manufacture of HCFC-22, trifluoromethane (HFC-23 or CHF<sub>3</sub>) is generated as a by-product (IPCC 2002).

The reaction is carried out in a continuous flow reactor, usually under high pressure (up to 500 psig) and temperatures in the order of 45 to 200°C. Although the reaction is exothermic, heat is added to increase the flow of vapours leaving the reactor. The vapour stream contains HCFC-22 (CHClF<sub>2</sub>), HFCF-21 (CHCl<sub>2</sub>F), HFC-23 (CHF<sub>3</sub>), HCl, excess CHCl<sub>3</sub>, HF and some entrained catalyst. Subsequent processing of the vapour stream involves several separations to remove/recover by-products and to purify HCFC-22. Unreacted chloroform, entrained catalyst and underfluori-



nated intermediates (i.e. HCFC-21) from the vapour stream are condensed and returned to the reactor. The major emission point for HFC-23 is the condenser vent, where HFC-23 is discharged into the atmosphere after being separated from HCFC-22 (IPCC 2002).

Two HCFC-22 producers (Dupont Canada and Allied-Signal) operated in Canada during the 1980s and early 1990s. They ceased their HCFC-22 production between 1990 and 1993. According to the data records transferred from the Use Patterns and Control Implementation Section (UPCIS) of Environment Canada to the GHG Division, Dupont Canada produced some HCFC-22 in 1989, but none in the years 1990 and after. Allied-Signal only submitted its 1990–1992 production data to UPCIS because it stopped its operation in 1993 (2007 email from Y. Bovet).<sup>17</sup>

HCFC-22 can be used as refrigerant, as a blend component in foam blowing and as a chemical feedstock for manufacturing synthetic polymers (IPCC 2002). However, due to its ozone-depleting properties, developed countries have scheduled the phase-out of HCFC-22 for the coming years. In Canada, there should be no more manufacturing or import of equipment containing HCFC-22 as of Jan. 1, 2010 (HRAI 2008).

#### 4.19.2. Methodological Issues

To estimate HFC-23 emissions from HCFC-22 production, the total HCFC-22 production was multiplied by the IPCC Tier 1 default emission factor of 0.04 t HFC-23 / t HCFC-22 produced (IPCC/OECD/IEA 1997). It was assumed that destruction (through thermal oxidation) or transformation of HFC-23 was not practiced in Canada. The 1990–1992 production data were collected by the UPCIS from HCFC producers (2007 email from Y. Bovet).<sup>19</sup>

#### 4.19.3. Uncertainties and Time-Series Consistency

Uncertainty in the HFC-23 emission estimates has not been assessed. However, it was believed that the production data reported by HCFC-22 producers were reasonably accurate. The major source of uncertainty could be the Tier 1 default emission factor, because the correlation between the quantity of HFC-23 emitted and the HCFC-22 produc-

tion rate can vary with plant infrastructure and operating conditions (IPCC 2002).

#### 4.19.4. Category-Specific QA/QC and Verification

Informal checks (such as data transcription checks, calculation checks, and unit conversion checks) were done on the category of HCFC-22 Production. No issues of importance were detected.

#### 4.19.5. Category-Specific Recalculations

There were no recalculations for this category.

#### 4.19.6. Category-Specific Planned Improvements

There are currently no improvements planned for this category.

### 4.20. Consumption of Halocarbons (CRF Category 2.F)

Hydrofluorocarbons (HFCs) and, to a very limited extent, perfluorocarbons (PFCs) are serving as alternatives to ozone-depleting substances being phased out under the Montreal Protocol. Application areas of HFCs and PFCs include refrigeration and air conditioning; fire suppression; aerosols; solvent cleaning; foam blowing; and other applications (such as semiconductor manufacturing in the case of PFCs).

Before the ban on the production and use of CFCs came into effect in 1996, as a result of the Montreal Protocol, very few HFCs were produced and used. The only HFCs produced were HFC-152a, a component of the refrigerant blend R-500; and HFC-23, a by-product of HCFC-22 production (discussed in the previous section). As such, emissions from HFC consumption were considered negligible for the 1990–1994 period. HFC-134a began to be produced in 1991 and a variety of other HFCs are now also being produced (IPCC/OECD/IEA 1997). All HFCs consumed in Canada are imported in bulk or in products (e.g. refrigerators). There is no known production of HFCs in Canada.

<sup>17</sup> Bovet Y. 2007. Personal communication (email from Bovet Y. to A. Au, Greenhouse Gas Division, dated November 8, 2007). Environment Canada, UPCIS.

PFCs have been primarily used as cooling/heating agents in specialized market segments and in electronic safety testing. However, emissions from the consumption of PFCs are minor relative to the by-product emissions of PFCs from aluminium production (discussed in the section on aluminium production). Like HFCs, all PFCs consumed in Canada are imported in bulk or in products. There is no known PFC manufacturing in Canada.

### 4.20.1. Methodological Issues

HFC emission estimates for 1995 were based on data gathered from an initial HFC survey conducted by the Use Patterns and Control Implementation Section (UPCIS) of Environment Canada in 1996. Environment Canada has revised subsequent surveys to obtain more detailed activity data. The 1998, 1999, 2001 and 2005 HFC surveys were the source of activity data for emission estimates for the years 1996–2000 and 2004 (2004–2006 emails from Y. Bovet and Y. Guilbault).<sup>18</sup> In some cases, one survey was done to collect data for two years. HFC sales data for 2001–2003 were also collected in 2005 from major HFC importers in Canada (Cheminfo Services 2005c). These data were provided by market segment, such that the total quantity used for each type of application could be determined.

HFC import and sales data for 2005–2009 were collected by the GHG Division through a voluntary data submission process. In this process, requests for data were sent to the main importers of bulk HFCs and to companies that import/export HFC-containing products. For 2009, the distribution list for data collection was expanded, as the Division became aware of other players (either importers of bulk HFCs or importers/exporters of items with HFCs) in the market by looking at HFC import data collected by the Canadian Border Services Agency (CBSA).<sup>19</sup> For the years for which data collection was carried out by the Division, the overall response rate has been above 80% (i.e. over 80% of the companies to which a request was sent were able to provide their data).

In terms of data on imports of bulk HFCs, since the 2007 and 2008 data were not provided by one of the major

importers, it was assumed that the company would have the same 2006–2007 and 2006–2008 growth rates (%) in its imports as the other reporting companies. The same company also did not respond to the voluntary data submission survey for 2009 data. As such, its 2009 HFC import was assumed to stay at the 2008 level. As mentioned above, there were other companies added to the distribution list of the 2009 data collection. The new reporting companies added approximately 3% to the total calculated from the 2009 data reported by the other companies.

In terms of data on import/export of HFC-containing products, in the cases where data were not available from companies, it was assumed that the non-reported quantities stayed at the levels of the most recent years for which data were available. For instance, 1995 and 1999–2003 data on the quantities of HFCs contained in imported and exported products, except imported and exported vehicles, were not available. As such, 1995 HFC quantities in imported and exported products were assumed to be zero. For 1999–2003, these quantities were assumed to stay at 1998 levels and at the 2000 level for imported/exported vehicles, since the 1999 and 2000 amounts of HFCs found in imported and exported vehicles, and the 1998 amounts of HFCs found in other items were provided by the UPCIS. Similar to the situation for data on bulk HFCs, the distribution list for HFC item data collection has been expanded for 2009 data. The new companies added around 9% to the total calculated from the 2009 data reported by the others. Since detailed 1995 HFC data were not available, the IPCC Tier 2 method could not be applied. Instead, a modified Tier 1 methodology was used to obtain a representative estimate of the actual 1995 HFC emissions for the following groups: aerosols; foams; air conditioning original equipment manufacture (AC OEM); AC service; refrigeration; and total flooding systems. To estimate 1996–2008 HFC emissions at the national level, an IPCC Tier 2 methodology was applied. A more detailed description of the Tier 1 and Tier 2 methods used is provided in the following subsections.

Once the emission estimates at the national level were obtained, they were distributed by province/territory based on proxy variables, such as gross output of accommodation and food services for commercial refrigeration and number of households for residential refrigeration. The details of the proxy variables used and assumptions made can be made available upon request.

<sup>18</sup> Bovet Y. and Y. Guilbault. 2004–2006. Personal communications (emails received from Bovet Y. and Guilbault Y. to A. Au, Greenhouse Gas Division, during the years 2004–2006). UPCIS.

<sup>19</sup> It should be noted that HFC data from the CBSA cannot be used for GHG inventory purposes, as these HFC data are collected and categorized only under three types: HFC-134a, HFC-152a and others. Also, the data are not presented by use type. However, company-specific data from the CBSA are a useful tool for expanding the distribution list for the HFC data collection conducted by the GHG Division and for cross-checking the data submitted to the Division.

The IPCC Tier 2 methodology was used to estimate emissions from the consumption of PFCs for the years 1995–2008. Details of the method are found in the following subsections. The 1995–2000 activity data were obtained through the 1998 and 2001 PFC surveys conducted by Environment Canada. As 2001–2004 data were unavailable, emission estimates were developed, by and large, based on the assumption that the use quantities in various applications stayed constant since 2000. The GHG Division conducted a collection of 2005–2007 PFC use data from major consumers of PFCs. Over 50% of the companies to which a request was sent reported their data to the Division. The collected data were used in the emission estimation. For the companies for which data were not available, it was assumed that PFC use stayed constant at the 2000 levels. Moreover, as companies that have been closed or that stopped using PFCs were identified throughout the process, the corresponding activity data used to develop estimates were adjusted accordingly. For 2008–2009, data on bulk PFC were collected from major importers. However, because Consumption of PFCs is not a key category (not a significant GHG source), data on (imported/exported) items containing PFCs were not collected and it was assumed that they stayed at the levels of the most recent years for which data were available.

#### 4.20.1.1. 1995 HFC Emission Estimates

The following subsections provide explanations on the emission factors used and the assumptions made to develop 1995 HFC emission estimates for AC OEM, AC service, refrigeration, foam blowing, aerosol products and total flooding systems.

##### **Air Conditioning Original Equipment Manufacture (AC OEM)**

To estimate emissions from AC OEM, the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) suggest a 2–5% loss rate. For Canada, a rate of 4% was assumed.

##### **AC Service**

Since it was assumed that for AC systems, HFCs were mostly used to replace operating losses, a loss rate of 100% was applied.

##### **Refrigeration**

It was assumed that all refrigeration in Canada falls under

the “commercial and industrial” category, since it was the dominant emission source. It was also assumed that the quantity of HFCs reported under “refrigeration” represented the amount used for initial and subsequent recharging of equipment. Therefore,

Equation 4–17:

$$\text{HFC (refrig)} = \text{Charge} + \text{Operating Loss}$$

The IPCC considers that operating loss is approximately 0.17(charge) (IPCC/OECD/IEA 1997). Therefore, assuming the total charge remains constant for the short term,

$$\text{HFC (refrig)} = \text{Charge} + 0.17(\text{Charge}) = 1.17(\text{Charge})$$

or

$$\text{Charge} = \frac{\text{HFC (refrig)}}{1.17}$$

Assuming assembly leakage was minimal,

$$\text{Emissions} = \text{Operating Loss} = 0.17(\text{Charge})$$

Thus,

Equation 4–18:

$$\text{Emissions} = 0.17 \times \left( \frac{\text{HFC (refrig)}}{1.17} \right)$$

##### **Foam Blowing**

For 1995, it was assumed that all foam blowing was of open cell type. In other words, an emission factor of 100% was applied.

##### **Aerosol Products**

For aerosol products, the IPCC Good Practice Guidance (IPCC 2000) suggests a default EF of 50% of the initial charge per year. It was assumed that 1994 production was 50% of that of 1995. Hence, emissions from aerosol products manufactured in 1994, occurring in 1995, would be equivalent to approximately 25% of the 1995 consumption level. Therefore, a factor of 80% was applied to the 1995 consumption to estimate HFC emissions from aerosol products in 1995.

##### **Fire Suppression – Total Flooding Systems**

For 1995, it was assumed that all fire suppression equipment to which HFCs were introduced was of the total flooding type. Hence, a factor of 35% (IPCC/OECD/IEA 1997) was applied to estimate HFC emissions from fire suppression.

#### 4.20.1.2. 1996–2009 HFC and 1995–2009 PFC Emission Estimates

The following subsections provide explanations on the emission factors used and the assumptions made to develop 1996–2009 HFC and 1995–2009 PFC emission estimates.

##### Refrigeration and AC System Assembly

The Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) recommend the use of the following equation for the estimation of emissions from the assembly of residential refrigeration, commercial refrigeration, stationary AC, and mobile AC systems:

Equation 4–19:

$$E_{\text{assembly},t} = \text{Charge}_t \times k$$

where:

$E_{\text{assembly},t}$	=	emissions during system manufacture and assembly in year t
$\text{Charge}_t$	=	quantity of refrigerant charged into new systems in year t
k	=	assembly losses in percentage of the quantity charged

The k value was chosen from a range of values that were provided for each equipment category in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) (see Table 4–7).

##### Annual Leakage

The equation below, given in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997), was used to calculate HFC and PFC emissions from leakage:

Table 4–7 Percentage of Losses during Assembly (k) for Various Applications

For HFC estimates		For PFC estimates	
Application Type	k Values (%)	Application Type	k Values (%)
Residential Refrigeration	2.0	Refrigeration (including ultra low temperature refrigeration)	3.5
Commercial Refrigeration	3.5	Stationary AC	3.5
Stationary AC	3.5	Mobile	4.5
Mobile AC	4.5		

Table 4–8 Annual Leakage Rates (x) for Various Applications

For HFC estimates:		For PFC estimates:	
Application Type	x Values (%)	Application Type	x Values (%)
Residential Refrigeration	1.0	Refrigeration (including ultra low temperature refrigeration)	17.0
Commercial Refrigeration	17.0	Stationary AC	17.0
Stationary AC	17.0	Mobile AC	30.0
Mobile AC	15.0		

Equation 4–20:

$$E_{\text{operation}, t} = \text{Stock}_t \times x$$

where:

$E_{\text{operation}, t}$	=	quantity of HFCs/PFCs emitted during system operations in year t
$\text{Stock}_t$	=	quantity of HFCs/PFCs stocked in existing systems in year t
$x$	=	annual leakage rate in percentage of total HFC/PFC charge in the stock

The term “Stock<sub>t</sub>” includes the amount of HFCs/PFCs contained in equipment manufactured in Canada, the amount of HFCs/PFCs in imported equipment, and the amount of HFCs used for servicing equipment and excludes the amount of HFCs/PFCs in exported equipment. It was assumed that no leakage occurred in the year of manufacturing. The Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) give a range of values for the annual leakage rate (x) for each of the different equipment categories. The annual leakage rate chosen for each category is shown in Table 4–8.

## System Disposal

It was assumed that there were no HFC/PFC emissions from the disposal of refrigeration and stationary AC systems during 1995–2009, since these systems have a lifetime of 15 years (IPCC default value) and HFC use began only in 1995. For the disposal of mobile AC systems with a slightly shorter lifetime of 12 years (the IPCC default average value), it was assumed that there were no recovery and recycling technologies in place and, therefore, 100% of the quantities remaining in systems built in 1995 would be emitted in 2008.

## Foam Blowing

The IPCC Tier 2 methodology presented in the revised 1996 IPCC guidelines (IPCC/OECD/IEA 1997) was used to estimate HFC (1996–2009) and PFC (1995–2009) emissions from foam blowing. Foams are grouped into two main categories: open cell and closed cell.

### Open Cell Foam Blowing

In the production of open cell foam, 100% of the HFCs used are emitted (IPCC/OECD/IEA 1997). There has been no known PFC use in open cell foam blowing. Open cell foam production categories that release HFC emissions include the following:

- Cushioning—Automobiles;
- Cushioning—Others;
- Packaging—Food;
- Packaging—Others; and
- Other Foam Uses.

### Closed Cell Foam Blowing

During the production of closed cell foam, approximately 10% of the HFCs/PFCs used are emitted (IPCC/OECD/IEA 1997). The remaining quantity of HFCs/PFCs is trapped in the foam and is emitted slowly over a period of approximately 20 years. The IPCC Tier 2 equation (as shown below) was used to calculate emissions from closed cell

Equation 4–21:

$$E_{\text{foam}, t} = 10\% \times \text{Qty}_{\text{manufacturing}, t} + 4.5\% \times \text{Orig. Charge}$$

where:

$E_{\text{foam}, t}$	=	emissions from closed cell foam in year t
$\text{Qty}_{\text{manufacturing}, t}$	=	quantity of HFCs/PFCs used in manufacturing closed cell foam in year t
Orig. Charge	=	original charge blown into the foam

The following are closed cell foam production categories that emit HFC emissions:

- Thermal Insulation—Home and Building;
- Thermal Insulation—Pipe;
- Thermal Insulation—Refrigerator and Freezer; and
- Thermal Insulation—Other.

## Fire Extinguishers

The IPCC Tier 2 methodology of the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) was used to calculate HFC emissions from portable fire extinguishers and total flooding systems from 1996 onward. There has been no known PFC use in fire-extinguishing equipment.

### Portable Fire-Extinguishing Equipment

The IPCC Tier 2 methodology in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) estimated emissions as 60% of HFCs used in newly installed equipment.

### Total Flooding Systems

The IPCC Tier 2 methodology provided in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) estimated emis-



sions from total flooding systems as 35% of the HFCs used in newly installed fire-extinguishing systems.

### Aerosols/Metered Dose Inhalers

The IPCC Tier 2 methodology presented in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) was used to calculate HFC emissions from aerosols for 1996 onwards. The emission estimate for the current year is equal to half of the HFCs used in aerosols in the current year plus half of the HFCs used in aerosols in the previous year. The amount of HFCs used each year is equal to the amount of HFCs used to produce aerosols and the amount of HFCs in imported aerosol products and excludes the amount of HFCs in exported aerosol products.

Since no data on PFCs used in aerosols were gathered from Environment Canada's PFC surveys, it was assumed that PFC emissions coming from its use in aerosols were negligible.

### Solvents

The IPCC Tier 2 methodology presented in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) was used to estimate HFC and PFC emissions from solvents. The emission estimate for the current year is equal to half of the HFCs/PFCs used as solvents in the current year plus half of the HFCs/PFCs used as solvents in the previous year. The amount of HFCs/PFCs used each year is equal to the amount of HFCs/PFCs produced and imported as solvents and excludes the amount of HFCs/PFCs exported as solvents. HFCs/PFCs used as solvents include the following categories:

- electronics industries;
- laboratory solvents; and
- general cleaning.

### Semiconductor Manufacture

There are two main uses of PFCs in the semiconductor manufacturing industry: plasma etching of silicon wafers and plasma cleaning of chemical vapour deposition chambers.

IPCC Tier 2b methodology, as shown below, was used to estimate PFC emissions from the semiconductor manufacturing industry:

Equation 4-22:

$$E_{SC} = E_{FC} + E_{CF4}$$

where:

$E_{SC}$	=	total PFC emissions from semiconductor
$E_{FC}$	=	emissions resulting from the use of PFCs (see Equation 4-23: below)
$E_{CF4}$	=	CF <sub>4</sub> emitted as a by-product during the use of PFCs (see Equation 4-24 below)

Equation 4-23:

$$E_{FC} = (1-h) \times \sum_p [FC_{i,p} \times (1-C_{i,p}) \times (1-a_{i,p} \times d_{i,p})]$$

where:

$h$	=	fraction of fluorocarbon remaining in shipping container (heel) after use
$p$	=	process type (plasma etching or chemical vapour deposition chamber cleaning)
$FC_{i,p}$	=	quantity of fluorocarbon $i$ fed into the process type $p$
$C_{i,p}$	=	use rate (fraction destroyed or transformed) for each fluorocarbon $i$ and process type $p$
$a_{i,p}$	=	fraction of gas volume $i$ fed into the process $p$ with emission control technologies
$d_{i,p}$	=	fraction of fluorocarbon $i$ destroyed in the process $p$ by the emission control technologies

Equation 4-24:

$$E_{CF4} = (1-h) \times \sum_p [B_{i,p} \times FC_{i,p} \times (1-a_{i,p} \times d_{i,p})]$$

where:

$B_{i,p}$	=	fraction of gas $i$ transformed into CF <sub>4</sub> for each process type $p$
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and other terms are as defined above.

Default values for variables used in the above equations are shown in Table 4-9 (IPCC 2000).

As no information on emission control technologies for these processes was available,  $a_{i,p}$  was assumed to equal to 0 and  $d_{i,p}$  to 1. Also,  $h$  was assumed to equal 0.1, as suggested in IPCC (2000).



Table 4–9 PFC Emission Rates<sup>1</sup>

Process	IPCC Default Emission Fractions			
	CF <sub>4</sub>	C <sub>2</sub> F <sub>6</sub>	C <sub>3</sub> F <sub>8</sub>	c-C <sub>4</sub> F <sub>8</sub>
(1-C) Plasma Etching	0.7	0.4	0.4	0.3
(1-C) Chemical Vapour Deposition Chamber	0.8	0.7	0.4	ND
B Plasma Etching	NA	0.1	ND	NA
B Chemical Vapour Deposition Chamber	NA	0.1	0.2	NA

Notes:

1. Tier 2b, from IPCC (2000)

NA = not applicable.

ND = no data.

## Other Sources

Minor amounts of PFC emissions have been identified as related to PFC use in the electronics industry for emissive applications, including reliability testing (inert liquids), coolants (direct evaporative cooling for electric and electronic apparatuses and indirect coolants in closed-circuit electronic apparatuses), and precision cleaning (IPCC 2000). These emissions can be classified into two types of sources: *emissive* and *contained*.

*Emissive* sources include the following:

- electrical environmental testing;
- gross leak testing; and
- thermal shock testing.

Unidentified and miscellaneous PFC uses reported in the PFC survey were also considered as part of *emissive* sources. According to the IPCC Tier 2 methodology, 50% of PFCs used for the above purposes would be released during the first year and the remaining 50% released in the following year.

*Contained* sources consist of PFC use as an electronic insulator and a dielectric coolant for heat transfer in the electronics industry. The IPCC Tier 2 emission factors (IPCC 2000) are applied to the PFC use data obtained from the PFC survey to estimate PFC emissions from contained sources, as follows:

Equation 4–25:

$$E_{\text{contained},t} = (k \times Qty_t) + (x \times Stock_t) + (d \times Qty_t)$$

where:

$E_{\text{contained},t}$  = emissions from contained sources

$$E_{\text{contained},t} = (k \times Qty_t) + (x \times Stock_t) + (d \times Qty_t)$$

$Qty_t$  = quantity of PFC sale for use or manufacturing of contained sources in year  $t$

$Stock_t$  = quantity of PFCs in stock in year  $t$

$k$  = manufacturing emission rate (1% of annual sales)

$x$  = leakage rate (2% of stock)

$d$  = disposal emission factor (5% of annual sales)

## 4.20.2. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the category of HFC Consumption. It took into account the uncertainties associated with all the subcategories, such as residential/commercial refrigeration, stationary/mobile AC, etc. To determine the uncertainty for a subcategory, the uncertainties related to activity data (Cheminfo 2005c) and emission factors (Japan's Ministry of the Environment 2009) were used. It should be noted that the category uncertainty can vary throughout the time series because it is dependent on the magnitude of each of the subcategory emission estimates, which changes from year to year. The uncertainty associated with the category as a whole for the time series ranged from  $\pm 34\%$  to  $\pm 50\%$ .

A Tier 1 uncertainty assessment has also been performed for the category of PFC Consumption. Similar to HFC Consumption, the uncertainties related to activity data (IPCC 2006) and emission factors (Japan's Ministry of the Environment 2009) were taken into account in the assessment for PFC Consumption. The uncertainty associated with the category as a whole for the time series ranged from  $\pm 10\%$  to  $\pm 24\%$ .

### 4.20.3. Category-Specific QA/QC and Verification

Consumption of Halocarbons resulting in HFC emissions was a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). Some transcription errors (for 1999–2008 estimates) were detected from the Tier 1 QC process. Specifically, the 1999–2008 values have been recalculated, as a result of the use of linearly interpolated activity data (as recommended by the Expert Review Team, ERT, in a past inventory review report) in place of the assumed constant values for the years for which data were not available. However, due to a file linkage problem, the revised estimates have not been correctly transcribed to the national summary tables shown in Annex 12. To maintain consistency between the NIR and the CRF tables, the same “unrevised” 1999–2008 data are inputted to the CRF Reporter.

Informal checks (such as data transcription checks, calculation checks, and unit conversion checks) were done on the category of PFC Consumption. No issues of importance were detected.

### 4.20.4. Category-Specific Recalculations

Although not shown in the national summary tables or in the CRF tables, as mentioned in the previous section, the 1999–2008 values have been recalculated as a result of the use of linearly interpolated activity data in place of the assumed constant values for the years for which data were not available.

There was no recalculation for the category of PFC Consumption.

### 4.20.5. Category-Specific Planned Improvements

Transcription errors for the years 1999–2008 that were detected during the quality control process will be corrected in the next submission, so that the implementation of a previous recommendation from the ERT can be fully reflected.

Furthermore, some information on emission factors related to the use of HFCs has been obtained throughout the preparation of this inventory, but a follow-up with companies will be needed to acquire more details, such that well-founded country-specific HFC emission factors can be developed. The inventory agency is also looking into the development of a regulatory collection tool, in future, for gathering of HFC and PFC use data. This, it is hoped, will improve the response rates and decrease the uncertainty in emissions. Finally, efforts will be made to acquire more information on the recovery and recycling technologies in place for the disposal of equipment using HFCs.

## 4.21. Production and Consumption of SF<sub>6</sub> (CRF Categories 2.E & 2.F)

### 4.21.1. Source Category Description

In addition to magnesium production and casting, electrical equipment in electric utilities and semiconductor manufacturing are known sources of SF<sub>6</sub> emissions. In electric utilities, SF<sub>6</sub> is used as an insulating and arc-quenching medium in high-tension electrical equipment, such as electrical switchgear, stand-alone circuit breakers and gas-insulated substations.

There is currently no production of SF<sub>6</sub> in Canada; therefore, all Canadian supply of SF<sub>6</sub> is obtained through imports. From 1990 to 1996, more than 95% of total SF<sub>6</sub> imports came from the United States; however, in recent years, this percentage has declined, with an increase in SF<sub>6</sub> imports from Germany (Cheminfo Services 2002).

### 4.21.2. Methodological Issues

#### SF<sub>6</sub> Emissions from Electrical Equipment

As an improvement to the 2011 submission, a modified Tier 3 method was used to estimate SF<sub>6</sub> emissions from electrical equipment in utilities for certain years (i.e. 2006–2009) of the time series, in place of the previous top-down approach (which assumed that all SF<sub>6</sub> purchased from gas distributors replaces SF<sub>6</sub> lost through leakage). The SF<sub>6</sub> emission estimates by province for 2006–2009 were provided by the Canadian Electricity Association (CEA)—which represents electricity companies across

Canada, except for the province of Quebec—and Hydro Quebec (HQ). The emission data submitted by the CEA and HQ were prepared following the *SF<sub>6</sub> Emission Estimation and Reporting Protocol for Electric Utilities* (Environment Canada and Canadian Electricity Association 2008). The national SF<sub>6</sub> estimate for each year of 2006–2009 was the sum of all provincial estimates. The *Protocol* is the result of a collaborative effort between the GHG Division, the CEA and HQ. It is also part of the memorandum of understanding between the GHG Division and the CEA, and of the voluntary data submission agreement of HQ.

In summary, the *Protocol* explains how the (country-specific) modified Tier 3 method was derived from the IPCC Tier 3 life cycle methodology. It also explains the different options available for estimating the equipment life cycle emissions. These are equal to the sum of SF<sub>6</sub> used to top up the equipment and the equipment disposal and failure emissions (which are equal to nameplate capacity less recovered quantity for disposal emissions or to simply nameplate capacity for failure emissions). The options for the tracking of SF<sub>6</sub> consumed for top-ups are mass flow meters, weigh scales and cylinder count. For further details on the methodology, data uncertainty, data quality control, data verification by third party, transfer of information and data to the GHG Division, documentation and archiving, new information or data updates, and protocol reviews and amendments, please refer to the *Protocol* (available upon request at <http://www.ec.gc.ca/Publications/default.asp?lang=En&xml=5926D759-36A6-467C-AE05-077C5E6C12A2>) A more detailed description of the methodology is also provided in Annex 3.2

Estimates were not available from the CEA and HQ for the years 1990–2005 because a systematic manner to inventory the quantities of SF<sub>6</sub> used was only started in 2006. Hence, the application of the *Protocol* was not possible. Section 7.3.2.2 of the 2000 IPCC Good Practice Guidance suggests four approaches for such a situation: 1) overlap; 2) surrogate method; 3) interpolation; and 4) trend extrapolation. Each of these techniques has been evaluated for its applicability in this particular situation. The surrogate and trend extrapolation methods could not be used because according to the 2000 IPCC Good Practice Guidance, it is NOT good practice to use these approaches for a long period. The interpolation approach could not be used because data for intermittent years were needed. The overlap approach was determined to be the most appropriate option in this case for the following reasons:

- The overlap between two or more sets of annual emission estimates could be assessed. (In this case, the overlap was assessed between four sets of annual estimates – 2006, 2007, 2008 and 2009.)
- There was a consistent and proportional relationship between the estimates developed using the “old” (i.e. used in the previous submission) and the modified Tier 3 methods. (In this case, the national estimates developed using the modified Tier 3 method were consistently around 10–20% of those developed using the old method.)

Emissions at provincial/territorial levels were estimated based on the national emission estimates (obtained from the use of the overlap approach) and the percent of provincial shares (based on the reported 2006–2009 data)

## SF<sub>6</sub> Emissions from Semiconductor Manufacturing

The method applied to estimate SF<sub>6</sub> emissions from semiconductor manufacturing was similar to the one used for calculating PFC emissions. However, as there is no by-product CF<sub>4</sub> created during the use of SF<sub>6</sub> in the process, Equation 4–26 is not needed.

Equation 4–26:

$$\text{SF}_6 \text{ Emissions} = (1 - h) \times [\text{FC} \times (1 - C) \times (1 - a \times d)]$$

where:

h	=	fraction of SF <sub>6</sub> remaining in shipping container (heel) after use (%)
FC	=	quantity of SF <sub>6</sub> fed into the process (or sales) (t)
c	=	use rate (fraction destroyed or transformed) (%)
a	=	fraction of gas volume fed into the process with emission control technologies
d	=	fraction of SF <sub>6</sub> destroyed in the process by the emission control technologies

The value of “h” provided and confirmed by two major SF<sub>6</sub> gas distributors, Air Liquide and Praxair, was 12% (2006 emails from H. Rahal and A. Tardif).<sup>20</sup> The IPCC default value of 0.5 for “(1 – C)” was used. As it was assumed that there has been no emission control technology applied by this industry, the values of “a” and “d” were 0 and 1, respectively. The estimation technique is considered as Tier 2 type. It should be noted that efforts have been made

20 Rahal H. and Tardif A. 2006. Personal communications (emails from Rahal H. and Tardif A. to A. Au, Greenhouse Gas Division, dated November 22, 2006, and November 13, 2006, respectively). Praxair and Air Liquide, respectively.

during the preparation of this inventory to obtain country-specific (1 - C), a and d values. However, due to the scarcity of the collected data, the development of country-specific parameter values has not been possible.

Since only 1995–2000 sales data were obtained from major Canadian gas suppliers through a study conducted in 2005 (Cheminfo Services 2005a), it was assumed that the quantity sold per year during 1990–1994 was at the 1995 level. The sales per year for 2001–2003 were assumed to be the average value between 1995 and 2000. The SF<sub>6</sub> sales to semiconductor manufacturers in 2004–2009 were estimated by multiplying the SF<sub>6</sub> import (data purchased from Statistics Canada) by the sales distribution (in %) to semiconductor manufacturers. As explained above, sales data by market segment provided by some of the major SF<sub>6</sub> gas distributors were used to establish the percentage of sales attributed to each user/market type. It is noteworthy to mention that attempts have been made to collect SF<sub>6</sub> use data directly from manufacturers, but the response rate for the data-gathering exercise was rather low and the small amount of collected data would not bring in any improvement to the current estimation method.

### 4.21.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment was conducted for the category of Electrical Equipment. It should be noted, though, that the uncertainty assessment was done using 2007 data of the previous submission. Therefore, it is expected that emission estimates of this submission would have much lower uncertainty values. The uncertainty for the category as a whole was estimated at  $\pm 32.0\%$ .

A Tier 1 uncertainty assessment has also been performed for the category of Semiconductor Manufacturing. Although the assessment took into account uncertainties associated with the parameters of Equation 4–24, it should be mentioned that the most of these uncertainties were assumed values due to data unavailability. The uncertainty for the category as a whole was estimated at  $\pm 45\%$ . The uncertainty value is applicable to all years of the time series.

Depending on the years, the data source and methodology used for SF<sub>6</sub> from electrical equipment could vary, as explained in the section above. For SF<sub>6</sub> from semiconductor manufacturing, both data source and methodology used were consistent throughout the time series.

### 4.21.4. Category-Specific QA/QC and Verification

SF<sub>6</sub> Consumption in Electrical Equipment was a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected.

Informal checks (such as data transcription checks, calculation checks and unit conversion checks) were done on the category of Semiconductor Manufacturing. No issues of importance were detected.

### 4.21.5. Category-Specific Recalculations

With the implementation of the *SF<sub>6</sub> Emission Estimation and Reporting Protocol for Electric Utilities* (Environment Canada and Canadian Electricity Association 2008) and the use of the overlap approach, the whole time series (1990–2008) for SF<sub>6</sub> from electrical equipment has been recalculated. The 1990–2008 estimates have been revised by -72% to -89% (or -0.58 to -1.7 Mt CO<sub>2</sub> eq)

There was no recalculation for the category of SF<sub>6</sub> from Semiconductor Manufacturing.

### 4.21.6. Category-Specific Planned Improvements

As a planned improvement for the category of SF<sub>6</sub> from Electrical Equipment, an assessment of the uncertainty associated with the new emission data will be performed in collaboration with the Canadian electric utility companies.

## 4.22. Other and Undifferentiated Production (CRF Category 2.G)

### 4.22.1. Source Category Description

Emissions from this subsector are from the non-energy use of fossil fuels and are not accounted for under any of

the other subsectors of the Industrial Processes Sector. Examples of fuels in non-energy applications are the use of natural gas liquids (NGLs) and feedstocks in the chemical industry and the use of lubricants. All of them result in varying degrees of oxidation of the fuel, producing CO<sub>2</sub> emissions. Also included in this category are emissions from use of hydrocarbons (such as coal) as reductants for base metal smelting.

The use of fossil fuels as feedstock or for other non-energy uses is reported in an aggregated manner by Statistics Canada (#57-003) under “Non-Energy Use” for each individual fuel. In the event that CO<sub>2</sub> emissions resulting from non-energy fuel use are allocated to another category of the Industrial Processes Sector (as is the case for ammonia production, iron and steel production, and aluminium production), those emissions are subtracted from the total non-energy emissions to avoid double counting.

## 4.22.2. Methodological Issues

Emission rates for non-energy use of fuels were developed based on the total potential CO<sub>2</sub> emission rates and the IPCC default percentages of carbon stored in products. The total potential CO<sub>2</sub> emission rates were derived from the carbon emission factors shown in Jaques (1992), McCann (2000) and CIEEDAC (2006).

Fuel quantity data for non-energy fuel usage were reported by the *Report on Energy Supply-Demand in Canada* (RES-D – Statistics Canada #57-003). It should be noted that the RES-D data for any given year are preliminary and subject to revisions in subsequent publications. These data were multiplied by the emission rates shown in Annex 3 to estimate CO<sub>2</sub> emissions for this subsector.

This technique is considered to be a Tier 1 type method, as it is based on the use of national consumption data and average national emission factors. Methodological issues for calculating CO<sub>2</sub> emissions from the non-energy use of fossil fuels are not addressed specifically in the IPCC Good Practice Guidance (IPCC 2000).

Further details with respect to the calculation method used are provided in Annex 3.

## 4.22.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has also been performed

for the subsector of Other and Undifferentiated Production. The assessment took into account uncertainties associated with the activity data and emission factors (ICF Consulting 2004). The uncertainty for the category as a whole was estimated at  $\pm 21\%$ . It should be noted that the uncertainty assessment was done for only one year of the time series (2007).

The data sources and methodology used are consistent over the time series

## 4.22.4. Category-Specific QA/QC and Verification

Other and Undifferentiated Production was a key category that has undergone Tier 1 QC checks as developed in the QA/QC plan (see Annex 6). The checks performed were consistent with the Tier 1 General Inventory Level QC Procedures outlined in the IPCC Good Practice Guidance (IPCC 2000). No issues of importance were detected.

## 4.22.5. Category-Specific Recalculations

Unrounded activity data for 1990–2007 have been used in this submission, resulting in slightly revised emission estimates for these years. A comparison between previously submitted estimates and those in this submission shows that the differences are minute (i.e. less than 1%).

The 2008 estimate has also been slightly revised by 1.3% (or 0.14 Mt CO<sub>2</sub> eq) because updated activity data were obtained from Statistics Canada and were used in the estimation.

## 4.22.6. Category-Specific Planned Improvements

CH<sub>4</sub> and N<sub>2</sub>O emissions from Petrochemical Production are now reported under the Chemical Industry. As such, it is planned to also allocate CO<sub>2</sub> emissions from petroleum products used as feedstock in the petrochemical industry into the category of Petrochemical Production. This would mean a reallocation of emissions from the Other and Undifferentiated Production subsector to the Chemical Industry subsector.



# Chapter 5

## Solvent and Other Product Use (CRF Sector 3)

### 5.1. Overview

Although the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997) mention that solvents and related compounds can be significant sources of emissions of non-methane volatile organic compounds (NMVOCs), the Solvent and Other Product Use Sector accounts only for direct greenhouse gas (GHG) emissions. Annex 10 of the NIR provides details on emissions of NMVOCs and other indirect GHG emissions.

This Sector specifically includes emissions that are related to the use of N<sub>2</sub>O as an anaesthetic and propellant. Emissions from use of solvents in dry cleaning, printing, metal degreasing and a variety of industrial applications as well as household use are not estimated because, according to the Revised 1996 IPCC Guidelines, GHGs are not emitted in significant amounts from these types of uses (IPCC/OECD/IEA 1997). Process CO<sub>2</sub> emissions associated with the production of solvents are included in the Industrial Processes Sector.

As shown in Table 5–1, the GHG emissions from the Solvent and Other Product Use Sector contributed 0.26 Mt CO<sub>2</sub> eq to the 2009 national GHG inventory, compared

with 0.18 Mt CO<sub>2</sub> eq in 1990. These emissions represented 0.038% of the total Canadian GHG emissions in 2009. The emission trends, either long term or short term, were driven mainly by the domestic demand for N<sub>2</sub>O for anaesthetic or propellant purposes.

Most of the N<sub>2</sub>O sold in Canada is used as an anaesthetic. The second major application of N<sub>2</sub>O is as a propellant to generate pressure in aerosol products, with the largest application being pressure-packaged whipped cream. Demand for N<sub>2</sub>O in Canada for manufacturing this food product has been relatively stable since 1995 (Cheminfo Services 2006).

To ensure the correctness of the estimates, this Sector has undergone Tier 1 quality control checks. Further details on quality assurance and quality control and uncertainty assessment can be found in sections 5.1.4 and 5.1.3, respectively.

#### 5.1.1. Source Category Description

N<sub>2</sub>O is a clear, colourless, oxidizing liquefied gas with a slightly sweet odour, and is stable and inert at room temperature. Steam (H<sub>2</sub>O) and N<sub>2</sub>O are formed in a low-pressure and low-temperature reaction that decomposes ammonium nitrate (NH<sub>4</sub>NO<sub>3</sub>). While the steam is condensed out, the “crude” N<sub>2</sub>O is further purified, compressed, dried and liquefied for storage and distribution. Nitrous Oxide of Canada in Maitland, Ontario, is the only known producer of compressed N<sub>2</sub>O for commercial sales in Canada. It supplies N<sub>2</sub>O to two of the three primary N<sub>2</sub>O gas distributors that essentially account for the total commercial market in Canada. These companies sell cylinders of N<sub>2</sub>O to a relatively large number of sub-distributors. It is estimated that there may be 9000 to 12 000 final end-use customers for N<sub>2</sub>O in Canada, including dental offices, clinics, hospitals and laboratories (Cheminfo Services 2006).

Table 5–1 Solvent and Other Product Use Sector GHG Emission Summary, Selected Years

GHG Source Category	GHG Emissions (kt CO <sub>2</sub> eq)							
	1990	2000	2004	2005	2006	2007	2008	2009
Solvent and Other Product Use TOTAL	179	247	215	183	329	326	342	260
Use of N <sub>2</sub> O as an anaesthetic	151	211	182	155	278	276	289	220
Use of N <sub>2</sub> O as a propellant	27	36	33	28	51	51	53	40

Note: The sums of anaesthetic and propellant use may not add up to the solvent total due to rounding.



N<sub>2</sub>O is used in a limited number of applications, with an aesthetic use representing the vast majority of consumption in Canada. Use as a propellant in food products is the second largest type of end use in Canada. Other areas where N<sub>2</sub>O can be used include production of sodium azide<sup>1</sup> (a chemical that was used to inflate automobile airbags), atomic absorption spectrometry and semiconductor manufacturing. According to the distributors that were surveyed during the recent study, approximately 82% of their N<sub>2</sub>O sales volume is used in dentistry/medical applications, 15% in food processing propellants and only 3% for the other uses (Cheminfo Services 2006).

It is important to note that, of all applications in which N<sub>2</sub>O can be used, only the two major types are emissive. When N<sub>2</sub>O is used as an anaesthetic, it is assumed that none of the N<sub>2</sub>O is metabolized (IPCC 2006). In other words, the used N<sub>2</sub>O quickly leaves the body in exhaled breath (i.e. is emitted) as a result of the poor solubility of N<sub>2</sub>O in blood and tissues. When N<sub>2</sub>O is used as a propellant, only emissions coming from N<sub>2</sub>O used in whipped cream are estimated, because the amounts of N<sub>2</sub>O employed in other food products and in non-food products are considered negligible, according to the food industry and the gas producer and distributors. When the cream escapes from the can, the N<sub>2</sub>O gas expands and whips the cream into foam. As none of the N<sub>2</sub>O is reacted during the process, it is all emitted to the atmosphere (Cheminfo Services 2006).

### 5.1.2. Methodological Issues

Estimation of N<sub>2</sub>O emissions from this Sector was done based on sales data, following the consumption-based approach presented in the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997). Because it was virtually impossible to collect data from all end users, it was assumed that domestic sales equal domestic consumption.

Attempts were made to collect sales data, instead of purchase or consumption data, for all years. To obtain the sales volumes by end-use type, for 1990–2005, Canada's single N<sub>2</sub>O producer and the three major N<sub>2</sub>O gas distributors were surveyed through a study (Cheminfo Services 2006). Nitrous Oxide of Canada (NOC) was contacted to obtain its annual production and domestic sales data, but it was able to provide only rough estimates of historical data. Also, a questionnaire requesting sales volumes by

market segment was sent to each of the N<sub>2</sub>O distributors. However, these companies did not provide the full set of historical data that was requested. For the 2006–2009 data, requests for data were sent to NOC and the gas distributors directly by the GHG Division.

As a complete set of sales data covering 1990–2009 could not be gathered, the domestic sales of Canadian production data provided by NOC and N<sub>2</sub>O import data purchased from Statistics Canada<sup>2</sup> were used to estimate the total domestic sales volumes (or consumption) of N<sub>2</sub>O. The sales data by market segment and qualitative information gathered from the producer and distributors were used to develop the patterns of sales by application (Cheminfo Services 2006). The sales pattern for 2006–2009 was assumed to be the same as the one for 2005, which was determined during the study in 2006 (Cheminfo Services 2006). To calculate the amounts of N<sub>2</sub>O sold for anaesthetic and propellant purposes, the total domestic sales volume was multiplied by the percentage of each of these provided in the sales patterns.

To estimate emissions coming from usage of anaesthetic at the national level, it was assumed that 100% of the quantity used in anaesthetic applications was emitted, as explained previously in the source category description section. This is the recommended emission rate as per the 2006 IPCC guidelines.

To estimate emissions coming from N<sub>2</sub>O use in food products (i.e. whipped cream cans) at the national level, it was assumed that 100% of the quantity used in the whipped cream manufacturing was emitted, as explained previously in the source category description section. The same assumption was also made by the U.S. EPA.

Summing the emission estimates for anaesthetic and propellant would give the national emission totals for the Solvent and Other Product Use Sector.

The national emission estimates were divided by the national total population to yield an emissions per capita factor. This factor was then multiplied by the population in each province and territory to estimate emissions at provincial/territorial levels. The 1990–2000 annual population statistics were obtained from Statistics Canada (#91-213) and 2001–2009 data from Statistics Canada (#91-215).

<sup>1</sup> N<sub>2</sub>O was used by ICI Chemicals between 1990 and 1997 as a reactant for producing sodium azide. However, a different raw material has been used in its place since 1998.

<sup>2</sup> 1990–2005 data were downloaded from [http://www.statcan.ca/trade/scripts/trade\\_search.cgi](http://www.statcan.ca/trade/scripts/trade_search.cgi), but this URL is no longer functional. For 2006–2009 data: Bourgeois P. 2010. Personal communications (email from Bourgeois to E. Renault, Greenhouse Gas Division, on December 2, 2010). Statistics Canada.

The IPCC *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000) does not provide any recommendations on the estimation of N<sub>2</sub>O emissions.

### 5.1.3. Uncertainties and Time-Series Consistency

A Tier 1 uncertainty assessment has been performed for the Solvent Sector. It took into account the uncertainties associated with domestic sales, import, sales patterns and emission factors. The uncertainty associated with the Sector as a whole was evaluated at  $\pm 19\%$ . It should be noted, though, that the uncertainty assessment was done for only one year of the time series (2007). However, it was expected that the uncertainty for this Sector would not vary considerably from year to year, especially for more recent years, as the data sources and methodology applied were the same.

### 5.1.4. Quality Assurance, Quality Control and Verification

This Sector was a key category which has undergone Tier 1 quality control checks as developed in the quality assurance/quality control plan (see Annex 6). The checks performed were consistent with the Tier 1 general inventory level quality control procedures outlined in the IPCC good practice guidance (IPCC 2000). No issues of importance were detected from the Tier 1 quality control process.

### 5.1.5. Recalculations

Recalculations were necessary for the entire 1990–2008 time series, as the emission rate for anaesthetic use of N<sub>2</sub>O was changed from 97.5% to 100%, as recommended by the 2006 IPCC guidelines (IPCC 2006). This raised values for the entire time series by 2.5%, and 3 kt to 7 kt of CO<sub>2</sub> eq over that time period.

### 5.1.6. Planned Improvements

Updated historical N<sub>2</sub>O production data will be used in the 2012 submission to improve the N<sub>2</sub>O emission estimates. Also, the use patterns of N<sub>2</sub>O will be reviewed based on data and inputs to be supplied by gas distributors.

# Chapter 6

## Agriculture (CRF Sector 4)

### 6.1. Overview

Emission sources from the Agriculture Sector include animal production—namely, enteric fermentation ( $\text{CH}_4$ ), manure management ( $\text{N}_2\text{O}$ ,  $\text{CH}_4$ ) and crop production including agricultural soils ( $\text{N}_2\text{O}$ ) and field burning of crop residues ( $\text{CH}_4$  and  $\text{N}_2\text{O}$ ). Carbon dioxide emissions from and removals by agricultural lands are reported under the Land Use, Land-Use Change and Forestry (LULUCF) Sector under the Cropland category (see Chapter 7).

The largest sectors in Canadian agriculture are beef cattle (non dairy), swine, and cereal and oil seed crops. There is also a large poultry industry and a large dairy industry, but they are controlled to supply national market demand. Sheep are raised, but production is highly localized and small compared to the beef, swine, dairy and poultry industries. Other animals are produced for commercial purposes, namely buffalo,<sup>1</sup> llamas, alpacas, horses and goats, but production is small.

Canadian agriculture is highly regionalized due to historic and climatic influences. Approximately 70% of beef cattle and more than 90% of wheat, barley and canola are produced on the Prairies in a semi-arid to sub-humid ecozone. On the other hand, approximately 70% of dairy cattle, 60% of swine and poultry, 95% of corn and 90% of soybeans are produced on the humid mixed-wood plains ecozone in Eastern Canada.

In 1990 there were 10.5 million non-dairy cattle in Canada, 1.4 million dairy cattle, 10.2 million swine and 100 million poultry. Beef cattle and swine populations peaked in 2005

at 14.8 and 15 million head but have decreased to 13 and 12 million head, respectively, while poultry populations had increased to 133 million head. Dairy cattle populations have decreased steadily from 1.4 million head in 1990 to 1 million head in 2009.

Since 1990, cropping practices have changed in Canada, with increasing canola production from 3 Mt to 12 Mt, corn production from 7 Mt to 16 Mt, and soybean production from 1.3 Mt to 5 Mt, and decreasing wheat production from 32 Mt to 27 Mt. Synthetic nitrogen consumption has increased from 1.2 Mt N in 1990 to 1.9 Mt N in 2009, while the area under summerfallow has decreased by 4.5 megahectares (Mha) and the regions using conservation tillage have increased by 12.4 Mha.

As a result of those changes, total greenhouse gas (GHG) emissions from the Canadian Agriculture Sector have increased from 47 Mt  $\text{CO}_2$  eq in 1990, to 56 Mt  $\text{CO}_2$  eq in 2009 (Table 6–1). This represents an increase of 19% between 1990 and 2009, mainly due to higher populations of beef cattle and swine (23% and 19%, respectively), as well as an increase in the consumption of synthetic nitrogen fertilizers (59%).

For the period of 2004 to 2008, there were no significant changes in total emissions, as increases in emissions due to increased use of nitrogen fertilizer and crop residue decomposition were offset by reductions in emissions from animal production with the decrease in animal populations (Enteric Fermentation and Manure Management, Table 6–1). In 2009, emissions from soils decreased slightly, reversing the trend of the past three years. As a result of this decrease, in combination with the continued decline in emissions from livestock due to decreasing animal populations, an overall decrease of 2 Mt  $\text{CO}_2$  eq for total agricultural emissions was observed in 2009, relative to the previous years.

Recalculations occurred in this submission due to corrections made in the livestock database that subsequently had impacts across the entire Agriculture Sector of the inventory. These changes resulted in a decrease in the estimate of emissions by 4.1 Mt  $\text{CO}_2$  eq or 6.5% for 2008. Furthermore, a combination of corrections to milk data for the period of 1990 to 1999 and reductions in emission estimates and manure excretion estimates from the two cattle subcategories Slaughter Heifers and Steers, modified the long-term emission trend. The 14 Mt  $\text{CO}_2$  eq (29%) increase in agricultural emissions from 1990 to 2008 reported in the

<sup>1</sup> The Intergovernmental Panel on Climate Change (IPCC) category “buffalo” is being used in Canada to represent the North American bison (*Bison bison*) that is raised for meat production using methods similar to beef cattle.

Table 6–1 Short- and Long-Term Changes in GHG Emissions from the Agriculture Sector

GHG Source Category		GHG Emissions (kt CO <sub>2</sub> eq)							
		1990	2000	2004	2005	2006	2007	2008	2009
Agriculture TOTAL		47 000	55 000	58 000	58 000	57 000	57 000	58 000	56 000
Enteric Fermentation		16 000	20 000	21 000	22 000	21 000	21 000	20 000	19 000
—CH <sub>4</sub>	Dairy Cattle	3 300	2 800	2 700	2 700	2 700	2 600	2 600	2 600
	Beef Cattle <sup>2</sup>	12 000	16 000	18 000	18 000	18 000	17 000	17 000	16 000
	Others <sup>3</sup>	610	910	1 100	1 100	1 100	1 000	1 000	1 000
Manure Management		5 700	6 900	7 300	7 400	7 300	7 100	980	950
—CH <sub>4</sub>	Dairy Cattle	700	610	590	580	580	560	560	560
	Beef Cattle <sup>2</sup>	670	760	830	840	810	790	770	740
	Swine	1 100	1 400	1 600	1 600	1 600	1 500	1 300	1 300
	Poultry	72	90	92	92	92	92	92	92
	Others <sup>4</sup>	23	35	38	38	38	38	38	38
—N <sub>2</sub> O	All Animal Types	3 100	4 000	4 200	4 300	4 200	4 100	4 000	3 900
Agricultural Soils		25 000	29 000	29 000	28 000	29 000	30 000	31 000	30 000
Direct Sources		14 000	15 000	15 000	15 000	15 000	16 000	17 000	16 000
	Synthetic Nitrogen Fertilizers	5 900	7 700	7 600	7 000	7 100	8 100	8 500	8 600
	Manure Applied as Fertilizers	1 800	2 000	2 100	2 200	2 100	2 100	2 000	1 900
	Crop Residue Decomposition	4 700	4 700	5 000	5 200	5 400	5 200	6 200	5 600
	Cultivation of Organic Soils	63	63	63	63	63	63	63	63
	Conservation Tillage <sup>5</sup>	-300	-740	-850	-890	-850	-940	-1 200	-1 100
	Summerfallow	1 400	1 100	870	830	720	700	760	660
	Irrigation	280	340	350	350	340	350	390	360
Pasture, Range, and Paddock Manure		2 200	3 100	3 300	3 400	3 300	3 200	3 200	3 000
Indirect Sources		8 700	10 000	11 000	10 000	10 000	11 000	11 000	11 000
Crop Residue Burning (CH <sub>4</sub> & N <sub>2</sub> O)		210	120	32	41	44	35	45	45

Notes:

1. Totals may not add up due to rounding.

2. Beef Cattle includes dairy heifers.

3. Others Enteric Fermentation, includes buffalo, goat, horse, lamb, llama/alpaca, sheep and swine.

4. Others, Manure Management, includes buffalo, goat, horse, lamb, llama/alpaca and sheep.

5. The negative values reflect a reduced N<sub>2</sub>O emission due to the adoption of conservation tillage.

2010 inventory report has been reduced to an increase of 11.5 Mt CO<sub>2</sub> eq (25%). Some of the recalculations carried out on the livestock database were integrated into the 2010 inventory submission only for the year 2008, (Table 6–2), resulting in time-series inconsistencies that were reported in the 2010 inventory report. These corrections and others have now been completed for the entire time series.

Biological nitrogen fixation by the legume-rhizobium association is reported as not occurring. This decision is supported by Rochette and Janzen (2005) (and reflected in the IPCC 2006 Guidelines), who concluded that there is no evidence that measurable amounts of N<sub>2</sub>O are produced during the nitrogen fixation process. Methane emissions from rice production in Canada do not occur. For the first time, Canada has reported emissions of CH<sub>4</sub> and N<sub>2</sub>O from field burning of agricultural residues. Prescribed burning of savannas is not practised in Canada. Finally, emissions of GHG from on-farm fuel combustion are included in the Energy Sector (Chapter 3).

For each emission source category, a brief introduction and a brief description of methodological issues, uncertainties and time-series consistency, quality assurance / quality control (QA/QC) and verification, recalculations, and planned improvements are provided in this chapter. The detailed inventory methodologies and sources of activity data are described in Annex 3.3.

## 6.2. Enteric Fermentation (CRF Category 4.A)

### 6.2.1. Source Category Description

In Canada, animal production varies from region to region. In western Canada, beef production dominates, combining both intensive production systems with high animal densities finished in feedlots, and low-density, ranch-style, pasturing systems. Most dairy production occurs in

Table 6–2 Proposals for Improvement by the ERT for Canada's 2009 Submission and Actions Taken in the 2010 Submission.

Correction or Improvement	2010 submission	2011 submission
1. Correction to Quebec milk data input. Fat data for this province had been input as a percentage as opposed to a quantity per day.	No correction	Corrected for time series, when errors occurred, 1990 – 1999.
2. Correction to equations used to calculate the net energy mobilized during weight loss ( $Ne_{mob}$ ), having an impact on the calculations of gross energy intake (GE).	Corrected for year 2008, but not for the 1990 – 2007 time series.	Corrected for entire time series 1990 – 2009.
3. Correction in database structure used to weigh subannual production stages to develop an annual emission factor	No correction	Corrected for entire time series 1990 – 2009.
4. Corrections to cattle weight data. Inconsistencies were observed between the database and published reports used in the development of the methodology.	No correction	Corrected for entire time series 1990 – 2009.
5. Implementation issue with slaughter heifers and steers. Emissions were corrected for the duration of time for cattle in feedlots to be finished. Correction reduced time of emissions from enteric fermentation and total quantity of manure produced by this subcategory of animals.	No correction	Corrected for entire time series 1990 – 2009.
6. Update to swine population data to eliminate inconsistencies between activity data sources.	Updated for year 2008, but not for the 1990 – 2007 time series.	Updated for entire time series 1990 – 2009.
7. Integration of regionally distributed manure management systems from Marinier et al. (2005) supplied estimates of the distribution of manure.	No change	Integrated for entire time series, 1990 – 2009.
8. Use of regional, weighted non dairy cattle emission factor for buffalo, as opposed to steer. The use of this emission factor takes into account all stages of the buffalo production, which is similar to beef production in Canada.	No correction	Corrected for entire time series 1990 – 2009.
9. "Integration of methodology for the calculation of emissions related to the burning of agricultural residues."	Not integrated on original submission, integrated for resubmission for entire time series, 1990 – 2008	Integrated for entire time series, 1990 – 2009.

eastern Canada in high-production, high-density facilities. Eastern Canada also has traditionally produced swine in high-density, intensive production facilities. Over the past 20 years, some swine production has shifted to western Canada. Other animals that produce  $CH_4$  by enteric fermentation are raised, such as buffalo, goats, horses, llamas/alpacas and sheep; however, populations of these animals have traditionally been low.

Methane ( $CH_4$ ) is produced during the normal digestive process of enteric fermentation by herbivores. Microorganisms break down carbohydrates and proteins into simple molecules for absorption through the gastro-intestinal tract and  $CH_4$  is produced as a by-product. This process results in an accumulation of  $CH_4$  in the rumen that is emitted

by eructation and exhalation. Some  $CH_4$  is released later in the digestive process by flatulation, but this accounts for less than 5% of total emissions. Ruminant animals, such as cattle, generate the most  $CH_4$ .

## 6.2.2. Methodological Issues

The diversity of animal production systems and regional differences in production facilities complicate emission estimation. For each animal category/subcategory,  $CH_4$  emissions are calculated, by province, by multiplying the animal population of a given category/subcategory by its corresponding emission factor.

For cattle,  $CH_4$  emission factors are estimated using the



Intergovernmental Panel on Climate Change (IPCC) Tier 2 methodology, based on the equations provided by IPCC Good Practice Guidance. A national study by Boadi et al. (2004) broke down cattle subcategories, by province, into subannual production stages and defined their physiological status, diet, age class, sex, weight, growth rate, activity level and production environment. These data were integrated into IPCC Tier 2 equations to produce annual emission factors for each individual animal subcategory that take into account provincial production practices. The data describing each production stage were obtained by surveying beef and dairy cattle specialists across the country.

Increased milk production and fat content with herd improvements in dairy cattle over the 1990–2009 time period are reflected in a gradual increase in CH<sub>4</sub> emission factors from this animal category. Increased milk production results in a greater requirement for energy of lactation (NE<sub>l</sub>) and greater food consumption. In beef cattle, herd improvements have led to increased mature body weight and increased consumption of feed, due to increased energy requirements for maintenance and growth (NE<sub>m</sub> and NE<sub>g</sub>) for the larger body size. As a result, emission factors have increased from 1990 to 2009 for this animal category (see Table A3-7, Annex 3.3).

For non-cattle animal categories, CH<sub>4</sub> emissions from enteric fermentation continue to be estimated using the IPCC Tier 1 methodology. Poultry are excluded from enteric fermentation estimates, since no emission factors are available.

Activity data consist of domestic animal populations for each animal category/subcategory, by province, and are obtained from Statistics Canada (Annex 3.3, Table A3-1). The data are based on the *Census of Agriculture*, conducted every 5 years and updated annually by semi-annual or quarterly surveys for important animal categories.

### 6.2.3. Uncertainties and Time-Series Consistency

The uncertainties associated with CH<sub>4</sub> emissions from enteric fermentation were determined using the Monte Carlo technique, based on the IPCC Tier 2 methodology for the years 2004 and 2008 (Hutchinson et al. 2007). Hutchinson et al. (2007) assumed, based on expert opinion, that uncertainties associated with animal populations ranged from  $\pm 1\%$  for poultry,  $\pm 2\%$  for sheep and lambs,  $\pm 3\%$  for

dairy cattle,  $\pm 5\%$  for non-dairy cattle,  $\pm 10\%$  for swine, and  $\pm 15\%$  for horses, goats and llamas. Emission factor uncertainty was  $\pm 5\%$  for dairy and varied from  $\pm 8\%$  to  $\pm 17\%$  for beef cattle, depending on the animal subcategory. Uncertainties for enteric fermentation emission estimates were  $\pm 5\%$  for dairy cows,  $\pm 9\%$  for beef cattle,  $\pm 13\%$  for swine,  $\pm 16\%$  for sheep and lambs, and  $\pm 24\%$  for other animals. For all animals other than cattle, uncertainty was based on emission factors taken from the IPCC Tier 1 defaults for non-cattle categories estimated to have  $\pm 20\%$  uncertainty (IPCC/OECD/IEA 1997). Total uncertainty on the enteric fermentation value was estimated at  $\pm 11\%$ .

The methodology and parameter data used in the calculation of emission factors are consistent throughout the entire time series (1990–2009) with the exception of milk production for dairy. Milk production from 1990 to 1999 in Ontario and the western provinces is estimated based on production data from eastern Canada.

### 6.2.4. QA/QC and Verification

Enteric Fermentation, as a key category, has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC (2000) Good Practice Guidance. The activity data, methodologies and changes are documented and archived in both paper and electronic forms. The IPCC Tier 2 emission factors for cattle, derived from Boadi et al. (2004), have been reviewed by independent experts (McAllister and Basarab 2004).

The Enteric Fermentation source category is a part of a Tier 2 QA/QC for the Agriculture Sector (MacDonald et al. 2011). This analysis included i) rebuilding and analyzing the database model, by retracing all documented changes to the model and comparing calculation methodology; ii) a complete review of method rationale, implementation and activity data selection; iii) a review and compilation of Canadian research related to methane production from enteric fermentation; and iv) an assessment of the methodology in terms of its strengths and weaknesses.

The analysis of the database model and review of method rationale and implementation resulted in a number of recalculations, outlined in Section 6.2.5. The review and compilation of Canadian research on enteric fermentation suggested that emission factors do agree, in general, with the emission factors from the IPCC Tier 2 model used by Canada in the inventory. No specific bias can be clearly identified in the enteric emission estimate. Improvements



to the cattle model require the development of country-specific Tier 3 calculation methodology and cannot be achieved through improvement to specific Tier 2 parameters. Details of this review can be found in Annex 3.3, Section 3.3.3.2.

### 6.2.5. Recalculations

Significant recalculations resulted from the Tier 2 analysis of the enteric fermentation database. Some of the issues with the database were identified in the 1990 to 2008 inventory submission but were not introduced in the complete time series until this submission (Table 6–2). Corrections to the database shifted estimates either up or down, depending on the source of error, the animal category, and the year (Table 6–3). In the case of dairy cattle, there was an overall increase of 0.58 Mt CO<sub>2</sub> eq or 21% in 1990, that was largely associated with errors in milk data in the province of Quebec. In 2007, the increase was smaller, 0.18 Mt CO<sub>2</sub> eq or 7.4%, and was largely attributed to the correction to gross energy. There was only a 1.1% correction in 2008, as most of the corrections had been completed for the database in the 2010 submission.

The Tier 2 analysis also resulted in large recalculations to the Non-dairy Cattle category. Most of the same corrections observed for the Dairy Cattle category were observed for the Non-dairy Cattle category, with the exception of

the correction to milk data that was not applicable to the Non-dairy category. The most significant recalculation for non-dairy cattle was due to a method implementation issue in the calculation of emissions from slaughter heifers and steers that did not affect the Dairy Cattle category. The Boadi et al. (2004) cattle production model calculates an emission factor in kg/head/yr, based on daily gross energy intake according to equation 4.14 of the *IPCC Good Practice Guidance* (IPCC 2000); however the actual annual emission factor for an animal subcategory (for example, Slaughter Heifers and Steers) is the sum of the emission factors for each of the sub-annual production stages weighted by their time in that production stage. During the implementation of this method in the national inventory, the emission factor for all slaughter heifer and steer production stages were weighted to one full year. According to the Boadi et al. production stage model, these animal subcategories consist of animals fed over a short period and sent to slaughter (short-keeps) and animals that are back-grounded over a period of time and then sent to slaughter (long-keeps). The average lifespan of short-keep animals is 3 to 4 months, while long-keeps have an average lifespan of 9 to 10 months. These animals do not emit methane for a full year. This error explains in part the expert review team's (ERT) concerns over the relatively high emission factor reported by Canada for non-dairy cattle that had been expressed after the 2010 submission.

Table 6–3 Recalculations for Enteric Fermentation Estimates from Cattle – Absolute and Percent Change<sup>1</sup>

Recalculation	1990		2007		2008	
	Mt CO <sub>2</sub> eq	%	Mt CO <sub>2</sub> eq	%	Mt CO <sub>2</sub> eq	%
<b>Dairy Cattle</b>	0.584	21.3	0.179	7.41	-0.028	-1.09
Database errors – milk data – Quebec	0.379	13.9	NO	NO	NO	NO
Database errors – Net energy weight loss	0.207	7.58	0.182	7.53	-0.027	-1.03
Database errors – cattle weights <sup>2</sup>	-0.001	-0.042	-0.001	-0.042	-0.001	-0.046
Database errors – Animal classification and weighting <sup>2</sup>	-0.001	-0.042	-0.001	-0.042	-0.001	-0.046
Activity data – Population data changes <sup>2</sup>	NO	NO	-0.001	-0.042	0.001	0.031
<b>Beef Cattle</b>	-1.22	-8.98	-2.04	-10.7	-2.20	-11.6
Database errors – cattle weights	-0.028	-0.209	-0.047	-0.243	-0.044	-0.234
Database errors – Animal classification and weighting	-0.216	-1.59	-0.355	-1.85	-0.336	-1.78
Database errors – Net energy weight loss	0.193	1.42	0.317	1.65	NO	NO
Activity data – Population data changes	NO	NO	-0.037	-0.195	-0.001	-0.001
Implementation of slaughter heifers & steers	-1.17	-8.61	-1.92	-10.0	-1.82	-9.64
<b>Total Cattle</b>	-0.636	-3.90	-1.86	-8.64	-2.23	-10.4

Notes:

1. Totals may not add up due to rounding.

2. Small recalculations were estimated by difference and averaged.

NO-not occurring

The impact of this correction and other database corrections resulted in an overall decrease in non-dairy cattle emission estimates for all years. However, the decrease was less in 1990 compared to 2007 and 2008 (1.2 Mt CO<sub>2</sub> eq or 9.0% in 1990, 2.0 Mt CO<sub>2</sub> eq or 11% in 2007 and 2.2 Mt CO<sub>2</sub> eq or 12% in 2008) as populations of finishing animals were smaller relative to total beef populations in 1990 than in 2007 and 2008. Therefore, the relative impact of the correction to heifer and steer emissions was smaller in 1990. This difference, combined with the attenuating effect of the large recalculation in dairy cattle for 1990, resulted in a change in the trend in enteric fermentation emissions. With recalculations, emissions increased by only 24% between 1990 and 2008, compared to 33% as was indicated in the 2010 inventory submission.

Other small recalculations included corrections to swine populations for the years 2005 to 2007. Inconsistencies were observed between electronic tables and published reports available from Statistics Canada. These recalculations resulted in changes in emission estimates from an increase of 0.5% or 0.002 Mt CO<sub>2</sub> eq in 2005 to a decrease of 1.9% or 0.009 Mt CO<sub>2</sub> eq in 2007 compared to the emission estimates submitted in 2010.

### 6.2.6. Planned Improvements

Analysis of the enteric fermentation model for non-dairy cattle and the identification of the implementation issue with slaughter heifers and steers suggested that a review of the slaughter heifer and steer portion of the Boadi et al. (2004) cattle model is required. Consultations have begun with experts in the beef industry, and a review of slaughter cattle numbers will be carried out in the coming year. If population numbers and feedlot duration require revision, a process will be put in place to revise the slaughter animal portion of the inventory model that will be implemented over the short to medium term.

Work has been done on an updated Tier 2 uncertainty analysis for enteric fermentation emissions. Karimi-Zandashty et al. (2011) are completing an analysis of the 2008 inventory year, using new data from Statistics Canada and a revised model structure from Hutchinson et al. (2007). This study highlights the importance of assumptions of independence and correlation between parameters used in the model—in particular, IPCC default parameters. Further work is required to finalize uncertainty analyses, as distinct differences were noted between the calculation methodology and the model structure used by Karimi-Zandashty et

al. (2011) for the uncertainty estimates and the current NIR database.

In general, the Tier 2 QC found the enteric fermentation methodology to be robust; improvements are mainly dependant on the ability to collect more complete data on animal nutrition across the different regions of Canada. Consultations are ongoing with national animal production experts to develop a time series that accounts for changes in feed ration digestibility for certain animal categories to be incorporated over the medium term.

## 6.3. Manure Management (CRF Category 4.B)

In Canada, the animal waste management systems (AWMS) typically used in animal production include 1) liquid, 2) solid storage and drylot, and 3) pasture and paddock, but include some other systems such as composters and biodigestors. No manure is burned as fuel.

Both CH<sub>4</sub> and N<sub>2</sub>O are emitted during handling and storage of livestock manure. The magnitude of emissions depends upon the quantity of manure handled, its characteristics, and the type of manure management system. Generally, poorly aerated manure management systems generate high CH<sub>4</sub> emissions but relatively low N<sub>2</sub>O emissions, whereas well-aerated systems generate high N<sub>2</sub>O emissions but relatively low CH<sub>4</sub> emissions.

Manure management practices vary regionally and also by animal category. Dairy, poultry and swine production occur in modern high-density production facilities. Dairy and swine produce large volumes of liquid manure while poultry production produces solid manure, both of which are spread on a limited landbase. Feedlot beef production results in large volumes of drylot and solid manure, whereas low-density pasturing systems for beef result in widely dispersed manure in pastures and paddocks. Production systems for other animals, such as buffalo, goats, horses, llamas/alpacas and sheep are generally in medium-density production facilities producing mainly solid manure.

### 6.3.1. CH<sub>4</sub> Emissions from Manure Management (CRF Category 4.B (a))

#### 6.3.1.1. Source Category Description

Shortly after manure is excreted, the decomposition process begins. In well-aerated conditions, decomposition is an oxidation process producing CO<sub>2</sub>; however, if little oxygen is present, carbon is reduced, resulting in the production of CH<sub>4</sub>. The quantity of CH<sub>4</sub> produced depends on manure characteristics and on the type of manure management systems. Manure characteristics are in turn linked to animal category and diets.

#### 6.3.1.2. Methodological Issues

Methane emissions from manure management are calculated for each animal category/subcategory by multiplying its population by the corresponding emission factor (see Annex 3.3 for detailed methodology). The animal population data are the same as those used for the enteric fermentation emission estimates (Section 6.2.2). Manure management systems for each animal category were taken from Marinier et al. (2004) for each province, taking into account regional differences in production practices and manure storage systems. A more complete description of the derivation of the distribution factor for manure management systems is contained in Appendix A3.3, Section A3.3.4.4.

Methane emission factors for manure management are estimated using the IPCC Tier 2 methodology. All parameters were taken from expert consultations in Boadi et al. (2004) and Marinier et al. (2004, 2005) or from the IPCC (2006) Guidelines. For dairy and beef cattle, the Boadi et al. (2004) Tier 2 animal production model was used to derive gross energy of consumption (GE) from which volatile solids (VS) were estimated using Equation 4.16 of the Good Practice Guidance and manure ash contents from Marinier et al. (2004). Emission factors were derived using the CH<sub>4</sub> producing potential (B<sub>0</sub>) and CH<sub>4</sub> conversion factors (MCF) taken from the IPCC (2006) Guidelines, and manure management systems (MS) from Marinier et al. (2005). An increase in emission factors over the period of 1990 to 2002 (see Table A3-12 in Annex 3.3) reflects higher gross energy intake for dairy cattle due to increased milk productivity and for beef cattle due to increased live body weights (see Section 6.2.2).

Methane emission factors for all other livestock used VS taken from Marinier et al. (2004) based on dry matter intake, ash content and digestible energy derived from expert consultations. Manure management systems (MS) were taken from Marinier et al. (2005) and CH<sub>4</sub> producing potential (B<sub>0</sub>) and CH<sub>4</sub> conversion factors (MCF) were taken from the IPCC 2006 Guidelines. For swine, sheep and poultry, different parameters were used for subcategories based on size class for swine and sheep as well as for turkeys, broilers and layers in the poultry category.

#### 6.3.1.3. Uncertainties and Time-Series Consistency

The uncertainty associated with CH<sub>4</sub> emissions from manure management for the year 2004 was determined using a Monte Carlo technique based on the IPCC Tier 2 methodology (Hutchinson et al. 2007). Population uncertainty is reported in Section 6.2.3 for the 2004 uncertainty estimate. Estimated emission factor uncertainty was ±40% for dairy and from 28% to 41% for non-dairy cattle, depending on the animal subcategory. Uncertainties for manure management emission estimates were ±36% for dairy cows, ±28% for beef cattle, ±26% for swine, and 18–24% for other animals. Total uncertainty on the manure management value was estimated at ±29%.

The methodology and parameter data used in the calculation of emission factors are consistent for the entire time series (1990–2008) with the exception of milk production for dairy. Milk production from 1990 to 1999 in Ontario and the western provinces is estimated based on production data from eastern Canada.

#### 6.3.1.4. QA/QC and Verification

Methane emissions from manure management have undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC good practice. The activity data and methodologies are documented and archived in both paper and electronic forms. The IPCC Tier 2 CH<sub>4</sub> emission factors for manure management practices by all animal categories derived from Marinier et al. (2004) have been reviewed by independent experts (Patni and Desjardins 2004). These documents have been archived in both paper and electronic form.

The Manure Management source category is also a part of a Tier 2 QA/QC for the Agriculture Sector (MacDonald et al. 2010). This analysis included i) rebuilding and analyzing

the database model, by retracing all documented changes to the model and comparing calculation methodology; ii) a complete review of method rationale, implementation and activity data selection; iii) a review and compilation of Canadian literature related to methane production from enteric fermentation; and iv) an assessment of the methodology in terms of its strengths and weaknesses.

The analysis of the database model and review of method rationale and implementation resulted in a number of recalculations, outlined in Section 6.3.1.5 and Table 6–2. The review and compilation of Canadian research on manure management demonstrated the scarcity of data on emissions from stored manure in Canada. No specific bias can be clearly identified in the IPCC Tier 2 model parameters due to the high variability in research results and the lack of supporting information for research carried out on manure storage installations. There is no clear standard to evaluate if IPCC parameters are appropriate for estimating emissions from manure management systems. Details of this review can be found in Annex 3.3, section A3.3.4.7. More standardized and detailed research is required in Canada to improve upon the current Tier 2 methodology.

### 6.3.1.5. Recalculations

Significant recalculations resulted from the Tier 2 analysis of the manure management database. As was the case with enteric fermentation, some of the issues with the database were identified in the 1990 to 2008 inventory submission (Table 6–2). Corrections to the database shifted estimates either up or down, depending on the source of error, the year and the variables being corrected (Table 6–4). In the Dairy Cattle category, there was an overall increase in emissions of roughly 0.13 Mt CO<sub>2</sub> eq or 21% in 1990, largely due to errors in milk data in the province of Quebec. In 2007, the recalculation was smaller, an increase of 0.05 Mt CO<sub>2</sub> eq or 10% due to the correction to gross energy, and there was a smaller increase of 1.5% correction for the year 2008, as most of the corrections had been completed for the database in the 2010 submission.

The Tier 2 analysis also resulted in recalculations to the Non-dairy Cattle category, with decreases of 4.2% or 0.03 Mt CO<sub>2</sub> eq in 1990, 6.2% or 0.05 Mt CO<sub>2</sub> eq in 2007 and 6.6% or 0.06 Mt CO<sub>2</sub> eq in 2008. Details on the nature of these recalculations are outlined in Section 6.2.5. The largest recalculation was due to a reduction in the total quantity of manure produced by slaughter heifers and steers. The impact of these corrections resulted in an overall decrease in non-dairy emission estimates for all years. However, the

Table 6–4 Recalculations for manure management estimates from cattle, absolute and percent change<sup>1</sup>

Recalculation	1990		2007		2008	
	Mt CO <sub>2</sub> eq	%	Mt CO <sub>2</sub> eq	%	Mt CO <sub>2</sub> eq	%
<b>Dairy Cattle</b>	0.131	21.3	0.052	10.3	0.008	1.52
Database errors – milk data – Quebec	0.067	11.66	NO	NO	NO	NO
Database errors – Net energy weight loss	0.055	7.04	0.033	6.42	-0.002	-0.437
Database errors – cattle weights <sup>2</sup>	<-0.001 <sup>3</sup>	-0.038	<-0.001	-0.035	<-0.001	-0.038
Database errors – Animal classification and weighting <sup>2</sup>	<-0.001	-0.038	<-0.001	-0.035	<-0.001	-0.038
Database errors – manure management systems	0.025	4.34	0.020	3.96	0.011	2.021
Activity data – Population data changes <sup>2</sup>	NO	NO	<-0.001	-0.035	<0.001	0.013
<b>Beef Cattle</b>	-0.029	-4.18	-0.052	-6.21	-0.055	-6.62
Database errors – cattle weights	-0.001	-0.108	-0.001	-0.158	-0.001	-0.125
Database errors – Animal classification and weighting	-0.005	-0.76	-0.009	-1.10	-0.007	-0.873
Database errors – Net energy weight loss	0.011	1.55	0.019	2.27	NO	NO
Activity data – Population data changes	NO	NO	-0.001	-0.102	<-0.001	<-0.001
Implementation of slaughter heifers & steers	-0.03	-4.87	-0.06	-7.1	-0.047	-5.62
<b>Total Cattle</b>	0.102	0.08	<-0.001	<-0.001	-0.047	-3.38

Notes:

1. Totals may not add up due to rounding.

2. Small recalculations were estimated by difference and averaged.

3. Values are less than 3 significant figures but +/- sign is correct.

NO-not occurring

decrease was smaller in 1990 compared to 2007 and 2008, as populations of finishing animals were smaller relative to total beef populations in 1990 than in 2007 and 2008.

Overall, the combination of these recalculations to dairy and beef cattle resulted in an increase in the emission estimate for cattle of 0.10 Mt CO<sub>2</sub> eq, or 0.1 % in 1990, no change in 2007 and a decrease of 0.05 Mt CO<sub>2</sub> eq or 3.4% in 2008. These recalculations, mainly due to the attenuating effect of the large recalculation in dairy cattle for 1990, resulted in a change in trend in manure management emissions; emissions increased by only 10% between 1990 and 2008, compared to 16%, as was indicated in the 2010 inventory submission.

Small recalculations were made for the Buffalo category, as provincial non-dairy cattle emission factors were used as opposed to the emission factor for steers. This change resulted in an increase in emissions from this category of 0.001 Mt CO<sub>2</sub> eq, (15%) but had a less than 0.1% impact on manure management emissions. Updates to swine populations for 2005 to 2007 resulted in changes in emission estimates from an increase of 0.3% or 0.01 Mt CO<sub>2</sub> eq in 2005 to a decrease of 1.1% or 0.03 Mt CO<sub>2</sub> eq in 2007.

### 6.3.1.6. Planned Improvements

Analysis of the manure management model suggested that improvements could be made to the values used for the distribution of manure management systems (AWMS) based on Statistics Canada farm environmental surveys. Preliminary work has begun to analyze whether this data source can be used for inventory purposes.

As pointed out in Section 6.2.6, work has been done on an updated Tier 2 uncertainty analysis for manure management emissions in Karimi-Zandashty et al. (2011). Further work is required to finalize uncertainty analyses.

As noted in Section 6.2.6, consultations are continuing on incorporating a time series for digestible energy used in the calculation of volatile solids and for certain animal categories that will be incorporated over the medium term.

## 6.3.2. N<sub>2</sub>O Emissions from Manure Management (CRF Category 4.B (b))

### 6.3.2.1. Source Category Description

The production of N<sub>2</sub>O during storage and treatment of animal waste occurs during nitrification and denitrification of nitrogen contained in the manure. Nitrification is the oxidation of ammonium (NH<sub>4</sub><sup>+</sup>) to nitrate (NO<sub>3</sub><sup>-</sup>), and denitrification is the reduction of NO<sub>3</sub><sup>-</sup> to N<sub>2</sub>O or N<sub>2</sub>. In general, the quantity of N<sub>2</sub>O produced increases with greater aeration of the manure. Manure from beef cattle, sheep, lamb, goats and horses are mainly handled with a solid and drylot system, which is the manure management system that emits the most N<sub>2</sub>O. Nitrous oxide emissions from manure on pasture, range and paddock by grazing animals are reported separately (see Section 6.4.2 - Manure on Pasture, Range and Paddock).

### 6.3.2.2. Methodological Issues

There have been very little published data on N<sub>2</sub>O emissions from manure management storage in Canada or in regions with comparable practices and climatic conditions as Canada. Emissions of N<sub>2</sub>O from manure management are estimated using the IPCC Tier 1 methodology. Emissions are calculated for each animal category by multiplying the animal population of a given category by its nitrogen excretion rate and by the emission factor associated with the animal waste management system (AWMS).

The animal population data are the same as those used for the Enteric Fermentation category estimates (Section 6.2) and CH<sub>4</sub> Emissions from Manure Management (Section 6.3.1). The average annual nitrogen excretion rates for domestic animals are taken from the IPCC (2006) Guidelines. The amount of manure nitrogen subject to losses because of leaching and volatilization of NH<sub>3</sub> and NO<sub>x</sub> during storage is adjusted by animal type and manure management system according to the default values provided in the IPCC 2006 Guidelines.

The fraction of nitrogen available for conversion into N<sub>2</sub>O is estimated by applying system-specific emission factors to the manure nitrogen handled by each management system. The IPCC 2006 default emission factors for a developed country with a cool climate are used to estimate manure nitrogen emitted as N<sub>2</sub>O for each type of AWMS.



### 6.3.2.3. Uncertainties and Time-Series Consistency

Uncertainties associated with N<sub>2</sub>O emission estimates from manure management result from uncertainties associated with estimates of animal populations from the *Census of Agriculture* and range from  $\pm 1\%$  to  $\pm 15\%$ , as noted in Sections 6.2.3 and 6.3.1.3. Uncertainties associated with rates of nitrogen excretion are  $\pm 20\%$  (IPCC 2006), with types of AWMS are  $\pm 20\%$  (Marinier et al. 2004), and with the emission factors are  $\pm 20\%$  (IPCC 2006). The overall level and trend uncertainties associated with this source of emission were estimated to be  $\pm 25\%$  and  $\pm 21\%$ , respectively (Hutchinson et al. 2007).

The same methodology, emission factors, and data sources are used for the entire time series (1990–2009).

### 6.3.2.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodology, and changes to methodologies are documented and archived in both paper and electronic form.

### 6.3.2.5. Recalculations

Recalculations have been carried out for this source because of the reduction in the quantity of manure N excreted by slaughter heifers and steers (see Section 6.2.5). As a result, emissions from this source have decreased by 0.43 Mt CO<sub>2</sub> eq in 1990 and 0.62 Mt CO<sub>2</sub> eq in 2008. As the reduction in emissions due to the recalculation was greater for 2008, this reduced the trend in N<sub>2</sub>O emissions from AWMS from 32% to 30%.

### 6.3.2.6. Planned Improvements

Data from direct measurements of N<sub>2</sub>O emissions from manure management in Canada are scarce. Recent scientific advances in analytical techniques allow direct measurements of N<sub>2</sub>O emissions from point sources. However, it will likely take several years before N<sub>2</sub>O emissions can be reliably measured and verified for various manure management systems in Canada.

## 6.4. N<sub>2</sub>O Emissions from Agricultural Soils (CRF Category 4.D)

Emissions of N<sub>2</sub>O from agricultural soils consist of direct and indirect emissions as well as emissions from animal manure deposited on pasture, range and paddock. The emissions of N<sub>2</sub>O from anthropogenic nitrogen inputs occur directly from the soils to which the nitrogen is added, and also indirectly through two pathways: i) volatilization of nitrogen from synthetic fertilizer and manure as NH<sub>3</sub> and NO<sub>x</sub> and its subsequent deposition off-site; and ii) leaching and runoff of synthetic fertilizer, manure and crop residue N. Changes in crop rotations and management practices such as summerfallow, tillage and irrigation, can also affect direct N<sub>2</sub>O emissions by altering mineralization of organic nitrogen, nitrification and denitrification.

### 6.4.1. Direct N<sub>2</sub>O Emissions from Soils (CRF Category 4.D.1)

Direct sources of N<sub>2</sub>O from soils include the application of synthetic nitrogen fertilizers and animal manure, crop residue decomposition, and cultivation of histosols. In addition, Canada also reports three country-specific sources of emissions/removals driven by soil organic matter decay from tillage practices, summerfallow and irrigation. Emissions/removals from these sources are estimated similarly based on nitrogen inputs from the application of synthetic nitrogen fertilizers and animal manure and crop residue nitrogen.

#### 6.4.1.1. Synthetic Nitrogen Fertilizers

##### 6.4.1.1.1. Source Category Description

Synthetic fertilizers add large quantities of nitrogen to agricultural soils. This added nitrogen undergoes transformations, such as nitrification and denitrification, which release N<sub>2</sub>O. Emission factors associated with fertilizer application depend on many factors, such as soil types, climate, topography, farming practices and environmental conditions (Gregorich et al. 2005; Rochette et al. 2008b).

##### 6.4.1.1.2. Methodological Issues

Canada has developed a country-specific, Tier 2 methodology to estimate N<sub>2</sub>O emissions from synthetic nitrogen fertilizer application on agricultural soils, which takes into



account moisture regimes and topographic conditions. Emissions of  $\text{N}_2\text{O}$  are estimated by ecodistrict and are scaled up at provincial and national levels. The amount of nitrogen applied is obtained from yearly fertilizer sales, which are available from regional fertilizer associations and compiled at the national level. All synthetic nitrogen fertilizers sold by retailers are assumed to be applied for crop production in Canada; the quantity of fertilizers applied to forests is deemed negligible. More details on the inventory method can be found in Annex 3.3.

#### 6.4.1.1.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $\text{N}_2\text{O}$  emission estimates from synthetic nitrogen fertilizer applications result from the uncertainty in estimates of nitrogen fertilizer application ( $\pm 20\%$ ); estimates of  $\text{EF}_{\text{BASE}}$  ( $\pm 25\%$ ), an ecodistrict-based  $\text{N}_2\text{O}$  emission factor; and estimates of  $\text{RF}_{\text{TEXTURE}}$ , a ratio factor adjusting  $\text{EF}_{\text{BASE}}$  for soil texture ( $\pm 30\%$ ). These terms and emission calculations are explained in Annex 3.3. The overall level and trend uncertainties associated with this source of emissions were estimated to be  $\pm 21\%$  and  $\pm 19\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.1.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

While Statistics Canada conducts QC checks before the release of synthetic nitrogen fertilizer consumption data, the Greenhouse Gas Division of Environment Canada carries out its own Tier 2 QC checks through historical records and consultations with regional and provincial agricultural industries.

Emissions of  $\text{N}_2\text{O}$  associated with synthetic fertilizer nitrogen applications on agricultural soils in Canada vary on a site-by-site basis, but there is a close agreement between the IPCC default emission factor of 1% (IPCC 2006) and the measured emission factor of 1.2% in eastern Canada, excluding emissions during the spring thaw period (Gregorich et al. 2005).

### 6.4.1.1.5. Recalculations

In this submission, there has been recalculation involved in this source of emission estimates because of lower quantities of animal manure produced (see Section 6.3.2.5). This results in a change in the redistribution of synthetic nitrogen fertilizers among ecodistricts, and leads to recalculations in the range of  $\pm 1$  kt  $\text{CO}_2$  eq for the entire time series.

#### 6.4.1.1.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this source.

### 6.4.1.2. Manure Applied as Fertilizer

#### 6.4.1.2.1. Source Category Description

The application of animal manure as fertilizer to agricultural soils can increase the rate of nitrification and denitrification and result in enhanced  $\text{N}_2\text{O}$  emissions. Emissions from this category include all the manure managed by drylot, liquid and other animal waste management systems. Manure deposited on pasture by grazing animals is accounted for in Section 6.4.2.

#### 6.4.1.2.2. Methodological Issues

Similar to the methodology used to estimate emissions from synthetic nitrogen fertilizers, the method used to estimate  $\text{N}_2\text{O}$  emissions from animal manure applied to agricultural soils is a country-specific IPCC Tier 2 method that takes into account moisture regimes (long-term growing season precipitation and potential evapotranspiration) and topographic conditions. Emissions are calculated by multiplying the amount of manure nitrogen applied to agricultural soils by an emission factor for each ecodistrict, and summed at the provincial and national levels. All manure that is handled by AWMS, except for the manure deposited on pasture, range and paddock from grazing animals, is assumed to be subsequently applied to agricultural soils (see Section 6.4.2).

#### 6.4.1.2.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $\text{N}_2\text{O}$  emission estimates from animal manure applied as fertilizers result from uncertainties associated with estimates of manure nitrogen based on types of animal population ( $\pm 1\% \sim \pm 15\%$ ), average animal manure nitrogen excretion rate ( $\pm 20\%$ ), manure

nitrogen loss ( $\pm 20\%$ ),  $RF_{\text{TEXTURE}}$  ( $\pm 30\%$ ), and  $EF_{\text{BASE}}$  ( $\pm 25\%$ ). The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated to be  $\pm 32\%$  and  $\pm 28\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.2.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are documented and archived in both paper and electronic forms.

#### 6.4.1.2.5. Recalculations

Recalculations have been carried out for this source because of the reduction in the total amount of manure N that is applied to the agricultural soils as noted in Section 6.3.2.5. Emissions from this source have decreased by 0.10 Mt CO<sub>2</sub> eq in 1990 and 0.16 Mt CO<sub>2</sub> eq in 2008.

#### 6.4.1.2.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this source.

### 6.4.1.3. Biological Nitrogen Fixation

#### 6.4.1.3.1. Source Category Description

Biological nitrogen fixation by the legume-rhizobium association is not considered a source of N<sub>2</sub>O emissions by Canada. The decision to exclude this category as an emission source is supported by the findings of Rochette and Janzen (2005) that there is no evidence that measurable amounts of N<sub>2</sub>O are produced during the nitrogen fixation process. Canada reports this source as “not occurring.” However, the contribution of legume nitrogen to N<sub>2</sub>O emissions from crop residue decomposition is still included (see Section 6.4.1.4).

### 6.4.1.4. Crop Residue Decomposition (CRF Category 4.D.4)

#### 6.4.1.4.1. Source Category Description

When a crop is harvested, a portion of the crop is left on Canada’s 2011 UNFCCC Submission

the field to decompose. The remaining plant matter is a nitrogen source for nitrification and denitrification and thus can contribute to N<sub>2</sub>O production.

#### 6.4.1.4.2. Methodological Issues

Emissions are estimated using an IPCC Tier 2 approach based on the amount of nitrogen contained in crop residue multiplied by the emission factor at the ecodistrict level and scaled up to the provincial and national levels. The amount of nitrogen contained in crop residues from both nitrogen-fixing and non-nitrogen-fixing crops is estimated using country-specific crop characteristics (Janzen et al. 2003). Emission factors are determined using the same approach as for synthetic fertilizer nitrogen application, using moisture regimes and topographic conditions.

#### 6.4.1.4.3. Uncertainties and Time-Series Consistency

Uncertainties associated with N<sub>2</sub>O emission estimates from crop residue decomposition result from uncertainties associated with estimates of crop residue nitrogen returned to the soil based on crop production data ( $\pm 15\%$ ), above- and below-ground crop residue nitrogen concentration ( $\pm 15\%$ ),  $RF_{\text{TEXTURE}}$  ( $\pm 30\%$ ), and  $EF_{\text{BASE}}$  ( $\pm 25\%$ ). The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated to be  $\pm 23\%$  and  $\pm 20\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.4.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are documented and archived in both paper and electronic form.

#### 6.4.1.4.5. Recalculations

Recalculations have been carried out for this source because of the inclusion of emission estimates from crop residue burning. The amount of crop residue N contained in the burnt residue is removed from crop residue decomposition to avoid double counting. As a result, emissions from this source have decreased by 0.09 Mt CO<sub>2</sub> eq in 1990 and 0.02 Mt CO<sub>2</sub> eq in 2008.

#### 6.4.1.4.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this source.

### 6.4.1.5. Cultivation of Organic Soils (Histosols)

#### 6.4.1.5.1. Source Category Description

Cultivation of organic soils (histosols) for crop production usually involves drainage, lowering the water table and increasing aeration, which enhance the decomposition of organic matter and nitrogen mineralization. The enhancement of decomposition upon the cultivation of histosols can result in greater denitrification and nitrification, and thus higher N<sub>2</sub>O production (Mosier et al. 1998).

#### 6.4.1.5.2. Methodological Issues

The IPCC Tier 1 methodology is used to estimate N<sub>2</sub>O emissions from cultivated organic soils. Emissions of N<sub>2</sub>O are calculated by multiplying the area of cultivated histosols by the IPCC default emission factor.

Areas of cultivated histosols at a provincial level are not surveyed in the *Census of Agriculture*. Consultations with numerous soil and crop specialists across Canada have resulted in an estimated area of 16 kha of cultivated organic soils in Canada, a constant level for the period 1990–2009 (Liang et al. 2004a).

#### 6.4.1.5.3. Uncertainties and Time Series Consistency

Uncertainties associated with N<sub>2</sub>O emission estimates from the cultivation of histosols result from uncertainties associated with area estimates of cultivated histosols (±50%) and emission factors (±50%). The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated to be ±67% and ±65%, respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.5.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are

documented and archived in both paper and electronic form.

#### 6.4.1.5.5. Recalculations

There has been no recalculation in this source of emission estimates over the last two years.

#### 6.4.1.5.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source.

### 6.4.1.6. Changes in N<sub>2</sub>O Emissions from Adoption of No-Till and Reduced Tillage

#### 6.4.1.6.1. Source Category Description

This category is not derived from additional nitrogen inputs (i.e. fertilizer, manure or crop residue); rather, it is implemented as modifications to N<sub>2</sub>O emission factors due to the change from conventional to conservation tillage practices—namely, reduced tillage (RT) and no-tillage (NT).

#### 6.4.1.6.2. Methodological Issues

Compared with conventional or intensive tillage (IT), direct seeding or NT as well as RT change several factors that influence N<sub>2</sub>O production, including decomposition of soil organic matter, soil carbon and nitrogen availability, soil bulk density, and water content (McConkey et al. 1996, 2003; Liang et al. 2004b). As a result, compared with conventional tillage, conservation tillage (i.e. RT and NT) generally reduces N<sub>2</sub>O emissions for the Prairies (Malhi and Lemke 2007), but increases N<sub>2</sub>O emissions for the non-Prairie regions of Canada (Rochette et al. 2008a). The net result across the country amounts to a small reduction in emissions. This reduction is reported separately, as a negative estimate (Table 6–1).

Changes in N<sub>2</sub>O emissions resulting from the adoption of NT and RT are estimated through modifications of emission factors for synthetic fertilizers, manure nitrogen applied to cropland, and crop residue nitrogen decomposition. This subcategory is kept separate from the fertilizer and crop residue decomposition source categories to preserve the transparency in reporting; however, this separation causes negative emissions to be reported. An empirically derived tillage factor ( $F_{TILL}$ ), defined as the ratio

of mean  $\text{N}_2\text{O}$  fluxes on NT or RT to mean  $\text{N}_2\text{O}$  fluxes on IT ( $\text{N}_{2\text{O}_{\text{NT}}}/\text{N}_{2\text{O}_{\text{IT}}}$ ), represents the effect of NT or RT on  $\text{N}_2\text{O}$  emissions (see Annex 3.3).

#### 6.4.1.6.3. Uncertainties and Time-Series Consistency

Uncertainties associated with changes in  $\text{N}_2\text{O}$  emission estimates from adoption of NT and RT result from uncertainties associated with area estimates of NT and RT from the *Census of Agriculture* ( $\pm 15\%$ ),  $F_{\text{TILL}}$  ( $\pm 20\%$ ), and  $\text{EF}_{\text{BASE}}$  ( $\pm 25\%$ ). The overall level and trend uncertainties associated with this source of emission/removal estimates from 1990 to 2009 have not been assessed.

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.6.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are documented and archived in both paper and electronic form.

#### 6.4.1.6.5. Recalculations

There is no change in the area of conservation tillage, but recalculations have occurred in this category as a result of changes in nitrogen inputs from manure and crop residue decomposition (see Sections 6.2.5 and 6.4.1.4.5). These recalculations have changed avoided emissions slightly by  $-1.6 \text{ kt CO}_2 \text{ eq}$  in 1990 and  $9.4 \text{ kt CO}_2 \text{ eq}$  in 2008.

#### 6.4.1.6.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source. However, work is ongoing to develop level and trend uncertainty estimates using the IPCC Tier 2 method.

### 6.4.1.7. $\text{N}_2\text{O}$ Emissions Resulting from Summerfallowing

#### 6.4.1.7.1. Source Category Description

This category is not derived from additional nitrogen input but reflects changes in soil conditions that affect  $\text{N}_2\text{O}$  emissions. Summerfallow (SF) is a farming practice typically

used in the Prairie region to conserve soil moisture by leaving the soil unseeded for an entire growing season in a crop rotation. During the fallow year, several soil factors may stimulate  $\text{N}_2\text{O}$  emissions relative to a cropped situation, such as higher soil water content, higher soil temperature, and greater availability of soil carbon and nitrogen (Campbell et al. 1990, 2005).

#### 6.4.1.7.2. Methodological Issues

Experimental studies have shown that  $\text{N}_2\text{O}$  emissions in fallow fields are not statistically different from emissions on continuously cropped fields (Rochette et al. 2008b). Omitting areas under SF in calculations of  $\text{N}_2\text{O}$  emissions because no crops are grown or fertilizer applied could lead to underestimating total  $\text{N}_2\text{O}$  emissions. The emissions from SF land are therefore calculated through a country-specific method by summing emissions from fertilizer nitrogen, manure nitrogen application to annual crops and crop residue nitrogen for a given ecodistrict and multiplying the sum by the proportion of that ecodistrict area under SF (Rochette et al. 2008b). A more detailed description of the approach is provided in Annex 3.3. This subcategory is reported separately from the Synthetic Nitrogen Fertilizers, Animal Manure Applied as Fertilizers and Crop Residue Decomposition source categories to enhance the transparency of the reporting process.

#### 6.4.1.7.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $\text{N}_2\text{O}$  emission estimates from summerfallow result from uncertainties associated with area estimates of summerfallow from the *Census of Agriculture* (cropland area: up to  $\pm 10\%$ ; the fraction of cropland that is under summerfallow ( $\text{FRAC}_{\text{FALLOW}}$ ): up to  $\pm 10\%$ ) and  $\text{EF}_{\text{BASE}}$ :  $\pm 25\%$ . The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated to be  $\pm 24\%$  and  $\pm 21\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.7.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are

documented and archived in both paper and electronic form.

#### 6.4.1.7.5. Recalculations

There is no change in the area of summerfallow, but recalculations have occurred in this category as a result of changes in nitrogen inputs from manure nitrogen excretion from slaughter heifers and steers (see Section 6.2.5), and crop residue burning (see Section 6.4.1.4.5). These recalculations have increased emissions slightly—by 94 kt CO<sub>2</sub> eq in 1990 and 17 kt CO<sub>2</sub> eq in 2008.

#### 6.4.1.7.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source.

### 6.4.1.8. N<sub>2</sub>O Emissions from Irrigation

#### 6.4.1.8.1. Source Category Description

Similar to tillage practices and summerfallow, the effect of irrigation on N<sub>2</sub>O emissions is not derived from additional nitrogen input but reflects changes in soil conditions that affect N<sub>2</sub>O emissions. Higher soil water content under irrigation increases the potential for N<sub>2</sub>O emissions through increased biological activity, reducing soil aeration (Jambert et al. 1997), thus enhancing denitrification.

#### 6.4.1.8.2. Methodological Issues

The methodology is country specific and is based on the assumptions that 1) irrigation water stimulates N<sub>2</sub>O production in a way similar to rainfall water and 2) irrigation is applied at rates such that amounts of precipitation plus those of irrigation water are equal to the potential evapotranspiration at the local conditions. Consequently, the effect of irrigation on N<sub>2</sub>O emissions from agricultural soils was estimated using an EF<sub>BASE</sub> estimated at a P/PE = 1 (or EF<sub>BASE</sub> = 0.017 N<sub>2</sub>O-N/kg N) for the irrigated areas of a given ecodistrict.

#### 6.4.1.8.3. Uncertainties and Time-Series Consistency

Uncertainties associated with N<sub>2</sub>O emission estimates from irrigation result from uncertainties associated with synthetic nitrogen ( $\pm 20\%$ ) and animal manure nitrogen inputs ( $\pm 20\%$ ), crop residue nitrogen ( $\pm 15\%$ ), area estimates of irrigated cropland from the *Census of Agriculture*

(up to  $\pm 10\%$ ), as well as EF<sub>BASE</sub> ( $\pm 25\%$ ). The overall level and trend uncertainties associated with this source of emission estimates have not been assessed at this time.

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.1.8.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data and methodology are documented and archived in both paper and electronic form.

#### 6.4.1.8.5. Recalculations

Recalculations have occurred in this category because of reduced nitrogen inputs: 1) resulting from slaughter heifers and steers (see Section 6.2.5), and 2) the inclusion of emissions from crop residue burning (see Section 6.4.1.4.5). As a result, emissions from this source have changed slightly in a range of  $\pm 4$  kt CO<sub>2</sub> eq for the entire time series.

#### 6.4.1.8.6. Planned Improvements

The level and trend uncertainties associated with this source of estimates are planned to be assessed in future submissions.

## 6.4.2. Manure on Pasture, Range and Paddock (CRF Category 4.D.2)

### 6.4.2.1. Source Category Description

When manure is excreted on pasture, range and paddock by grazing animals, nitrogen in the manure undergoes transformations, such as ammonification, nitrification and denitrification. During these transformation processes, N<sub>2</sub>O can be emitted.

### 6.4.2.2. Methodological Issues

The emissions from manure excreted by grazing animals are calculated using the IPCC Tier 1 methodology (IPCC/OECD/IEA 1997). Emissions are calculated for each animal category by multiplying the number of grazing animals for that category by the appropriate nitrogen excretion



rate and by the fraction of manure nitrogen available for conversion to  $\text{N}_2\text{O}$ .

The animal population data are the same as those used in the Enteric Fermentation category (see Section 6.2). The number of grazing animals for each category or subcategory is calculated using the total population multiplied by the fraction of animals grazed on pasture, range and paddock. The nitrogen excretion rates are based on the IPCC (2006) defaults. The fraction of manure nitrogen available for conversion to  $\text{N}_2\text{O}$  is calculated as the percentage of total manure nitrogen produced on pasture, range and paddock multiplied by the IPCC 2006 default values of 0.02 kg  $\text{N}_2\text{O-N/kg N}$  for cattle and 0.01 kg  $\text{N}_2\text{O-N/kg N}$  for sheep/lamb, goat and horse.

#### 6.4.2.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $\text{N}_2\text{O}$  emission estimates from animal manure on pasture, range and paddock result from uncertainty in number of grazing animals ( $\pm 1\%$  to  $\pm 15\%$ ); manure nitrogen excretion rate ( $\pm 20\%$ ); fraction of manure nitrogen on pasture, range and paddock ( $\pm 20\%$ ); as well as emission factors ( $-25\% \sim +150\%$ ). The overall level and trend uncertainties associated with these estimates from 1990 to 2009 were estimated to be  $\pm 19\%$  and  $\pm 21\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.2.4. QA/QC and Verification

The activity data, methodologies, and changes to methodologies are documented and archived in both paper and electronic form. QC checks and cross-checks have been carried out to identify data entry errors and calculation errors. In general, there are very few data available on the quantity of  $\text{N}_2\text{O}$  emissions from the manure on pasture, range and paddock from grazing animals in Canada. Therefore, it is extremely difficult to verify how well the IPCC emission factors reflect Canadian conditions.

#### 6.4.2.5. Recalculations

Recalculations have occurred in this category because of the reduction in the total amount of manure nitrogen deposited on pasture, range and paddock from slaughter heifers and steers (see Section 6.2.5). As a result, emissions from this source have decreased by 0.42 Mt  $\text{CO}_2$  eq in 1990

and 0.62 Mt  $\text{CO}_2$  eq in 2008.

#### 6.4.2.6. Planned Improvements

Emissions of  $\text{N}_2\text{O}$  from the Manure Deposited on Pasture, Range and Paddock by grazing animals for Canada is a key source category according to the level assessment. Because of its importance, Canada has initiated a three-year research project, starting in September 2009, to quantify  $\text{N}_2\text{O}$  emissions from animal manure on pasture, range and paddock by grazing dairy and beef cattle. It is expected that, by the end of this project, Canada will be able to estimate  $\text{N}_2\text{O}$  emissions for this source using country-specific emission factors.

### 6.4.3. Indirect Emissions of $\text{N}_2\text{O}$ from Soils (CRF Category 4.D.3)

A fraction of the nitrogen from both synthetic fertilizer and manure that is applied to agricultural fields is transported off-site through volatilization in the form of  $\text{NH}_3$  and  $\text{NO}_x$  and subsequent redeposition or leaching, erosion and runoff. The nitrogen that is transported from the agricultural field in this manner provides additional nitrogen for subsequent nitrification and denitrification to produce  $\text{N}_2\text{O}$ .

#### 6.4.3.1. Volatilization and Redeposition of Nitrogen

##### 6.4.3.1.1. Source Category Description

When synthetic fertilizer or manure is applied to cropland, a portion of the nitrogen is lost through volatilization in the form of  $\text{NH}_3$  or  $\text{NO}_x$ , which can be redeposited elsewhere and undergo further transformation, resulting in  $\text{N}_2\text{O}$  emissions off-site. The quantity of this volatilized nitrogen depends on a number of factors, such as rates of fertilizer and manure nitrogen application, fertilizer types, methods and time of nitrogen application, soil texture, rainfall, temperature, and soil pH.

##### 6.4.3.1.2. Methodological Issues

Leached or volatilized N may not be available for the process of nitrification and denitrification for many years, particularly in the case of N leaching into groundwater. There are few published scientific data that actually determine  $\text{N}_2\text{O}$  emissions from volatilization and redeposition of  $\text{NH}_3$  and  $\text{NO}_x$ . Even though Indirect Soil  $\text{N}_2\text{O}$  Emissions from Ag-



gricultural Soils are a key source category for level and trend assessments for Canada, there are difficulties in defining the duration and boundaries for this source of emissions because no standardized method for deriving the IPCC Tier 2 emission factors is provided by the IPCC Guidelines.

The IPCC Tier 1 methodology is used to estimate indirect  $N_2O$  emissions due to volatilization and redeposition of nitrogen from synthetic N fertilizers and animal manure. The amount of synthetic fertilizer and manure nitrogen is multiplied by the fraction of N that is volatilized as  $NH_3$ -N and  $NO_x$ -N and then by an emission factor. The amount of nitrogen applied is obtained from yearly fertilizer sales data, which are available from Statistics Canada, and from the amounts of manure nitrogen excreted by animals (see Annex 3.3). The amount of nitrogen that volatilizes is assumed to be 10% of the total amount of synthetic fertilizer applied and 20% of the applied manure nitrogen to cropland (IPCC 2006). The default IPCC emission factor, 0.01 kg  $N_2O$ -N/kg N, is used to derive the  $N_2O$  emission estimate (IPCC/OECD/IEA 1997).

#### 6.4.3.1.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $N_2O$  emission estimates from volatilization of  $NH_3$  and  $NO_x$  due to applications of synthetic and manure nitrogen result from uncertainty in estimates of synthetic fertilizer nitrogen consumption ( $\pm 20\%$ ), fraction of volatilized  $NH_3$  and  $NO_x$  from synthetic nitrogen fertilizers ( $\pm 20\%$ ), animal populations and categories ( $\pm 1\%$  to  $\pm 15\%$ ), manure nitrogen excretion rate ( $\pm 20\%$ ), fraction of volatilized  $NH_3$  +  $NO_x$  from animal manure ( $\pm 20\%$ ), as well as emission factors ( $-50\%$  to  $+300\%$ ). Indirect  $N_2O$  Emissions are among the source categories with the highest uncertainty in the inventory, with little prospect for short-term improvements. The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated at  $\pm 40\%$  and  $\pm 34\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.3.1.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (see details and references in Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes

to methodologies are documented and archived in both paper and electronic form.

#### 6.4.3.1.5. Recalculations

Recalculations have occurred in this category because of the reduction in the total amount of manure nitrogen either directly deposited on pasture, range and paddock or applied to the agricultural soils from slaughter heifers and steers (see Section 6.2.5). As a result, emissions from this source have decreased by 0.12 Mt  $CO_2$  eq in 1990 and 0.19 Mt  $CO_2$  eq in 2008.

#### 6.4.3.1.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source.

### 6.4.3.2. Leaching, Erosion, and Runoff

#### 6.4.3.2.1. Source Category Description

When synthetic fertilizer or manure nitrogen is applied to cropland, a portion of the nitrogen is lost through leaching, erosion and runoff. The magnitude of this loss depends on a number of factors, such as application rate and method, crop type, soil texture, rainfall and landscape. This portion of lost nitrogen can further undergo transformations, such as nitrification and denitrification, and can produce  $N_2O$  emissions off-site.

#### 6.4.3.2.2. Methodological Issues

There are few published scientific data that determine  $N_2O$  emissions from leaching, erosion and runoff in Canada. Similar to  $N_2O$  emissions from volatilization and redeposition of ammonia and  $NO_x$ , this source is poorly defined because no standardized method for deriving the IPCC Tier 2 emission factors is provided by the IPCC Guidelines.

A modified IPCC Tier 1 methodology is used to estimate indirect  $N_2O$  emissions from leaching, runoff and erosion of fertilizers, manure and crop residue nitrogen from agricultural soils. Indirect  $N_2O$  emissions from runoff and leaching of nitrogen at the ecodistrict level are estimated using  $FRAC_{LEACH}$  multiplied by the amount of synthetic fertilizer nitrogen, non-volatilized manure nitrogen and crop residue nitrogen and by an emission factor of 0.025 kg  $N_2O$ -N/kg N (IPCC 2000).

The default value for the fraction of nitrogen that is lost

through leaching and runoff ( $\text{FRAC}_{\text{LEACH}}$ ) in the Revised 1996 Guidelines is 0.3; however,  $\text{FRAC}_{\text{LEACH}}$  can reach values as low as 0.05 in regions where rainfall is much lower than potential evapotranspiration (IPCC 2006), such as in the Prairie region of Canada. Accordingly, it is assumed that  $\text{FRAC}_{\text{LEACH}}$  would vary among ecodistricts from a low of 0.05 to a high of 0.3. For ecodistricts with no moisture deficit during the growing season (May through October), the maximum  $\text{FRAC}_{\text{LEACH}}$  value of 0.3 recommended by the IPCC 2006 Guidelines is assigned. The minimum  $\text{FRAC}_{\text{LEACH}}$  value of 0.05 is assigned to ecodistricts with the greatest moisture deficit. For the remaining ecodistricts,  $\text{FRAC}_{\text{LEACH}}$  is estimated by the linear extrapolation of the two end-points described above.

#### 6.4.3.2.3. Uncertainties and Time-Series Consistency

Uncertainties associated with  $\text{N}_2\text{O}$  estimates from leaching, runoff and erosion of nitrogen from synthetic nitrogen fertilizers, animal manure and crop residue nitrogen result from uncertainties associated with estimates of synthetic fertilizer nitrogen consumption ( $\pm 20\%$ ), manure nitrogen excretion rate ( $\pm 20\%$ ), animal populations ( $\pm 1\% \sim \pm 15\%$ ), crop residue nitrogen ( $\pm 15\%$ ),  $\text{FRAC}_{\text{LEACH}}$  ( $\pm 50\%$ ), as well as the leaching/runoff emission factor  $\text{EF}_{\text{LEACH}}$  ( $-48\% \sim +200\%$ ). The overall level and trend uncertainties associated with this source of emission estimates from 1990 to 2009 were estimated to be  $\pm 32\%$  and  $\pm 29\%$ , respectively (Hutchinson et al. 2007).

The same methodology and emission factors are used for the entire time series (1990–2009).

#### 6.4.3.2.4. QA/QC and Verification

This category has undergone Tier 1-level QC checks as elaborated in the QA/QC plan (refer to details and references in Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data, methodologies, and changes to methodologies are documented and archived in both paper and electronic form.

#### 6.4.3.2.5. Recalculations

Recalculations have been carried out for this source mainly because of the reduction in the total amount of manure nitrogen from slaughter heifers and steers (see Section 6.2.5). The inclusion of emission estimates from crop residue burning also results in a small reduction in emissions

because the amount of crop residue nitrogen contained in the burnt residue is removed from crop residue decomposition (see Section 6.4.1.4.5). As a result, emissions from this source have decreased by 0.21 Mt  $\text{CO}_2$  eq in 1990 and 0.24 Mt  $\text{CO}_2$  eq in 2008.

#### 6.4.3.2.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source.

## 6.5. $\text{CH}_4$ and $\text{N}_2\text{O}$ Emissions from Field Burning of Agricultural Residues (CRF Category 4.F)

### 6.5.1. Source Category Description

Crop residues are sometimes burned in Canada, as a matter of convenience and disease control through residue removals, even though this practice has declined in recent years because of concerns over soil quality and environmental issues. Crop residue burning is a net source of  $\text{CH}_4$ , CO,  $\text{NO}_x$  and  $\text{N}_2\text{O}$  (IPCC/OECD/IEA 1997).

### 6.5.2. Methodological Issues

There are no published data on emissions of  $\text{N}_2\text{O}$  and  $\text{CH}_4$  from field burning of agricultural residues in Canada. Thus, the IPCC default emission factors and parameters from the Good Practice Guidance (IPCC 2000) were used for estimating emissions from this source.

A complete time series of activity data on the type and the percent of each crop residue subject to field burning was developed based on Statistics Canada's *Farm Environmental Management Survey* (FEMS)<sup>2</sup> and on expert consultations (Coote et al. 2008).

Crop-specific parameters, such as moisture content of the crop product and ratio of above-ground crop residue to crop product, required for estimating the amount of crop residue burned were obtained from Janzen et al. (2003).

2 <http://www.statcan.gc.ca/cgi-bin/imdb/p2SV.pl?Function=getSurvey&SDDS=5044&lang=en&db=imdb&adm=8&dis=2#a4>

### 6.5.3. Uncertainties and Time-Series Consistency

The uncertainties associated with CH<sub>4</sub> and N<sub>2</sub>O emissions from field burning of agricultural residues were determined using an IPCC Tier 1 method.

The most uncertain quantities associated with CH<sub>4</sub> and N<sub>2</sub>O emissions from field burning of agricultural residues are the amount of field crop residues burned and emission factors. The uncertainty in the amount of crop residues burned is estimated, based on the area of specific seeded crop, to be  $\pm 50\%$  (Coote et al. 2008). The uncertainties associated with the emission factors are not reported in the 2006 IPCC Guidelines but are assumed to be similar to those associated with burning of Savanna and grassland:  $\pm 40\%$  for CH<sub>4</sub> and  $\pm 48\%$  for N<sub>2</sub>O (IPCC 2006).

### 6.5.4. QA/QC and Verification

CH<sub>4</sub> and N<sub>2</sub>O emissions from field burning of agricultural residues have undergone Tier 1-level QC checks as elaborated in the QA/QC plan (Annex 6) in a manner consistent with IPCC Good Practice Guidance. The activity data and methodologies are documented and archived in both paper and electronic forms.

### 6.5.5. Recalculations

Since this is the first year that Canada has reported emissions of CH<sub>4</sub> and N<sub>2</sub>O from field burning of agricultural residues, no recalculations are involved in the 1990–2008 time series.

### 6.5.6. Planned Improvements

There is no immediate plan in place aimed at improving emission estimates from this source.

# Chapter 7

## Land Use, Land-use Change and Forestry (CRF Sector 5)

### 7.1. Overview

The Land Use, Land-use Change and Forestry (LULUCF) Sector reports greenhouse gas (GHG) fluxes between the atmosphere and Canada's managed lands, as well as those associated with land-use change. The assessment includes emissions and removals of CO<sub>2</sub>, additional emissions of CH<sub>4</sub>, N<sub>2</sub>O and CO due to wildfires and controlled burning, and N<sub>2</sub>O released following land conversion to cropland. All emissions from and removals by the LULUCF Sector are excluded from the national totals.

In 2009, the estimated net GHG flux in the LULUCF Sector, calculated as the sum of CO<sub>2</sub><sup>1</sup> emissions and removals and non-CO<sub>2</sub> emissions, amounted to removals of 12 Mt. If these were included in the national totals, they would decrease the total Canadian GHG emissions by about 2%. Table 7-1 provides the net flux estimates for 1990 and recent years in the major LULUCF Sector categories and subcategories.

In view of the high interannual variability displayed by some categories and its effect on the sectoral trends, the reader is cautioned against interpreting the figures in Table 7-1 as trends. The full time series of LULUCF Sector estimates is available in Table 10 of the common reporting format (CRF) series.

The Forest Land category has the largest influence on sectoral totals. The net fluxes are negative (removals) in 12 of the 20 years of the time series, and positive (emissions) for the remainder. Years with a net positive flux increased in frequency in the latter part of the time series. Net emissions are particularly large in years where large areas of managed forests were burned by wildfire. As a consequence, the interannual variability is high, with net

1 Unless otherwise indicated, all emissions and removals are in CO<sub>2</sub> equivalents.

Table 7-1 LULUCF Sector Net GHG Flux Estimates, Selected Years

Sectoral Category	Net GHG Flux (kt CO <sub>2</sub> eq) <sup>4</sup>							
	1990	2000	2004	2005	2006	2007	2008	2009
<b>Land Use, Land-Use Change and Forestry TOTAL<sup>1</sup></b>	<b>-67 000</b>	<b>-62 000</b>	<b>97 000</b>	<b>54 000</b>	<b>65 000</b>	<b>51 000</b>	<b>-17 000</b>	<b>-12 000</b>
<b>a. Forest Land</b>	<b>-93 000</b>	<b>-74 000</b>	<b>88 000</b>	<b>46 000</b>	<b>58 000</b>	<b>45 000</b>	<b>-22 000</b>	<b>-17 000</b>
Forest Land Remaining Forest Land	-92 000	-73 000	89 000	47 000	58 000	46 000	-21 000	-16 000
Land Converted to Forest Land	-1 000	-1 000	-1 000	-1 000	-900	-900	-800	-800
<b>b. Cropland</b>	<b>11 000</b>	<b>-140</b>	<b>-3 400</b>	<b>-4 300</b>	<b>-4 800</b>	<b>-5 400</b>	<b>-6 300</b>	<b>-6 900</b>
Cropland Remaining Cropland	-1 500	-6 900	-9 400	-10 000	-11 000	-11 000	-12 000	-12 000
Land Converted to Cropland	13 000	6 700	6 000	5 700	5 800	5 800	5 600	5 500
<b>c. Grassland</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>
Grassland Remaining Grassland	NE	NE	NE	NE	NE	NE	NE	NE
Land Converted to Grassland	NE	NE	NE	NE	NE	NE	NE	NE
<b>d. Wetlands</b>	<b>5 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>3 000</b>	<b>2 000</b>
Wetlands Remaining Wetlands	1 000	2 000	2 000	2 000	2 000	2 000	2 000	2 000
Land Converted to Wetlands	4 000	800	800	900	800	700	700	600
<b>e. Settlements</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>	<b>9 000</b>
Settlements Remaining Settlements	-100	-200	-200	-200	-200	-200	-200	-200
Land Converted to Settlements	9 000	9 000	9 000	9 000	9 000	9 000	9 000	9 000
Forest conversion (memo item) <sup>2</sup>	26 000	20 000	19 000	19 000	19 000	19 000	19 000	18 000
Grassland conversion (memo item) <sup>2,3</sup>	300	200	200	200	200	200	200	200

Notes:

1. Totals may not add up due to rounding. Annex 9 describes the rounding protocol.

2. Already included in land converted to cropland, land converted to wetlands, and land converted to settlements; and in cropland remaining cropland and wetlands remaining wetlands (for residual emissions post-20 years, 10 years for reservoirs).

3. Includes conversion of agricultural grassland to cropland and of tundra to settlement.

4. Negative sign indicates removal of CO<sub>2</sub> from the atmosphere.

5. NE = Not estimated.

category totals fluctuating between –110 Mt (1992) and 170 Mt (1995). These fluctuations are carried over to the LULUCF Sector totals, which vary between net emissions and net removals, depending on the net flux from managed forests.

Over the entire period, the Cropland category displays a steady trend towards decreasing emissions, with removals every year since 2000—up to 6.9 Mt in 2009. The 50% decline of emissions from land converted to cropland and growing removals by cropland remaining cropland equally contribute to the 18 Mt reduction in net emissions over the 1990–2009 period.

Over the 1990–2009 period, net fluxes in the Wetlands category (managed peatlands and flooded lands) fluctuate between 2 Mt and 5 Mt. Emissions from land converted to wetlands declined from a little less than 4 Mt to 0.6 Mt during the period. Current emissions from flooded lands account for 51% of all emissions in the Wetlands category compared to a share of 83% in 1990.

With this submission, Canada continues the implementation of a multi-year effort to substantially improve its estimates in the LULUCF Sector.<sup>2</sup> The contribution of the best Canadian expertise to this effort occurred within Canada's national, multidisciplinary framework for monitoring, accounting and reporting emissions and removals in managed lands. The Monitoring, Accounting and Reporting System (MARS) framework provides a means for coordinating, planning and integrating the activities of many groups of scientists and experts across several government levels and research institutions.

Work within MARS for LULUCF is expected to continue. In addition to enhanced collaboration within the framework, planned improvements include the completion of formal and documented uncertainty estimates in all LULUCF categories and quantification of missing land use and land-use change categories.

<sup>2</sup> First described in the NIR of the 2004 submission and implemented in the 2006 submission.

Table 7–2 List of Changes and Corresponding Implementation Date

List of Changes	Implementation	
	2010 Submission	2011 Submission
<b>Forest Land</b>		
Modification of algorithm used to simulate the growth of young forest stands	2008 inventory year	Entire time series
Re-calibration C transfer parameters for fire	2008 inventory year	Entire time series
Harvest statistics updates	2008 inventory year	2006–2008 inventory years
Additional insect monitoring data	2008 inventory year	2005–2008 inventory years
Wildfire area updates	2008 inventory year	2006–2008 inventory years
<b>Land Converted to Forest Land</b>		
Review of afforestation data	2008 inventory year	Entire time series
<b>Forest Conversion</b>		
Updated deforestation monitoring with inclusion of 2000–2008 mapping time period	2008 inventory year (Selected regions & categories)	Entire time series (All regions)
Detailed Q/C review and update of previous estimates	2008 inventory year (Selected regions & categories)	Entire time series (All regions)
Additional samples and records data	2009 inventory year (Selected regions & categories)	Entire time series (All regions)
<b>Cropland</b>		
After FLCL occurs in Western Canada, this land remains in the perennial category for 20 years		Entire time series
<b>Wetland</b>		
Modification of model used to estimate area of managed peatlands	2008 inventory year	Entire time series
Single temporal CO <sub>2</sub> Reservoir emission curve replaced the two regional curves		Entire time series
Reservoir activity data updates		Entire time series

This marks the second year that LULUCF estimates are reported under both the Convention and the Kyoto Protocol. The same definition and estimation approaches apply to all LULUCF estimates, hence there is no separate section in this chapter devoted to estimates reported under the Kyoto Protocol. Discrepancies between estimates reported under the Convention and the Kyoto Protocol are discussed in Annex 11.

The remainder of this chapter highlights the salient features of each LULUCF Sector category. Section 7.2 gives an overview of the representation of managed lands; each subsequent section provides a short description of a land category (Sections 7.3–7.7). A special section (Section 7.8) is devoted to the cross-category estimates of forest conversion to other lands.

This submission does not incorporate any change in response to the recommendations of the expert review team (ERT) that reviewed Canada's 2010 submission, as the review report was not yet available at the time this submission was prepared. However, the 2010 submission included several improvements that were only implemented to the estimates of the 2008 inventory year. In this year's submission, all improvements have been implemented over all applicable years (Table 7–2).

## 7.2. Land Category Definition and Representation of Managed Lands

In order to harmonize all land-based estimates, a common definitional framework was elaborated and adopted by all groups involved in estimate preparation. Definitions are consistent with the IPCC (2003) land categories, while remaining relevant to land management practices, prevailing environmental conditions and available data sources in Canada. This framework applies to all LULUCF estimates reported under the Convention and the Kyoto Protocol.

Forest land includes all areas of 1 ha or more where tree formations can reach 25% crown cover and 5 m in height *in situ*. Not all Canadian forests are under the direct influence of human activities, prompting the non-trivial question of what areas properly embody the “managed forests.” For the purpose of the GHG inventory, managed forests are those managed for timber and non-timber resources (including parks) or subject to fire protection. Annex 3.4

provides more detail on the implementation of the “managed forests” definition.

Agricultural land comprises both cropland and agricultural grassland. Cropland includes all lands in annual crops, summerfallow, and perennial crops (mostly forage, but also including berries, grapes, nursery crops, vegetables, and fruit trees and orchards). Agricultural grassland is defined as “unimproved” pasture or rangeland that is used only for grazing domestic livestock. It occurs only in geographical areas where the grassland would not naturally regrow to forest if abandoned: the natural shortgrass prairie in southern Saskatchewan and Alberta and the dry, interior mountain valleys of British Columbia. All agricultural land that is not grassland is *de facto* classified as cropland, including unimproved pastures where natural vegetation would be forest (eastern Canada and most of British Columbia).

Vegetated areas that do not meet the definition of forest land or cropland are generally classified as grassland: extensive areas of tundra in the Canadian north are considered un-managed grassland.

Wetlands are areas where permanent or recurrent saturated conditions allow the establishment of vegetation and soil development typical of these conditions and that are not already in forest land, cropland or agricultural grasslands. A national wetland inventory is under preparation (Fournier et al. 2007). Managed wetlands are those where human interventions have altered the water table—for example, peatlands drained for peat extraction or flooded lands (IPCC 2003).

Settlements include all built-up land: urban, rural residential, land devoted to industrial and recreational use; roads, rights-of-way and other transportation infrastructure; and resource exploration, extraction and distribution (mining, oil and gas). The diversity of this category has so far precluded an assessment of its extent in the Canadian landscape; however, it is often involved in land conversion, and the impact of forest land conversion to settlements is assessed in this GHG inventory.

Other land comprises areas of rock, ice or bare soil, and all land areas that do not fall into any of the other five categories. Currently, only emissions from the conversion of other land to reservoirs are reported, under the Wetlands category.



Table 7–3 Land Use and Land-use Change Matrix for the 2009 Inventory Year (Areas in kha)<sup>1</sup>

Initial Land Use	Final Land Use					
	Forest	Cropland	Grassland	Wetlands	Settlements	Other
Forest	229 400	473	NO	66	481	NO
Cropland	98	47 040	NO	NE	NE	NO
Grassland	NO	3	NE	NE	1	NO
Wetlands	NO	NE	NO	411 <sup>2</sup>	NE	NE
Settlements	NO	NE	NE	NO	NE	NO
Other	NO	NO	NO	36	NE	NE

Notes:

1. Non-diagonal cells refer to cumulative areas, i.e., total land converted over the last 20 years (10 years for reservoirs).

2. Only includes wetland areas for which emissions are reported in the CRF.

NE = Not estimated.

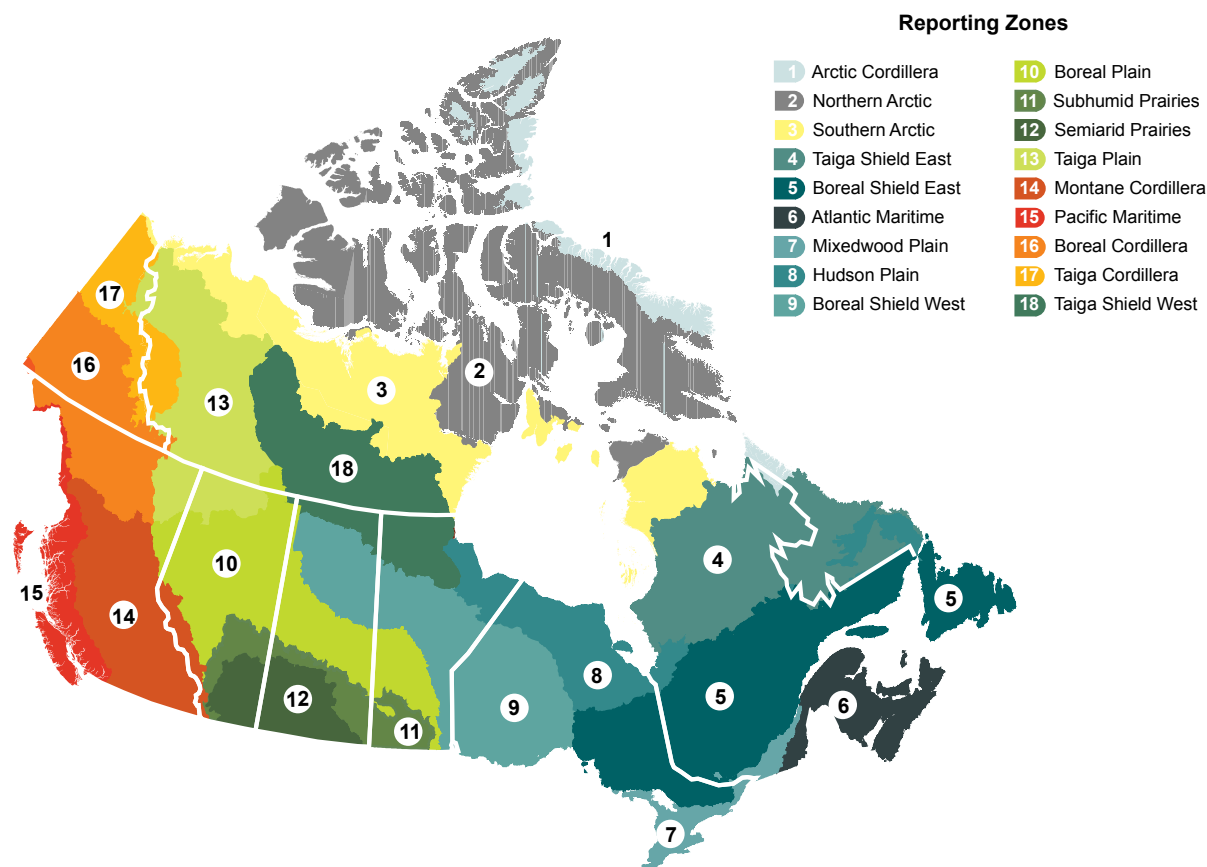
NO = Not occurring.

As a consequence of the land categorization scheme, some land-use transitions cannot occur—for example, forest conversion to agricultural grassland, since these by definition exclude areas where forests can grow naturally. Note that in theory the opposite can happen (i.e. grassland conversion to forest), although the direct human-induced conversion of agricultural grassland to forest has not been observed. Since grassland is defined as “native,” creation of

grassland is mostly not occurring.

Table 7–3 illustrates the land-use areas (diagonal cells) and cumulative land-use change areas (non-diagonal cells) in 2009. Cumulative land-use change areas are the total land areas converted over the past 20 years (10 years for reservoirs). The grassland diagonal cell refers to the total area of agricultural grassland, whereas grassland converted

Figure 7–1 Reporting Zones for LULUCF Estimates



to settlements refers to land conversion of un-managed tundra to settlements in northern Canada. Column totals equal the total land area as reported in the CRF for each category.

The MARS land monitoring system includes the conversion of un-managed forests and grassland to other land categories. Un-managed land converted to any use always becomes “managed”; once land has become managed, it does not revert to “un-managed” status, even if management practices are discontinued. Parks and protected areas are included in managed lands.

With a few exceptions (e.g. emissions due to liming), the LULUCF estimates as reported in the CRF tables are spatially attached to “reporting zones” (Figure 7–1). These reporting zones are essentially the same as Canada’s terrestrial ecozones (Marshall and Shut 1999), with three exceptions: the Boreal Shield and Taiga Shield ecozones are split into their east and west components to form four reporting zones; and the Prairies ecozone is divided into a semi-arid and a subhumid component. Estimates are reported for 17 of the 18 reporting zones, leaving out the northernmost ecozone of Canada: the Arctic Cordillera, where no direct human-induced GHG emissions and removals are detected for this sector. More details on the spatial estimation and reporting framework can be found in Annex 3.4.

The areas reported in the CRF tables represent those used for annual estimate development, but not always the total land area under a land category or subcategory in a specific inventory year. Hence areas of land converted to wetlands (reservoirs) represent a fraction of total reservoir areas (those flooded for 10 years or less), not the total area of reservoirs in Canada.

Similarly, the areas of land conversion reported in the CRF tables refer to the cumulative total land area converted over the last 20 years (10 years for reservoirs) and should not be confused with annual rates of land-use change. The trends observed in the land conversion categories of the CRF (e.g. land converted to forest land, land converted to cropland) result from the balance between land area newly converted to a category and the transfer of lands converted more than 20 years ago (10 years for reservoirs) into the “land remaining land” categories.

## 7.3. Forest Land

Forest and other wooded lands cover 389 million hectares (Mha) of Canadian territory; forest lands alone occupy 348 Mha (NRCan 2010). Managed forests, those under direct human influence, extend over 229 Mha, or 66% of all forests. Four reporting zones (Boreal Shield East, Montane Cordillera, Boreal Plains and Boreal Shield West) account for 68% of managed forests (Table 7–4).

In 2009, the net GHG balance of managed forest land amounted to removals of 17 Mt (Table 7–1 and CRF Table 5). This estimate includes net emissions and removals of CO<sub>2</sub>, as well as N<sub>2</sub>O, CO, and CH<sub>4</sub> emissions from wildfires in managed forests. For the purpose of UNFCCC reporting, managed forest lands are divided into forest land remaining forest land (229 Mha, net removals of 16 Mt) and land converted to forest land (0.10 Mha, net removals of 0.8 Mt) in 2009.

The managed forest GHG balance is not spatially homogeneous. In 2009, managed forests in the Montane Cordillera and Pacific Maritime reporting zones are two large net sources of GHG, while those in the Taiga Plains, Boreal Shield East, Boreal Shield West and Boreal Cordillera are four net sinks (Table 7–4). Note that the spatial distribution of emissions and removals is influenced by the occurrence and location of disturbances and would therefore not necessarily be constant in successive years.

### 7.3.1. Forest Land Remaining Forest Land

#### 7.3.1.1. Methodological Issues

Vegetation absorbs CO<sub>2</sub> from the atmosphere through photosynthesis, and some of this carbon is sequestered in standing vegetation (biomass), dead organic matter, and soils. CO<sub>2</sub> is returned to the atmosphere by vegetation respiration and the decay, through heterotrophic respiration, of organic matter. The natural CO<sub>2</sub> exchanges between the atmosphere and biota are large fluxes, globally recycling on the order of one seventh of the total atmospheric CO<sub>2</sub> content annually. These large gross fluxes result from the accumulation of minute processes dispersed over vast land areas. Only a small fraction of the carbon (C) taken up by photosynthesis accumulates in ecosystem C pools after all respiratory and disturbance C losses (Stinson et al. 2011).

Table 7–4 GHG Balance of Managed Forests by Reporting Zone, 2009<sup>1</sup>

Reporting Zone Number	Reporting Zone Name	Managed Forest Area (kha)	Net GHG Balance (Mt CO <sub>2</sub> eq)
1	Arctic Cordillera	–	NA
2	Northern Arctic	–	NA
3	Southern Arctic	–	NA
4	Taiga Shield East	1 100	1.8
5	Boreal Shield East	55 600	-34
6	Atlantic Maritime	15 400	-3.1
7	Mixedwood Plains	2 700	-7.8
8	Hudson Plains	302	-0.54
9	Boreal Shield West	28 800	-26
10	Boreal Plains	36 100	-5.2
11	Subhumid Prairies	1 840	-1.4
12	Semiarid Prairies	18	~0
13	Taiga Plains	20 000	-31
14	Montane Cordillera	35 400	100
15	Pacific Maritime	13 200	7
16	Boreal Cordillera	16 600	-16
17	Taiga Cordillera	412	~0
18	Taiga Shield West	1 830	-3

Notes:

1. Negative sign indicates removal of CO<sub>2</sub> from the atmosphere.

NA = Not applicable.

Human interactions with the land can directly alter the size and rate of these natural exchanges of GHGs, in both the immediate and long term. Land-use change and land-use practices in the past still affect current GHG fluxes to and from the terrestrial biosphere. This long-term effect is a unique characteristic of the LULUCF Sector, which makes it very distinct from other sectors, such as Energy.

While the intended focus is on anthropogenic impacts on the GHG balance, it is recognized that separating human from natural effects in the LULUCF Sector poses a unique challenge. Humans manipulate biological processes in a myriad of ways and intensities. What we observe is typically the outcome of these various manipulations and their combined interactions with an equally varied biophysical environment, including natural disturbances. Untangling the various cause-and-effect relationships in the long term and short term is still the object of complex scientific inquiries.

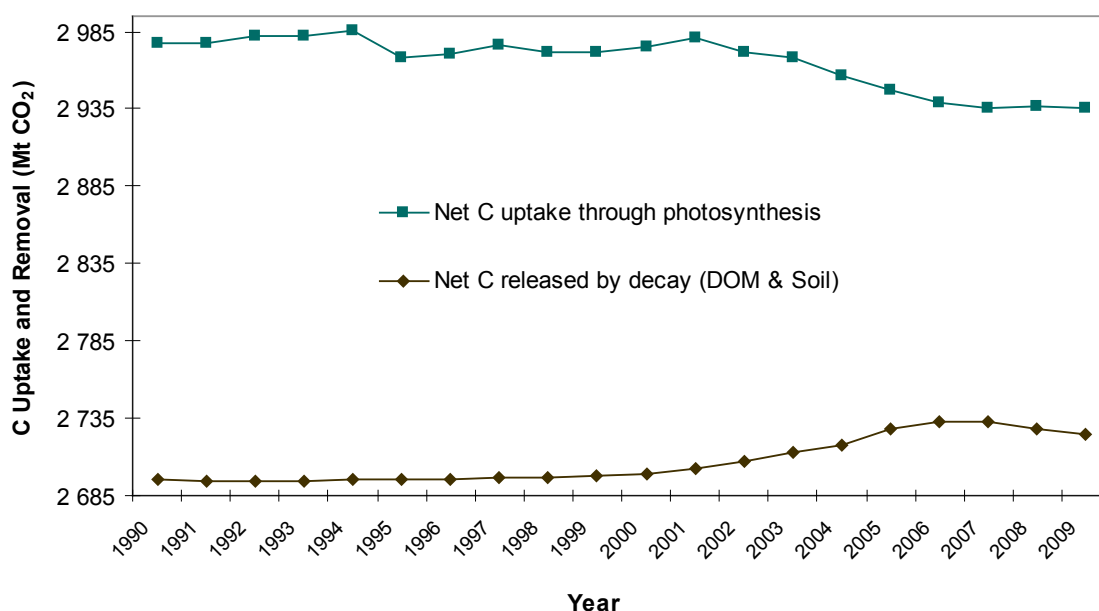
Canada emphasizes that while all efforts are made to provide IPCC-compliant GHG estimates, such estimates may not truly account for direct human effects, nor accurately reflect where and when emissions occur (Stinson et al. 2011).

Canada applies a Tier 3 methodology for estimating GHG emissions and removals in managed forests. Canada's

National Forest Carbon Monitoring, Accounting and Reporting System (NFCMARS – Kurz and Apps 2006) includes a model-based approach (Carbon Budget Model of the Canadian Forest Sector, CBM-CFS3 – Kull et al. 2006; Kurz et al. 2009). This model integrates forest inventory data and yield curves with spatially referenced activity data on forest management and natural disturbances (fires, insect infestations) to estimate forest carbon stocks, stock changes and CO<sub>2</sub> emissions and removals. The model uses regional ecological and climate parameters to simulate carbon transfers among pools, to the forest product sector and to the atmosphere. The conceptual approach remains that recommended by the IPCC (2003), in which net removals or emissions are calculated as the difference between CO<sub>2</sub> uptake by growing trees and emissions from forest management activities (harvesting), heterotrophic respiration and natural disturbances. The interested reader will find additional information on estimation methodology in Annex 3.4.

Carbon stock changes in managed forests are reported in CRF Table 5A, by reporting zone. For any given pool, carbon stock changes include not only exchanges of GHG with the atmosphere, but also the carbon transfers to and from pools, for example its transfer from living biomass to dead organic matter upon stand mortality. Therefore individual carbon stock changes give no indication of the

Figure 7–2 Large Annual Carbon Fluxes to and from the Atmosphere in Managed Forests, 1990–2009: Net Carbon Uptake (or Net Primary Production) and Release Due to Decay (From Heterotrophic Respiration)



net fluxes between carbon pools in managed forests and the atmosphere. The largest carbon fluxes to and from managed forests consist of net carbon uptake by growing trees and its release due to the decay of organic matter (–2900 and 2700 Mt, respectively, in 2009; see Figure 7–2). The upward trend in dead organic matter (DOM) decay since the year 2000 reflects the long-term, growing effect of past disturbances, especially insect epidemics that have left substantial quantities of decaying DOM. Over the last three years, insect epidemics have affected a total of over 18 Mha<sup>3</sup> of managed forests, with 84% being located in the Montane Cordillera reporting zone and corresponding to the epidemics of Mountain Pine Beetle. In contrast, much of the interannual variability of the GHG budget of managed forests hinges on the occurrence and severity of fires. During the 1990–2009 period, annual wildfire emissions fluctuated between 11 and 260 Mt. The consumption of dead organic matter by fires accounts for 76% of immediate emissions; much biomass is killed by forest fires and is thus transferred to the dead organic matter pool, but is not immediately burned. Hence, a large amount of the actual fuel load consists of dead wood and litter on the forest floor. On average, 8% of immediate fire emissions in CO<sub>2</sub> equivalents are in the form of CO, 7% as CH<sub>4</sub>, and 4% in the form of N<sub>2</sub>O.

In order to avoid double counting, estimates of C stock changes in CRF Table 5A exclude carbon emissions as CO<sub>2</sub>, CH<sub>4</sub> and CO due to biomass burning, which are reported in Table 5(V). Emissions and removals are automatically tallied in CRF Table 5.

In keeping with the current IPCC default methodology, CO<sub>2</sub> emissions from forest management activities comprise all the carbon contained in harvested roundwood and harvest residues. All carbon transferred out of managed forests as wood products is deemed an immediate emission. Under this approach, in 2009 the transfer of carbon from forests to harvested wood products (HWP) from forest management activities accounts for emissions of 114 Mt, a decrease of 23% compared to 1990 values. Three alternative approaches—atmospheric flow, production and stock change—have been preliminarily evaluated in Canada to attempt to correctly account for delayed emissions due to long-term carbon storage in HWPs. These approaches account for the carbon stored in HWPs and emissions from the eventual decay of products derived from domestically harvested, imported (stock change, atmospheric flow) or exported (production) wood in the current and previous years; they are therefore more spatially and temporally realistic than the current default, which does not account for emissions from HWPs where or when they actually occur. They differ with respect to their allocation of emissions and removals. A breakdown and brief discussion

3 May include areas repeatedly infested, e.g., a hectare infested in three successive years is counted as three hectares towards the 18 Mha.

of each of the accounting approaches, along with implications for Canada, are contained in Annex 3.4.

### 7.3.1.2. Uncertainties and Time-Series Consistency

#### Uncertainty Estimates

Conducting uncertainty analysis is a significant challenge in complex modelling frameworks such as the one used for modelling forest carbon dynamics, with multiple interactions and dependencies between current and historical events, and among the very large number of model inputs and parameters, required to simulate carbon dynamics.

Numerical techniques were used to quantify uncertainties about the outputs of the CBM-CFS3 (Metsaranta et al. 2011). Careful consideration was given to the identification and representation of uncertainty sources in the analysis (see Annex A3.4.2.4 for additional information). Modelling of the entire managed forests of Canada is not done as a single run, but in separate “project runs” whose output is subsequently assembled. For each “project,” 100 Monte Carlo runs were conducted using the base input data for the 2011 submission (covering the entire 1990–2009 time series). Confidence intervals were obtained for each inventory year, by randomly sampling 10 000 combinations of all the project runs for that year. Separate uncertainty estimates were produced for each gas.

The uncertainties about annual estimates are expressed as a 95% confidence interval, bound by 2.5<sup>th</sup> and 97.5<sup>th</sup> percentiles of the Monte Carlo run outputs (see Table 7–5, Table 7–6 and Table 7–7 for CO<sub>2</sub> fluxes and CH<sub>4</sub> and N<sub>2</sub>O emissions, respectively).

Over the entire time series, the range of the 95% confidence interval about the median CO<sub>2</sub> estimates averaged 73 Mt. Expressing this range in relative terms (as % of the estimate) can be misleading, as the relative uncertainty will be largest when the net CO<sub>2</sub> balance is closest to neutrality, and increasingly smaller as the net flux departs from neutrality. This does not represent varying uncertainty levels; it is an artefact of the combination of large fluxes cancelling each other, while their respective uncertainties do not. Also note the very small contribution of non-CO<sub>2</sub> emissions to total uncertainty.

More information is provided in Annex A3.4.2.4 on the general approach used to conduct this analysis.

Table 7–5 Estimates of the net Annual CO<sub>2</sub> Fluxes for Forest Land Remaining Forest land, 1990–2009, with 2.5<sup>th</sup> and 97.5<sup>th</sup> Percentiles

Inventory Year	Net CO <sub>2</sub> Flux (Kt)	2.5 <sup>th</sup> Percentile (Kt)	97.5 <sup>th</sup> Percentile (Kt)
1990	-97 000	-240 000	-96 000
1991	-74 000	-210 000	-63 000
1992	-110 000	-260 000	-110 000
1993	-51 000	-190 000	-37 000
1994	-50 000	-190 000	-36 000
1995	140 000	39 000	190 000
1996	-51 000	-180 000	-35 000
1997	-88 000	-230 000	-73 000
1998	79 000	-31 000	140 000
1999	-15 000	-130 000	9 900
2000	-75 000	-220 000	-56 000
2001	-72 000	-200 000	-48 000
2002	60 000	-47 000	120 000
2003	21 000	-76 000	52 000
2004	74 000	-32 000	110 000
2005	38 000	-74 000	63 000
2006	48 000	-43 000	75 000
2007	35 000	-65 000	64 000
2008	-28 000	-140 000	-3 500
2009	-25 000	-120 000	-780

Table 7–6 Estimates of the Annual CH<sub>4</sub> Emissions from Forest Land Remaining Forest Land, 1990–2009, with 2.5<sup>th</sup> and 97.5<sup>th</sup> Percentiles

Inventory Year	Net CH <sub>4</sub> Flux (Kt CO <sub>2</sub> eq)	2.5 <sup>th</sup> Percentile (Kt)	97.5 <sup>th</sup> Percentile (Kt)
1990	3 000	2 700	4 200
1991	5 000	4 200	8 300
1992	1 500	1 300	2 200
1993	6 000	5 000	9 000
1994	5 600	4 900	7 600
1995	19 000	17 000	27 000
1996	4 400	3 800	7 900
1997	1 600	1 400	2 600
1998	15 000	13 000	23 000
1999	6 000	5 700	9 900
2000	1 300	1 100	1 900
2001	2 800	2 300	4 900
2002	11 000	9 800	19 000
2003	8 500	7 800	14 000
2004	9 500	8 000	14 000
2005	5 500	4 700	8 400
2006	6 600	6 000	10 000
2007	6 800	5 800	10 000
2008	4 100	3 500	5 700
2009	5 600	4 800	9 300



Table 7–7 Estimates of the Annual N<sub>2</sub>O Emissions from Forest Land Remaining Forest land, 1990–2009, with 2.5<sup>th</sup> and 97.5<sup>th</sup> Percentiles

Inventory Year	Net N <sub>2</sub> O Flux (Kt CO <sub>2</sub> eq)	2.5 <sup>th</sup> Percentile (Kt)	97.5 <sup>th</sup> Percentile (Kt)
1990	1 800	2 700	4 200
1991	3 100	4 200	8 300
1992	950	1 300	2 200
1993	3 700	5 000	9 000
1994	3 500	4 900	7 800
1995	12 000	1 700	27 300
1996	2 700	3 800	7 900
1997	980	1 400	2 600
1998	9 200	1 300	23 000
1999	3 700	5 700	9 900
2000	800	1 100	1 900
2001	1 700	2 300	4 900
2002	7 000	9 800	19 000
2003	5 400	7 800	14 000
2004	5 900	8 000	14 000
2005	3 400	4 700	8 400
2006	4 100	6 000	10 000
2007	4 200	5 800	10 000
2008	2 500	3 500	5 700
2009	3 500	4 800	9 300

The uncertainty sources included in this analysis were essentially errors about input data and model parameters, as opposed to the model structure itself. Given the nature of these sources, the results of the uncertainty analysis are better understood as expressions of precision than estimates of accuracy. These results will be used, among other factors, to analyze and prioritize improvements in the estimation of anthropogenic emissions and removals in this category.

### Time Series Consistency

All estimates have been developed in a consistent manner, but some sources of activity data do not provide full coverage for the reporting period. Estimates for wildfire areas in 2004–2009 were derived from a composite of real-time, remotely sensed imagery and monitoring data collected by provincial resource management agencies produced by the Canadian Wildland Fire Information System.<sup>4</sup> Estimates for 1990–2003 were derived solely from the Canadian Forest Service (CFS) large fire database which comprises information from provincial resource management agencies.

The forest inventory data incorporated in the analyses were not all collected in the same year across the country; Annex 3.4 explains how forest inventory data from various sources were processed to provide complete, coherent and consistent forest data for 1990.

### 7.3.1.3. QA/QC and Verification

Tier 2 quality control (QC) checks (White and Dymond 2008; Dymond 2008), specifically address estimate development in the Forest Land category. Systematic and documented quality assurance / quality control (QA/QC) procedures are performed in four areas: workflow checks (manual), model checks (automated), benchmark checks (manual) and external reviews. Check results are systematically documented; an issue logging system identifies each issue and facilitates tracking and managing its resolution.

Environment Canada, while maintaining its own QA/QC procedures for estimates developed internally (refer to Annex 6), has implemented category-specific Tier 2 checks for estimates obtained from partners, as well as for all estimates and activity data contained in the LULUCF geodatabase and entered into the CRF reporter. These procedures and their outcome are fully documented in the centralized archives.

### 7.3.1.4. Recalculations

Recalculations include methodological changes and updated activity data. The largest recalculations occur in the following categories:

*Immediate Emissions from Wildfire:* There were downward recalculations of immediate CO<sub>2</sub> emissions for the years 1990–2005, which ranged from less than 1 Mt to 23 Mt CO<sub>2</sub> (the 23 Mt CO<sub>2</sub> recalculation represents a 10% change for 1995, a large fire year) and upward recalculations for the years 2006–2008 (from 3 to 10 Mt CO<sub>2</sub>). Recalculations of immediate emissions from fire occur as a result of the combined effects of methodological changes (which were applied in a consistent way for the whole time series) and updated wildfire area data (updated for 2006–2008). When examined on a forest area basis, there were downward recalculations of immediate CO<sub>2</sub> emissions for the years 1990–2007; these ranged from less than 0.002 kt CO<sub>2</sub>/ha burned to 0.025 kt CO<sub>2</sub>/ha burned. These recalculations were largely the result of recalibration of C transfer parameters in fire disturbance matrices, which incorporates new fire weather data into the averaging process used to derive these parameters. The effect of this methodological recal-

4 [http://cwfis.cfs.nrcan.gc.ca/en\\_CA/index](http://cwfis.cfs.nrcan.gc.ca/en_CA/index)



bration is offset by a change in area burned for 2006–2008 which caused upward recalculations during these years. Wildfire area data updates involved more detailed mapping of areas affected by fires. These updates led to significant increases in areas burned, especially in 2006 (by 100 kha or 18%) and 2007 (117 kha or 21%).

*Annual Processes:* Total recalculations within this category are due to the combined effects of recalculations in biomass CO<sub>2</sub> uptake and emissions from DOM decay. For the 1990–2008 time period, there was an overall increase in CO<sub>2</sub> uptake by biomass (by on average 30 Mt ha/year, which on a forest area basis accounts for an average increase of 0.0001 kt CO<sub>2</sub> ha/year or 1%) and an increase in DOM decay (by on average 32 Mt ha/year or 0.0001 kt CO<sub>2</sub> ha/year or 2%). The main cause of the increase in C uptake by biomass was the refinement of the algorithm used to simulate the growth of young forest stands. For the 2009 submission, the algorithm used to simulate the growth of very young stands was modified to reduce discontinuities as these stands grew to merchantable volume and were assigned standard growth curves. This algorithm was further refined with the use of linear and non-linear smoothing techniques. The dominant contribution to the increase in emissions due to DOM decay is due to recalibration of C transfer parameters for fire (as explained above), which leave more DOM to decay post-fire.

*Harvesting:* Recalculations in emissions due to harvesting are less than 2 Mt CO<sub>2</sub> eq (1%) for all years except 2006 and 2007, where emissions dropped by 3 and 17 Mt CO<sub>2</sub> eq, respectively (2% and 10%). For these years there were substantial updates to harvest statistics from provinces and territories.

In the earlier part of the time series many of the recalculations are due to method changes, while in the later part of the time series many recalculations are due to activity data updates (forest management information and wildfire area). The total magnitude of the combined recalculations range from -19 Mt CO<sub>2</sub> (2004) to 27 Mt CO<sub>2</sub> (2001).

### 7.3.1.5. Planned Improvements

Long-term planned improvements include enhancing the quality of forest inventory data, refining model parameters and expanding model scope. A more systematic approach is being considered for implementing improvements, to reduce the frequency of recalculations.

## 7.3.2. Land Converted to Forest Land

### 7.3.2.1. Category Description

This category includes all lands converted to forest land through direct human activity. Post-harvest tree planting is not included, nor is abandoned farmland where natural vegetation is allowed to establish; hence, the category more precisely refers to forest establishment where the previous land use was not forest (typically, abandoned farmland).

The total cumulative area of land converted to forest land declined from 170 kha in 1990 to 98 kha in 2009. The trend reflects the gradual transfer of lands afforested more than 20 years ago to the forest land remaining forest land category, and a dearth of recent data on current rates of forest establishment. Only 7% of all farmland converted to forest land over the last 20 years is located in the Prairies (Boreal Shield West, Boreal Plains and Subhumid Prairies reporting zones); 88% is in eastern Canada (Atlantic Maritime, Mixedwood Plains and Boreal Shield East reporting zones). However, caution must be exerted in regard to analysis of afforestation trends due to the absence of a reliable and comprehensive data source.

Net removals consequently declined throughout the period, from 1.0 Mt in 1990 to 0.8 Mt in 2009. Net carbon accumulation largely occurs in biomass (192 Gg C in 2009 – CRF table 5A); soil carbon sequestration is negligible and will remain so because this category is restricted to plantations that are younger than 20 years. For the same reason, and considering the relatively low net increment of planted trees in the early years, the subcategory as a whole is not expected to contribute significantly to the net greenhouse gas balance of forest lands.

### 7.3.2.2. Methodological Issues

Until a few years ago, national compilations of afforestation records in Canada were not available. The Feasibility Assessment of Afforestation for Carbon Sequestration (FAACS) initiative collected and compiled afforestation records for 1990–2002 (NRCan 2005a). In this period, softwood plantations, especially spruce and pine, accounted for 90% of the area planted. Activities for 1970–1989 and 2003–2008 were estimated based on activity rates observed in the FAACS data, complemented with information from the Forest 2020 Plantation Demonstration

Assessment (NRCan 2005b). No new afforestation activity data were identified for the 2009 inventory year.

GHG emissions and removals on lands newly converted to forests were estimated using CBM-CFS3, as described in Annex 3.4. Changes in soil carbon stocks are highly uncertain because of difficulties in locating data about the carbon stocks prior to plantation. It was assumed that the ecosystem would generally accumulate soil carbon at a slow rate; the limited time frame of this analysis and the scale of the activity relative to other land use and land-use change activities suggest that the impact of this uncertainty, if any, is minimal.

### 7.3.2.3. Uncertainties and Time-Series Consistency

It was not possible, owing to data and resource limitations, to develop formal uncertainty estimates for this category.

### 7.3.2.4. QA/QC and Verification

Tier 2 QC checks (Dymond 2008) specifically address estimate development in the Forest Land category. Environment Canada, while maintaining its own QA/QC procedures for estimates developed internally (refer to Annex 6), has implemented specific ones for estimates obtained from partners, as well as for all estimates and activity data contained in the LULUCF geodatabase and entered into the CRF reporter.

### 7.3.2.5. Recalculations

Recalculations include both updated activity data and methodological changes. Area estimates of land converted to forest land were reviewed, including a reanalysis and error correction of pre-2002 activity data, as well as the deletion of unsupported post-2002 afforestation activity derived from the Feasibility Assessment of Afforestation for Carbon Sequestration (FAACS). The lower CO<sub>2</sub> removals due to the smaller areas of land converted to forest land are partly offset by changes within the CBM-CFS3 (see Section 7.3.1.4, Recalculations), which will result in higher per hectare carbon removals.

### 7.3.2.6. Planned Improvements

There is currently limited access to information on afforestation activity. Efforts are underway to obtain data on afforestation activities in recent years from provincial and

territorial resource management agencies and to develop uncertainty estimates for areas of afforestation.

## 7.4. Cropland

Cropland covers approximately 47 Mha of the Canadian territory. In 2009, the net GHG balance in the Cropland category amounted to removals of 6.9 Mt CO<sub>2</sub> eq (Table 7-1 and CRF Table 5). For the purpose of reporting under the Convention (UNFCCC), Cropland is divided into cropland remaining cropland (net removals of about 12 Mt CO<sub>2</sub> eq in 2009) and land converted to cropland, due to either forest or grassland conversion (net emissions of about 5.5 Mt CO<sub>2</sub> eq and 0.009 Mt CO<sub>2</sub> eq respectively in 2009). The estimates in land “converted to cropland” include net emissions and removals of CO<sub>2</sub>, as well as N<sub>2</sub>O, CO and CH<sub>4</sub> emissions.

### 7.4.1. Cropland Remaining Cropland

Cultivated agricultural land in Canada includes areas of field crops, summerfallow, hayland, and tame or seeded pasture. Cropland is found mainly in the nine southernmost reporting zones. About 83% of Canada’s cropland is in the interior plains of western Canada, made up of the Semi-arid Prairies, the Subhumid Prairies and the Boreal Plains reporting zones.

Cropland remaining cropland includes CO<sub>2</sub> emissions/removals in mineral soils, CO<sub>2</sub> emissions from agricultural lime application and cultivation of organic soils, and CO<sub>2</sub> emissions/removals resulting from changes in woody biomass from specialty crops. An enhanced Tier 2 approach is used for estimating CO<sub>2</sub> emissions from and removals by mineral soils triggered by changes in land management practices. Table 7-8 summarizes the trend in emissions and removals for these categories.

#### 7.4.1.1. CO<sub>2</sub> Emissions and Removals in Mineral Soils

Mineral soils constitute the majority of cropland areas (> 99%). The amount of organic carbon retained in these soils is a function of primary production and rate of decomposition of soil organic carbon (SOC). Cultivation and management practices can lead to an increase or decrease in the organic carbon stored in soils. This change in SOC results in a CO<sub>2</sub> emission to or removal from the atmosphere.

Table 7–8 Base and Recent Year Emissions and Removals Associated with Various Land Management Changes on Cropland Remaining Cropland

Categories	Management Change (LMC)	Emissions/Removals (Kt CO <sub>2</sub> ) <sup>1</sup>				
		1990	2006	2007	2008	2009
<b>Total Cropland Remaining Cropland</b>		<b>-1 500</b>	<b>-11 000</b>	<b>-11 000</b>	<b>-12 000</b>	<b>-12 000</b>
<i>Cultivation of Histosols</i>		300	300	300	300	300
<i>Liming</i>		200	290	290	290	290
<i>Perennial Woody Crops</i>		50	20	20	10	10
<i>Total Mineral Soils</i>		-2 000	-11 000	-12 000	-12 000	-13 000
Change in Crop Mixture	Increase in Perennial	-1 200	-4 900	-5 200	-5 500	-5 800
	Increase in Annual	3 500	3 700	3 600	3 600	3 600
Change in Tillage	Conventional to Reduced Tillage	-850	-840	-820	-790	-770
	Conventional to No-till	-530	-3 800	-3 900	-4 100	-4 300
	Other	NO	-640	-710	-780	-840
Change in Summerfallow (SF)	Increase in SF	1 700	1 200	1 200	1 200	1 200
	Decrease in SF	-4 800	-7 700	-7 700	-7 800	-7 900
Land Conversion—Residual Emissions <sup>2</sup>		170	1 700	1 700	1 700	1 800

Notes:

1. Negative sign indicates removal of CO<sub>2</sub> from the atmosphere.2. Net residual CO<sub>2</sub> emissions from the conversion of forest land and grassland to cropland that occurred more than 20 years prior to the inventory year, including emissions from the decay of woody biomass and DOM.

In 1990, changes in mineral soil management amounted to a net CO<sub>2</sub> removal of about 2 Mt CO<sub>2</sub> eq (Table 7–8). This net sink steadily increased to about 13 Mt CO<sub>2</sub> eq in 2009, reflecting continuous efforts in reducing summerfallow and increasing conservation tillage (Campbell et al. 1996; Janzen et al. 1998; McConkey et al. 2003). The area of summerfallow declined by more than 50% over the 1990–2009 period, resulting in a net sink that increased from 3.1 Mt CO<sub>2</sub> eq in 1990 to 6.7 Mt CO<sub>2</sub> eq in 2009. The increase in net sink due to the adoption of conservation tillage practices (from 1.4 Mt CO<sub>2</sub> eq in 1990 to 5.9 Mt CO<sub>2</sub> eq in 2009) is substantiated by a net total increase of over 10 Mha in areas under no-till and reduced tillage over the 1990–2009 period. The net change in crop mixture has had a much more modest impact, from a source of 2.3 Mt CO<sub>2</sub> eq in 1990 to a sink of 2.2 Mt CO<sub>2</sub> eq in 2009.

The net increase in sink from changes in management practices over time was partially offset by an increase since 1990 in net residual CO<sub>2</sub> emissions from the decay of dead organic matter and SOC on land converted to cropland more than 20 years prior to the inventory year (emissions from land converted for less than 20 years are included under land converted to cropland). The increase since 1990 in these residual emissions is due to an accounting artefact. Since forest conversion monitoring goes back only to 1970, post-20-year residual emissions in 1990 only

accounted for the land converted in 1970. Residual emissions display an apparent increase because the temporal coverage increases with each inventory year. In the CRF tables, these emissions are split among the dead organic matter and soil pools.

## Methodological Issues

Following the IPCC Good Practice Guidance for LULUCF (IPCC 2003), the premise is that the changes in SOC are driven by changes in soil management practices. Where no change in management is detected, it is assumed that mineral soils are neither sequestering nor losing carbon.

VandenBygaart et al. (2003) compiled published data from long-term studies in Canada to assess the effect of agricultural management on SOC. This compendium provided the basis for selecting the key management practices and management changes likely to cause changes in soil carbon stocks. The availability of activity data (time series of management practices) from the *Census of Agriculture* was also taken into account. A number of management practices are known to increase SOC in cultivated cropland. They include a reduction in tillage intensity, intensification of cropping systems, adoption of yield-promoting practices and re-establishment of perennial vegetation (Janzen et al. 1997; Bruce et al. 1999). Other land management

changes, such as changes in irrigation, manure application and fertilization, are also known to have positive impacts on SOC. Lack of activity data for these land management changes (LMCs) associated with specific crops prevented their inclusion in the inventory at this time. Estimates of CO<sub>2</sub> changes in mineral soils were derived from the following LMCs:

- change in the proportion of annual and perennial crops;
- change in tillage practices; and
- change in area of summerfallow.

Carbon emissions and removals were estimated by applying country-specific carbon emission and removal factors multiplied by the relevant area of land that underwent a management change. Calculations were performed at a high degree of spatial disaggregation, namely by Soil Landscapes of Canada (SLC) polygons (see Annex 3.4.1). The carbon emission/removal factors represent the rate of SOC change per year and per unit area that underwent an LMC. The annual CO<sub>2</sub> emissions/removals by mineral soils undergoing a specific LMC are expressed as:

Equation 7-1:

$$\Delta C = F \times A$$

where:

$\Delta C$	=	change in soil carbon stock (Mg C)
$F$	=	average change in SOC subject to LMC (Mg C/ha)
$A$	=	area of LMC (ha)

In reality, the impact of LMC on SOC varies with initial conditions. The most accurate estimate of soil carbon stock change would therefore be derived by individually considering the cumulative effects of the long-term management history of each piece of land or farm field. Limits are imposed by the availability of activity data within the modelling framework. At this point, the inventory relies extensively on the *Census of Agriculture* for estimates of areas of LMC (i.e. changes in tillage, types of crop and fallow). The agricultural area of LMC was determined individually for 3264 SLC polygons having agricultural activities, each one having an area in the order of 1000–1 000 000 ha. This is the finest possible resolution of activity data, given the limitations imposed by confidentiality requirements of

census data. The census provides information about the area of each practice for each census year, so only the net area of change for each land management practice can be estimated. Estimates of these LMCs are as close to gross area of LMC as is feasible for regional or national analyses.

The validity of LMC estimates using census data relies on two key assumptions: additivity and reversibility of carbon factors. Additivity assumes that the combined effects of different LMCs or LMCs at different times would be the same as the sum of the effect of each individual LMC. Reversibility is the assumption that the carbon effects of an LMC in one direction (e.g. converting annual crops to perennial crops) is the opposite of the carbon effects of the LMC in the opposite direction (e.g. converting perennial crops to annual crops).

The various carbon factors associated with each particular situation (in both space and time) were derived using the CENTURY model (Version 4.0) by comparing output for scenarios “with” and “without” the management change in question. In specific instances, empirical data were used to complement the results of the CENTURY runs.

A more detailed description of methodologies for determining carbon factors and other key parameters can be found in Annex 3.4.

## Uncertainties and Time-Series Consistency

Uncertainty was estimated analytically with a Tier 1 approach. The uncertainties associated with estimates of CO<sub>2</sub> emissions or removals involve estimates of uncertainties for area and carbon factors of management changes for fallow, tillage and annual/perennial crops (McConkey et al. 2007).

The uncertainty about the area in a management practice for an ecodistrict varied inversely with the relative proportion it occupied of the total area of agricultural land in that ecodistrict. The relative uncertainty of the area of management practice (expressed as standard deviation of an assumed normal population) decreased from 10% to 1.25% of the area as the relative area of that practice increased.<sup>5</sup>

The uncertainties associated with carbon change factors for fallow, tillage and annual/perennial crops were partitioned in two main sources: 1) process uncertainty in

<sup>5</sup> T. Huffman, Agriculture and Agri-Food Canada, personal communication to Brian McConkey, 2007.

carbon change due to inaccuracies in predicting carbon change even if the situation of management practice was defined perfectly, and 2) situational uncertainty in carbon change due to variation in the location or timing of the management practice. More details about estimating process and situational uncertainties are presented in Annex 3.4. Uncertainty estimates associated with emissions/removals of CO<sub>2</sub> from mineral soils were developed by McConkey et al. (2007), who reported uncertainty values at  $\pm 19\%$  for the level and  $\pm 27\%$  for the trend. These uncertainty estimates have not been updated since 2009, but should still be applicable because there has been no change in the inventory method or activity data over the last two submissions.

Consistency in the CO<sub>2</sub> estimates is ensured through the use of the same methodology for the entire time series of estimates (1990–2009).

### QA/QC and Verification

Tier 1 QC checks, implemented by Agriculture and Agri-Food Canada (AAFC), specifically address estimate development in the cropland remaining cropland subcategory. Environment Canada, while maintaining its own QA/QC procedures for estimates developed internally (see Annex 6), has implemented additional QC checks for estimates obtained from partners, as well as for all estimates and activity data contained in its LULUCF geodatabase and entered into the CRF reporter. In addition, the activity data, methodologies and changes are documented and archived in both paper and electronic form.

Carbon change factors for LMCs used in the inventory were compared with empirical coefficients in Vanden-Bygaart et al. (2008). The comparison showed that empirical data on changes in SOC in response to no tillage were highly variable, particularly for eastern Canada. Nonetheless, the modelled factors were still within the range derived from the empirical data. For the switch from annual to perennial cropping, the mean empirical factor was 0.59 Mg C/ha per year, and this compared favourably with the range of 0.46–0.56 Mg C/ha per year in the modelled factors in western Canadian soil zones. For eastern Canada, only two empirical change factors were available, but they fell within the range of the modelled values (0.60–1.07 Mg C/ha per year empirical versus 0.74–0.77 Mg C/ha per year modelled). For conversion of crop fallow to continuous cropping, the modelled rate of carbon storage obtained (0.33 Mg C/ha per year) was more than twice the average

rate of  $0.15 \pm 0.06$  Mg C/ha per year derived from two independent assessments of the literature. This difference led to the decision to use empirically based factors for changes in summerfallow in the inventory. More details can be found in Annex 3.4.

In February 2009, Canada convened an international team of scientists and experts from Denmark, France, Japan, Sweden, the Russian Federation and the United States, to conduct a quality assurance assessment of the Canadian Agricultural Monitoring, Accounting and Reporting System (Can Ag-MARS). Some limitations of the current system were found with respect to activity data, which could possibly create some bias in the current carbon stock change estimates. In particular, the lack of a complete and consistent set of land-use data, and issues with the concept and application of pseudo-rotations, will be addressed in the next generation of Can Ag-MARS.

### Recalculations

Recalculations have been carried out on the 1990–2008 time series of estimates due mainly to the area update of forest land conversion to cropland. In addition, improvements in the inventory method for forest conversion to cropland for western Canada also resulted in recalculations of cropland emissions and removals indirectly through altering the spatial distribution of tillage practices, summerfallow and annual/perennial crops. As a result, there is a small increase in CO<sub>2</sub> removals by 62 kt in 1990 and 220 kt in 2008 for cropland remaining cropland.

### Planned Improvements

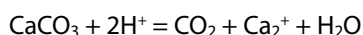
In the short term, work is ongoing to reduce uncertainties associated with the modelled carbon factors, through general improvements to factor methodologies, validation, and review of assumptions, where possible. Publication of scientific, peer-reviewed material is also underway. Improvements to the CENTURY model and the use of alternative models are also being explored, to improve the simulation of Canadian agricultural conditions. The quality of area statistics collected through the *Census of Agriculture* will be improved using land cover information.

#### 7.4.1.2. CO<sub>2</sub> Emissions from Lime Application

In eastern Canada, limestone and dolomite are often used for certain crops such as alfalfa to neutralize acidic soils, increase the availability of soil nutrients, in particular



phosphorus, reduce the toxicity of heavy metals, such as aluminium, and improve the crop growth environment. During this neutralization process, CO<sub>2</sub> is released in bicarbonate equilibrium reactions that take place in the soil:



The rate of release will vary with soil conditions and the compounds applied. In most cases where lime is applied, applications are repeated every few years. For the purposes of the inventory, it is assumed that the rate of lime addition is in near equilibrium with the rate of lime consumed from previous applications.

### Methodological Issues

Emissions associated with the use of lime were calculated from the amount and composition of the lime applied annually—specifically, the respective stoichiometric relationships that describe the breakdown of limestone and dolomite into CO<sub>2</sub> and other minerals. Methods and data sources are outlined in Annex 3.4.

### Uncertainties and Time-Series Consistency

The 95% confidence limits about data on the annual lime consumption in each province were estimated to be  $\pm 50\%$  (McConkey et al. 2007). This uncertainty was assumed to include the uncertainty about lime sales, uncertainty in proportion of dolomite to calcite, uncertainty of when lime sold is actually applied, and uncertainty in the timing of emissions from applied lime. The uncertainty in the emission factor was not considered because the chemical conversion is deemed complete, and the maximum value of the emission factor was used. The overall mean and uncertainties were estimated to be  $0.3 \pm 0.25$  Mt CO<sub>2</sub> eq for the level uncertainty and  $0.09 \pm 0.30$  Mt CO<sub>2</sub> eq for the trend uncertainty (McConkey et al. 2007).

The same methodology is used for the entire time series of emission estimates (1990–2009).

### QA/QC and Verification

This category has undergone Tier 1 QC checks (see Annex 6) in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

### Recalculations

There was no recalculation involved in emission estimates for this source category.

### Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this source.

#### 7.4.1.3. CO<sub>2</sub> Emissions from Cultivation of Organic Soils

### Category Description

In Canada, cultivated organic soils are defined as the conversion of organic soils to agriculture for annual crop production, normally accompanied by artificial drainage, cultivation and liming. Organic soils used for agricultural production in Canada include the Peaty Phase of Gleysolic soils, Fibrisols over 60 cm thick, and Mesisols and Humisols over 40 cm thick (AAFC 1998)

### Methodological Issues

The emissions from the cultivation of organic soils were calculated by multiplying the total area of cultivated histosols by the default emission factor of 5 Mg C/ha per year (IPCC 2006).

Areas of cultivated histosols are not provided by the *Census of Agriculture*; area estimates were based on the expert opinion of soil and crop specialists across Canada (Liang et al. 2004). The total area of cultivated organic soils in Canada (constant for the period 1990–2009) was estimated to be 16 kha or 0.03% of the cropland area.

### Uncertainties and Time-Series Consistency

The uncertainty associated with emissions from this source is due to the uncertainties from the area estimates for the cultivated histosols and the emission factor. The 95% confidence limits associated with the area estimate of cultivated histosols are assessed to be  $\pm 50\%$  (Hutchinson et al. 2007). The 95% confidence limits of the default emission factor are  $\pm 90\%$  (IPCC 2006). The overall mean and uncertainties associated with this source of emissions were estimated to be  $0.3 \pm 0.09$  Mt CO<sub>2</sub> eq for the level uncertainty and  $0 \pm 0.13$  Mt CO<sub>2</sub> eq for the trend uncertainty (McConkey et al. 2007).

The same methodology and emission factors are used for the entire time series of emission estimates (1990–2009).



## QA/QC and Verification

This category has undergone Tier 1 QC checks (see Annex 6) in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

## Recalculations

There was no recalculation involved in emission estimates for this source category.

## Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this source.

### 7.4.1.4. CO<sub>2</sub> Emissions and Removals in Woody Biomass

#### Category Description

Perennial woody biomass is found on cropland planted with vineyards, fruit orchards and Christmas trees. It also accumulates on abandoned cropland allowed to revert to natural vegetation. In the definitional framework adopted in Canada for LULUCF reporting, abandoned cropland is still considered “cropland” until there is evidence of a new land use; however, there is little information on the dynamics of cropland abandonment or recultivation. Owing to these data limitations, only vineyards, fruit orchards and Christmas trees are considered; for the time being changes in woody biomass from “abandoned cropland” on cropland remaining cropland are excluded.

#### Methodological Issues

Vineyards, fruit orchards and Christmas tree farms are intensively managed for sustained yields. Vineyards and fruit trees are pruned annually, and old plants are replaced on a rotating basis for disease prevention, stock improvement or introduction of new varieties. For all three crops, it is assumed that, because of rotating practices and the requirements for sustained yield, a uniform age-class distribution is generally found on production farms. Hence, there would be no net increase or decrease in biomass carbon within existing farms, as carbon lost from harvest or replacement would be balanced by gains due to new plant growth. The approach therefore was limited to detecting changes in areas under vineyards, fruit orchards and Christmas tree plantations and estimating the corresponding carbon stock changes in total biomass. More

information on assumptions and parameters can be found in Annex 3.4.

## Uncertainties and Time-Series Consistency

Upon a loss of area with perennial woody crops, all carbon in woody biomass is assumed to be immediately released. It is assumed that the uncertainty for carbon loss equals the uncertainty about mass of woody biomass carbon. The default uncertainty of  $\pm 75\%$  (i.e. 95% confidence limits) for woody biomass on cropland from the IPCC Good Practice Guidance (IPCC 2003) was used.

If the loss in area of fruit trees, vineyards or Christmas trees is estimated to have gone to annual crops, there is also a deemed perennial to annual crop conversion with associated uncertainty that contributes to carbon change uncertainty. For area of gain in fruit trees, vineyards or Christmas trees, the uncertainty in annual carbon change was also assumed to be the default uncertainty of  $\pm 75\%$  (i.e. 95% confidence limits) (IPCC 2003).

The overall mean and uncertainties associated with emissions or removals of CO<sub>2</sub> from woody specialty crops were estimated to be  $5 \pm 0.5$  kt CO<sub>2</sub> eq for the level uncertainty and  $-40 \pm 60$  kt CO<sub>2</sub> eq for the trend uncertainty (McConkey et al. 2007).

The same methodology was used for the entire time series of emission estimates (1990–2009).

## QA/QC and Verification

This category has undergone Tier 1 QC checks (see Annex 6) in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

## Recalculations

There was no recalculation involved in emission estimates for this source category.

## Planned Improvements

There is no immediate plan in place aimed at improving emission estimates for this category.

## 7.4.2. Land Converted to Cropland

This subcategory includes the conversion of forest land and grassland to cropland. Emissions from the conversion of forest land to cropland account for nearly 100% of the total emissions in this category, which have decreased from 13 Mt CO<sub>2</sub> eq in 1990 to 5.5 Mt CO<sub>2</sub> eq in 2009. Emissions from the conversion of grassland are relatively insignificant.

### 7.4.2.1. Forest Land Converted to Cropland

Clearing forest for use as agricultural land is an ongoing but declining practice in Canada, although agriculture remains an important cause of forest conversion (accounting for 43% of forest area conversion in 2009). The cumulative area of forest land converted to cropland was 1277 kha in 1990; in 2009, the cumulative area converted since 1989 was 473 kha. Methods to determine the area converted annually are common to all forest conversion to other land-use categories; they are outlined in Section 7.8 of this chapter, under the heading “Forest Conversion.” In 2009, immediate emissions from this year’s forest conversion accounted for 3.2 Mt CO<sub>2</sub> eq, or 57% of all forest land converted to cropland emissions, while residual emissions from events that occurred in the last 20 years accounted for the remainder. Ninety-five percent of emissions originate from the biomass and dead organic matter pools during and after conversion, with the remainder being attributed to the soil pool. The residual emissions from the decay of dead organic matter and soil organic matter will last for decades. This long-term effect is unique to the LULUCF Sector.

#### Methodological Issues – Dead Organic Matter and Biomass Pools

As stated above, emissions from the dead organic matter (DOM) and biomass pools account for almost all emissions due to the conversion of forests to cropland. Their estimation is performed in the same modelling environment as that used for forest land remaining forest land. A general description of this modelling environment was provided in Section 7.3.1.1; more information is provided in Annex 3.4.

#### Methodological Issues - Soils

Emissions from soils in this category include the net C stock change due to the actual conversion, a very small

net CO<sub>2</sub> source from change in management practices in the 20 years following conversion, and the N<sub>2</sub>O emissions from the decay of soil organic matter. The soil emissions from forest land conversion to cropland were calculated by multiplying the total area of conversion by the empirically derived emission factor along with modelling-based SOC dynamics (see Annex 3.4). As explained below, patterns of change in SOC after the conversion of forest to cropland clearly differ between eastern and western Canada.

#### Eastern Canada

Essentially, all land in the eastern part of the country was forested before its conversion to agriculture. Many observations, either in the scientific literature or the Canadian Soil Information System, of forest SOC comparisons with adjacent agricultural land in eastern Canada show a mean loss of carbon of 20% at depths to approximately 20–40 cm (see Annex 3.4). Average nitrogen change was –5.2%, equivalent to a loss of approximately 0.4 Mg N/ha. For those comparisons where both nitrogen and carbon losses were determined, the corresponding carbon loss was 19.9 Mg C/ha. Therefore, it was assumed that nitrogen loss was a constant 2% of carbon loss.

The CENTURY model (Version 4.0) is used to estimate the SOC dynamics from conversion of forest land to cropland in eastern Canada. More details of methodologies for determining the maximal carbon loss and its rate constant associated with the conversion of forest land can be found in Annex 3.4.

Following a Tier 2-type methodology, as was done for direct N<sub>2</sub>O emissions from agricultural soils (see Agriculture Sector, Chapter 6), emissions of N<sub>2</sub>O from forest conversion to cropland were estimated by multiplying the amount of carbon loss by the fraction of nitrogen loss per unit of carbon and by an emission factor (EF<sub>BASE</sub>). EF<sub>BASE</sub> was determined for each ecodistrict based on topographic and climate conditions (see Annex 3.3).

#### Western Canada

Much of the current agricultural land in western Canada (Prairies and British Columbia) was grassland in the native condition. Hence, forest land converted to cropland has been primarily of forest that lies on the fringe of former grassland areas.

The Canadian Soil Information System (CanSIS) represents the best available data source for SOC under forest and

agriculture. On average, these data suggest that there is no loss of SOC from forest conversion and that, in the long term, the balance between carbon input and SOC mineralization under agriculture remains similar to what it was under forest. It is important to recognize that along the northern fringe of western Canadian agriculture, where most forest conversion is occurring, the land is marginal for arable agriculture; pasture and forage crops are the dominant management practices.

For western Canada, no loss of SOC over the long term was assumed from forest land converted to cropland managed exclusively for seeded pastures and hayland. The carbon loss from forest conversion in western Canada results from the loss of above- and below-ground tree biomass and from loss or decay of other above- and below-ground coarse woody DOM that existed in the forest at the time of forest conversion. The average nitrogen change in western Canada for sites at least 50 years from breaking was +52% (see Annex 3.4), reflecting substantial added nitrogen in agricultural systems compared with forest management practices. However, recognizing the uncertainty about actual carbon-nitrogen dynamics for forest conversion, loss of forest land to cropland in western Canada was assumed not to be a source of N<sub>2</sub>O.

### Uncertainties and Time-Series Consistency

Greenhouse gas fluxes from forest land converted to cropland result from the combination of (i) burning or harvesting (immediate emissions from biomass and dead organic matter or transfers to HWP accounted for as immediate emissions, respectively; (ii) the organic matter decay and subsequent CO<sub>2</sub> emissions in the DOM pool; and (iii) the net carbon losses from SOC. Note that immediate CO<sub>2</sub> emissions always refer to area converted in the inventory year; residual emissions, while also occurring on land converted during the inventory year, mostly come from land converted over the last 20 years. Non-CO<sub>2</sub> emissions are produced only by burning, and occur during the conversion process.

Immediate and residual CO<sub>2</sub> emissions from the biomass and DOM pools represent the largest components of this category, and contribute the most to the category uncertainty. In all cases, uncertainty values are presented as the 95% confidence interval (immediate emissions – ±23%, residual emissions from the DOM pool – ±36%, and residual emissions from the soil pool – ±59%). Uncertainty values associated with non-CO<sub>2</sub> emissions were estimated

to be ±23% for the immediate emissions and ±37% for the residual emissions from the DOM pool.

Reflecting the estimation approach and procedures, uncertainty estimates were derived independently for the biomass and dead organic matter pools, and for soil organic matter. The uncertainty about activity data described in Section 7.8.2 was incorporated in all analyses.

### Biomass and Dead Organic Matter Pools

The fate of biomass and DOM upon forest conversion and the ensuing emissions are modelled in the same framework as that used for forest land; the corresponding uncertainty estimates were therefore also developed within this framework and with the same Monte Carlo runs that generated uncertainty estimates in the Forest Land category. The analysis was updated with the time series 1990–2009 of the 2011 submission. A description of the general approach is provided in Section 7.3.1.2; more information can be found in Section 3.4.2.4 of Annex 3.4.

### Soil Pool

The uncertainty about the net CO<sub>2</sub> flux from the soil pool was estimated analytically (McConkey et al. 2007). More information is provided in Annex 3.4.2.4 on the general approach used to conduct this analysis.

### QA/QC and Verification

This category has undergone Tier 1 QC checks (see Annex 6) in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). Quality checks were also performed externally by Agriculture and Agri-Food Canada, which derived the estimates of SOC change. The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

To address the question raised by the expert review team during the 2009 annual inventory review, Canada has provided additional materials in Annex 3.4 to support the methodology.

### Recalculations

Recalculations have been carried out to the 1990–2008 time series of estimates due mainly to the area update of forestland conversion to cropland (see Section 7.8.4). As a result, there is a reduction of emissions by 1.3 Mt CO<sub>2</sub> eq in 1990 and 1.6 Mt CO<sub>2</sub> eq in 2008. In addition, changes in land management practices are assumed not to

occur for 20 years after forest conversion to cropland in western Canada, leaving perennial or forage crops as the only crops to grow until the land is transferred to the Cropland Remaining Cropland category. This change resulted in a net increase of emissions by 11 kt CO<sub>2</sub> eq in 1990 and 47 kt CO<sub>2</sub> eq in 2008.

## Planned Improvements

Planned improvements described under Section 7.8, Forest Conversion, will also affect this category.

### 7.4.2.2. Grassland Converted to Cropland

Conversion of native grassland to cropland occurs in the Prairie region of the country and generally results in losses of SOC and soil organic nitrogen and emissions of CO<sub>2</sub> and N<sub>2</sub>O to the atmosphere. It is assumed that carbon losses from the above-ground or below-ground biomass or dead organic matter upon conversion are insignificant. This assumption largely results from the definitional framework of land categories (see Section 7.2). Total emissions in 2009 from soils amounted to 9 kt CO<sub>2</sub> eq, including carbon losses and N<sub>2</sub>O emissions from the conversion.

## Methodological Issues

A number of studies on changes of SOC and soil organic nitrogen in grassland converted to cropland have been carried out on the Brown, Dark Brown and Black soil zones of the Canadian Prairies. The average loss of SOC was 22%, and the corresponding average change in soil organic nitrogen was 0.06 kg N lost/kg C (see Annex 3.4).

The CENTURY model (Version 4.0) is used to estimate the SOC dynamics from breaking of grassland to cropland for the Brown and Dark Brown Chernozemic soils. More details of methodologies for determining the maximal carbon loss and its rate constant associated with the breaking of grassland can be found in Annex 3.4.

Similar to N<sub>2</sub>O emissions in forest converted to cropland, emissions of N<sub>2</sub>O in grassland converted to cropland were estimated by a Tier 2 methodology, multiplying the amount of carbon loss by the fraction of nitrogen loss per unit of carbon by a base emission factor (EF<sub>BASE</sub>). EF<sub>BASE</sub> is determined for each ecodistrict based on climate and topographic characteristics (see Annex 3.3.6).

## Uncertainty and Time-Series Consistency

The conversion from agricultural grassland to cropland

occurs, but within the land definitional framework the conversion in the other direction is not occurring (see Section 7.2). Therefore, the uncertainty of the area of this conversion cannot be larger than the uncertainty about the area of cropland or grassland. Hence, the uncertainty of the area of conversion was set to the lower of the uncertainties of the area of either cropland or grassland in each ecodistrict. The uncertainty of SOC change was estimated as in forest land conversion to cropland. The overall mean and uncertainty associated with emissions due to SOC losses on grassland conversion to cropland were estimated to be  $9 \pm 11$  kt CO<sub>2</sub> eq for the level uncertainty, and  $-63 \pm 42$  kt CO<sub>2</sub> eq for the trend uncertainty.

The same methodology and emission factors are used for the entire time series of emission estimates (1990–2009).

## QA/QC and Verification

This category has undergone Tier 1 QC checks (see Annex 6) in a manner consistent with IPCC Good Practice Guidance (IPCC 2000). The activity data, methodologies and changes to methodologies are documented and archived in both paper and electronic form.

## Recalculations

There was no recalculation involved in emission estimates for this source category.

## Planned Improvements

Canada plans to validate the modelled soil carbon change factors with the measured and published soil carbon change factors from grassland conversion.

## 7.5. Grassland

Agricultural grassland is defined under the Canadian LULUCF framework as pasture or rangeland on which the only land management activity has been the grazing of domestic livestock (i.e. the land has never been cultivated). It occurs only in geographical areas where the grassland would not naturally grow into forest if abandoned: the natural shortgrass prairie in southern Saskatchewan and Alberta and the dry, interior mountain valleys of British Columbia. Agricultural grassland is found in three reporting zones: Semi-Arid Prairies (5600 kha), Montane Cordillera (200 kha), and Pacific Maritime (4 kha). As with cropland, the change in management triggers a change in carbon stocks (IPCC 2003). Very little information is available on management practices on Canadian agricul-

tural grassland, and it is unknown whether grazed land is improving or degrading. Therefore, Canada reports this Grassland remaining grassland category as not estimated. More details on the rationale for not estimating this category are provided in Annex 3.4. The subcategory land converted to grassland, within the current definitional framework as explained in Section 7.2, is reported either as not estimated (wetlands converted to grassland) or as not occurring (Table 7–3).

## 7.6. Wetlands

In Canada, a wetland is land that is saturated with water long enough to promote anaerobic processes, as indicated by poorly drained soils, hydrophytic vegetation and various kinds of biological activity that are adapted to a wet environment—in other words, any land area that can keep water long enough to let wetland plants and soils develop. As such, wetlands cover about 14% of the land area of Canada (Environment Canada 2003). The Canadian Wetland Classification System groups wetlands into five broad categories: bogs, fens, marshes, swamps and shallow water (National Wetlands Working Group 1997).

However, for the purpose of this report and in compliance with land categories as defined in IPCC (2003), the Wetlands category should be restricted to those wetlands that are not already in the Forest, Cropland or Grassland categories. There is no corresponding area estimate for these wetlands in Canada.

In accordance with IPCC guidance (IPCC 2003), two types of managed wetlands are considered, where human intervention has directly altered the water table level and thereby the dynamics of GHG emissions/removals: peatlands drained for peat harvesting; and flooded land (namely, the creation of reservoirs). Owing to their differences in nature, GHG dynamics and the general approaches to estimating emissions and removals, these two types of managed wetlands are considered separately.

### 7.6.1. Managed Peatlands

#### 7.6.1.1. Source Category Description

Of the estimated 123 Mha of peatlands in Canada,<sup>6</sup> approximately 23 kha are, or were at some point in

the past, drained for peat extraction. Some 12 kha are currently being actively managed. The other 11 kha consist of peatlands that are no longer under production. In the Canadian context, generally only bog peatlands with a peat thickness of 2 m or greater and an area of 50 ha or greater are of commercial value for peat extraction (Keys 1992). Peat production is concentrated in the provinces of New Brunswick, Quebec and Alberta. Canada produces only horticultural peat.

Since the 1980s, virtually all peat extraction in Canada has relied on vacuum harvest technology; approximately 100 t/ha (wet basis) of horticultural peat is extracted with this technology (Cleary 2003). A drawback of the technology, as opposed to the traditional cut-block method, is poor natural vegetation regrowth in the post-production phase. In the 1990s, peatland restoration activities took on greater significance.

Peat extraction activities expanded during the 1990–2000 period, with a 47% increase in the land area under active peat extraction, from 9.5 kha in 1990 to 14 kha at the turn of the century. Owing to this expansion and to the significant contribution of vegetation clearing and decay to the overall GHG budget, emissions from managed peatlands show a significant increase over the first half of the assessment period (Figure 7–3).

Emissions from managed peatlands are reported under land converted to wetlands for the first 20 years after conversion and under wetlands remaining wetlands thereafter.

#### 7.6.1.2. Methodological Issues

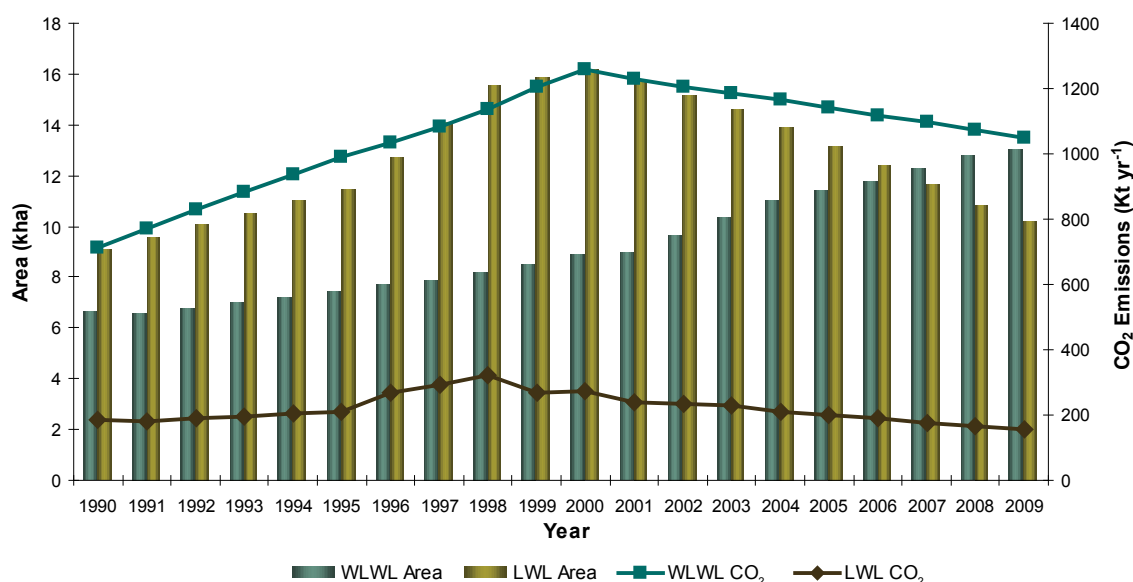
The general phases of peat extraction are 1) drainage, 2) vegetation clearing, 3) extraction, 4) stockpiling, 5) abandonment and 6) peatland restoration and establishment of natural vegetation. Due to drainage, CO<sub>2</sub> is the dominant GHG emitted from commercial peatlands and the only gas reported under this category. The main sources of emissions are vegetation clearing upon conversion, the continuing decay of dead organic matter and the rapid oxidation of exposed peat, resulting in a threefold increase in CO<sub>2</sub> emission rates (Waddington and Warner 2001). Estimates were developed using a Tier 2 methodology, based on domestic emission factors. They include emissions and removals during all five phases. More information on estimation methodology can be found in Annex 3.4.

Note that the methodology does not include carbon

<sup>6</sup> This area includes peatlands that would be classified as Forest, Cropland and Grassland in the IPCC land classification.



Figure 7-3 Areas of Managed Peatlands and CO<sub>2</sub> Emissions from These Lands, 1990–2009  
(LWL: Land Converted to Wetlands; WLWL: Wetlands Remaining Wetlands)



losses from the peat transported off-site; should these be included, total emissions from managed peatlands would significantly increase.

#### 7.6.1.3. Uncertainty and Time-Series Consistency

There was no formal uncertainty assessment for carbon emissions and removals in managed peatlands. The most important sources of uncertainty are discussed below.

Emission factors were derived from flux measurements made mostly over abandoned peatlands, which introduces significant uncertainty when applied to actively managed peatlands, and peat stockpiles. All measurements were conducted in eastern Canada, adding uncertainties to estimates for western Canada. A single estimate of preconversion forest biomass carbon density (20 t C/ha) was assumed; based on the characteristics of forest stands converted to peatland, an average 63% of above-ground biomass was deemed harvested at clearing.

Spatially referenced information on the areas of managed peatlands is currently not available; therefore these are modelled based on general information provided by the industry.<sup>7</sup> This introduces significant uncertainty about activity data. In addition, the fate of abandoned peatlands

is not monitored in Canada; older peat fields could have been converted to other uses. Therefore, the area estimate of abandoned peatlands is probably conservative.

#### 7.6.1.4. QA/QC and Verification

Annex 6 describes the general QA/QC procedures being implemented for Canada's GHG inventory; they apply to this category as well. Areas were derived in collaboration with the Canadian Sphagnum Peat Moss Association.

#### 7.6.1.5. Recalculations

Improvements in the modelling of area under peat production were implemented to the entire time series. These improvements prevent recalculations of the entire time series when activity data are incorporated for an additional year. Updated peat production data for 2007 and 2008 from the *Canadian Minerals Yearbook*<sup>8</sup> was incorporated.

Due to the unavailability of peat production data for the year 2009, at the time this submission was prepared, the average annual relative decrease in production from 2004 to 2008 was used in combination with the 2008 modelled area to estimate the annual area extracted in 2009.

7 Gerry Hood, Canadian Sphagnum Peat Moss Association, personal communication to D. Blain, Environment Canada, 2006.

8 <http://www.nrcan-rncan.gc.ca/mms-smm/busi-indu/cmy-amc/2008cmy-eng.htm>



### 7.6.1.6. Planned Improvements

Work is underway to decrease the uncertainty of the estimate of preconversion aboveground biomass (or biomass removal due to peat extraction) by analyzing geospatial information related to peatlands. Visual air photo interpretation will be used to identify area of activity along with the pre- and post-disturbance conditions of peat extraction sites.

## 7.6.2. Flooded Lands (Reservoirs)

This category includes in theory all lands that have been flooded regardless of purpose. Owing to methodological limitations, this submission includes only large hydro-electric reservoirs created by land flooding. Existing water bodies dammed for water control or energy generation were not considered if flooding was minimal (e.g. Manitoba's Lake Winnipeg, the Great Lakes).

Since 1970, land conversion to flooded lands occurred in reporting zones 4, 5, 8, 10 and 14. The total land area flooded for 10 years or less declined from 902 kha in 1990 to 92 kha in 2009. In 2009, 61% of the 92 kha of reservoirs flooded for 10 years or less were previously forested (mostly un-managed forests).

Total emissions from reservoirs declined from 4.2 Mt in 1990 to 1.3 Mt in 2009.

### 7.6.2.1. Methodological Issues

Two concurrent estimation methodologies were used to account for GHG fluxes from flooded lands—one for forest clearing and the other for flooding. When there was evidence of forest biomass clearing and removal prior to flooding, the corresponding carbon stock changes for all non-flooded carbon pools were estimated as in all forest conversion events, using the CBM-CFS3 (refer to Section 7.8 below and Annex 3.4). Emissions from the burning and decay of all non-flooded dead organic matter are reported under land converted to wetlands for the first 10 years post-clearing and in wetlands remaining wetlands beyond this period. The recent construction of large reservoirs in northern Quebec (Toulousteuc, Eastmain-1, Peribonka), whose impoundments were completed in 2005, 2006 and 2008, respectively, resulted in this type of forest clearing prior to flooding. Note that emissions from forest clearing in the general area surrounding future reservoirs (e.g.

for infrastructure development) are reported under forest conversion to settlements.

The second methodology is applied to estimate CO<sub>2</sub> emissions from the surface of reservoirs whose flooding has been completed. The default approach to estimate emissions from flooding assumes that all forest biomass carbon is emitted immediately (IPCC 2003). In the Canadian context, this approach would overestimate emissions from reservoir creation, since the largest proportion of any submerged vegetation does not decay for an extended period. A domestic approach was developed and used to estimate emissions from reservoirs based on measured CO<sub>2</sub> fluxes above reservoir surfaces, consistent with the descriptions of IPCC Tier 2 methodology (IPCC 2003, 2006) and following the guidance in Appendix 3a.3 of IPCC (2003). Annex 3.4 of this National Inventory Report contains more detail on this estimation methodology. In keeping with good practice, only CO<sub>2</sub> emissions are included in the assessment. Emissions from the surface of flooded lands are reported for a period of 10 years after flooding, in an attempt to minimize the potential double counting of dissolved organic carbon lost from managed lands in the watershed and subsequently emitted from reservoirs. Therefore, only CO<sub>2</sub> emissions are calculated for hydro-electric reservoirs where flooding had been completed between 1981 and 2009.

For each reservoir, the proportion of pre-flooding area that was forest is used to apportion the resulting emissions to the subcategories forest land converted to wetlands and other land converted to wetlands.

It is important to note that fluctuations in the area of lands converted to wetlands (reservoirs) reported in the CRF tables are not indicative of changes in current conversion rates, but reflect the difference between land areas recently flooded (less than 10 years before the inventory year) and older reservoirs (more than 10 years before the inventory year), whose areas are thus transferred out of the inventory. The reporting system does not encompass all the reservoir areas in Canada.

### 7.6.2.2. Uncertainties and Time-Series Consistency

For forest land converted to wetlands, refer to the corresponding subheading in Section 7.8, Forest Conversion. Annex 3.4 discusses the uncertainty associated with the Tier 2 estimation methodology.

Owing to current limitations in LULUCF estimation methodologies, it is not possible to fully monitor the fate of dissolved organic carbon and ensure that it is accounted for under the appropriate land category. The possibility of double counting in the Wetlands category is, however, limited to watersheds containing managed lands, which would exclude several large reservoirs in reporting zones 4 and 5.

### 7.6.2.3. QA/QC and Verification

Annex 6 describes the general QA/QC procedures being implemented for Canada's GHG inventory; they apply to this category as well.

Additional Tier 2 QC checks were performed on activity data, emission factors and methodology (for further explanation see Section 7.6.2.4 Recalculations).

For forest land converted to wetlands, also refer to the corresponding subheading in Section 7.8, Forest Conversion.

Canada's approach to estimating emissions from forest flooding is more realistic temporally than the default approach (IPCC 2003), which assumes that all biomass carbon on flooded forests is immediately emitted. Canada's method is more refined in that it distinguishes forest clearing and flooding; emissions from the former are estimated as in all forest clearing associated with land-use change. Further, in Canada's approach, emissions from the surface of reservoirs are derived from measurements, rather than from an assumption (decay of submerged biomass) that clearly is not verified.

### 7.6.2.4. Recalculations

Recalculations include both updated activity data and methodological changes. Over the time series, change in total emissions are not uniform, but range from an increase of 1.3 Mt CO<sub>2</sub> (45%) in 1993 to a decrease of 417 kt CO<sub>2</sub> (22%) in 2003.

Updates to activity data had the largest impact on recalculations and occurred as a result of quality control procedures supported by correspondence with industry representatives. These changes incorporate updates to dates that reservoir flooding was complete and to post-flooding reservoir area estimates. Changes to the year that flooding is completed have a large impact on the trend in the time series of emission estimates, as they shift the 10-year period of reporting. For example, the shift of the accounting

period of the La Forge 1 reservoir by two years (where the start of the accounting period shifted from 1995 to 1993) results in an increase in emissions in 1993 and 1994 and decreases emissions in 2003 and 2004. Additionally, the change in the date that flooding of the Opinaca reservoir was complete (from 1980 to 1984) results in reporting additional emissions from 1990 to 1993 in the 2011 submission, which increased total emissions during this period.

Carbon transfer parameters for forest clearing prior to reservoir flooding were updated. These changes more accurately reflect the fate of flooded and non-flooded organic matter after impoundment.

A single temporal emission curve replaced the two regional curves used to determine CO<sub>2</sub> emissions from the surface of reservoirs (see Annex 3.4). This change results from a statistical assessment revealing that the curve parameters for the Taiga and Boreal and Montane Cordillera regions were not significantly different from each other. Combining the datasets of the two regions results in a larger sample and improves the precision of annual emission factors. The emission factors derived from the single temporal emissions curve are greater than previous ones of the Taiga and Boreal zone (by 10–57 mg/m<sup>2</sup>/d CO<sub>2</sub> or ~1% change) and the Montane Cordillera zone (by 195–420 mg/m<sup>2</sup>/d CO<sub>2</sub> or 10–12% change). As currently emissions are reported from only one reservoir in the Montane Cordillera zone (Revelstoke), changes in emission factors resulted in small increases in emissions from the surfaces of reservoirs (increases in emissions never exceed 23 kt CO<sub>2</sub> or a 1% change).

### 7.6.2.5. Planned Improvements

Further refining estimates of CO<sub>2</sub> emissions from the surface of reservoirs partly rests upon the quantification of lateral transfers of dissolved carbon from the watershed. The monitoring of dissolved organic carbon as it travels through the landscape to the point of emission or long-term storage is beyond current scientific capabilities, and will require long-term investments in research. Efforts to ensure activity data are updated and validated will continue on an ongoing basis.

## 7.7. Settlements

The Settlements category is very diverse, and includes all roads and transportation infrastructure; rights-of-way for power transmission and pipeline corridors; residential, recreational, commercial and industrial lands in urban and rural settings; and land used for resource extraction other than forestry (oil and gas, mining).

In settlements remaining settlements, urban trees contribute very little to the national GHG budget. Estimates for 2009 indicate modest removals of less than 0.2 Mt.

For the purpose of this inventory, two types of land conversion to settlements were estimated: forest land conversion to settlements, and non-forest land conversion to settlements in the Canadian north. In 2009, 480 kha of lands converted to settlements accounted for emissions of a little more than 9 Mt. Forest land conversion to settlements represents 98% of these emissions. The conversion of cropland to settlements is known to occur in Canada; an approach to developing activity data and an estimation methodology is under development.

### 7.7.1. Settlements Remaining Settlements

This category includes estimates of carbon sequestration in urban trees. No modification has been made in activity data or methods since the last submission. The current approach considers only the removal activity of urban trees on the non-built-up portion of urban areas. This component, although approximate, makes a very minor contribution to the LULUCF Sector and represents a low priority for improvement.

#### 7.7.1.1. QA/QC and Verification

A detailed QC study was conducted on the data and approach used to estimate removals by urban trees. The result of this study was a correction of a minor error in the estimation model with suggested improvements for data and methods in future submissions.

## 7.7.2. Land Converted to Settlements

### 7.7.2.1. Source Category Description

In 2009, emissions from land conversion to settlements amounted to a little more than 9 Mt. While there are potentially several land categories, including forests, that have been converted to settlements, there are currently insufficient data to quantify areas or associated emissions for all types of land-use change. Significant efforts were invested in quantifying the areas of forest land converted to settlements, the second forest conversion type in importance after forest conversion to cropland. On average, during the 1990–2009 period 24, kha of forest land are converted annually to settlements, predominantly in the Boreal Plains, Montane Cordillera and Atlantic Maritime reporting zones. Forest land conversion accounts for nearly 100% of emissions reported under this category. A consistent methodology was developed for all forest conversion, which is outlined in Section 7.8.

This remainder of this section covers non-forest land conversion to settlements in the Canadian north, primarily the Arctic and Sub-Arctic regions and reporting zones 4 and 8. In 2009, the conversion of non-forest land to settlements in the Canadian north accounted for emissions of 0.15 Mt; this value is very similar in the entire trend from 1990. The major source of emissions in this category is associated with conversion of grassland to settlement land in reporting zone 13, the Taiga Plains.

### 7.7.2.2. Methodological Issues (Non-forest Land Converted to Settlements)

Resource development in Canada's vast northern ecumene is the dominant driver of land-use change. An accurate estimation of this direct human impact in northern Canada requires that activities be geographically located and the preconversion vegetation known—a significant challenge, considering that the area of interest extends over 557 Mha, intersecting with eight reporting zones (2, 3, 4, 8, 10, 13, 17 and 18). For all reporting zones except 4 and 8, various information sources and geographic data sets were used to identify areas of high land-use change potential and narrow down the geographical domain of interest. These areas were targeted for change detection analysis using 23 Worldwide Reference System Landsat frames from circa 1985, 1990 and 2000. The scenes cover more than 8.7 Mha,

or 56% of the area with high potential for land-use change. Lack of available imagery prevented the implementation of the system beyond 2000.

For reporting zones 4 and 8, a change enhancement and manual delineation approach was implemented for the 1975–2000 time period for the entire area.

Emissions include only the carbon in preconversion above-ground biomass. In spite of the existing relevant literature, the estimation of actual or average biomass density over such a large area is challenging and remains fraught with uncertainty.

### 7.7.2.3. Uncertainties and Time-Series Consistency

For forest land converted to settlements, refer to the corresponding subheading in Section 7.8, Forest Conversion.

The uncertainty about the area of non-forest land converted to settlements in the Canadian north is estimated at 20%; the uncertainty about the preconversion standing biomass varies between 35% and 50%. Annex 3.4 provides more information.

### 7.7.2.4. QA/QC and Verification

Annex 6 describes the general QA/QC procedures being implemented for Canada's GHG inventory; they apply to this category as well.

For forest land converted to settlements, refer to the corresponding subheading in Section 7.8, Forest Conversion.

### 7.7.2.5. Planned Improvement

Future efforts to improve estimates for this category will focus on improving estimates of above-ground biomass for pre-conversion condition for land-use change events in the Arctic and Sub-Arctic region; updating estimates of activity data for land-use change in the Arctic and Sub-Arctic region for the post-2000 time period; and improving the data and estimation approach used for removals associated with urban forests.

In addition, planned improvements described under Section 7.8, Forest Conversion, will also affect this category (see Section 7.8.5, Planned Improvements).

## 7.8. Forest Conversion

Forest conversion is not a reporting category, since it overlaps with the subcategories of land converted to cropland, land converted to wetlands and land converted to settlements; it is nevertheless reported as a memo item. This section will briefly discuss methodological issues specific to this type of land-use change and outline the general approach taken to estimate its extent, location and impact. A consistent approach was applied for all types of forest conversion, minimizing omissions and overlaps, while maintaining spatial consistency as much as possible.

In 2009, forest conversion to cropland, wetlands and settlements amounted to total emissions of 18 Mt, down from 26 Mt in 1990. This decline includes a 6 Mt decrease in immediate and residual emissions due to forest conversion to cropland and a 1.5 Mt decrease in emissions from forest conversion to reservoirs. Note that this assessment includes residual emissions more than 20 years after conversion (10 years for reservoirs) that are included in the "land remaining..." categories.

Care should be taken to distinguish annual deforestation rates (65 kha in 1990 and 44 kha in 2009) from the total area of forest land converted to other uses as reported in the CRF tables for each inventory year. The CRF figures encompass all forest land conversion for 20 years prior to the current inventory year (10 years for reservoirs) and hence are significantly higher than the annual rates of forest conversion to other land use.

Likewise, for the year 2009, emissions from the conversion of forest land differ from those due to deforestation reported under the Kyoto Protocol. This divergence is solely due to the differences in category definition, as opposed to methods or data. This is further explained in Annex 11.

It is also important to note that immediate emissions from forest conversion, which occur upon the conversion event, are only a fraction of the total emissions due to current and previous forest conversion activities reported in any inventory year; some of these "immediate" emissions are carbon transferred to forest products. In 2009, immediate emissions (7.2 Mt) represented only 39% of the total reported emissions due to forest conversion; the balance is accounted for by residual emissions due to current and prior events. Decay rates for dead organic matter are such that residual emissions continue beyond 20 years, after which they are reported in the carbon stock changes in

cropland remaining cropland and wetlands remaining wetlands.

With a current annual conversion rate of 25 kha, forest conversion to settlements accounts for the largest share of forest losses to other land categories. In 2009 conversion to cropland (19 kha) was the second most important cause of deforestation, representing 43% of all forest area lost. The occasional impoundment of large reservoirs (e.g. La Forge 1 in 1993) may convert large forest areas to wetlands (flooded land); because much of the pre-conversion C stocks are flooded, these punctual events may not release commensurate quantities of greenhouse gases.

Geographically, the highest rates of forest conversion occur in the Boreal Plains (reporting zone 10), which accounts for 50% of the total forest area lost in 2009.

Forest conversion affects both managed and un-managed forests. Losses of un-managed forests occur mainly in reporting zone 4 (Taiga Shield East) and are caused mostly by reservoir impoundment; they occur to a smaller extent in reporting zones 8 and 9.

### 7.8.1. Methodological Issues

Forest conversion to other land categories is still a prevalent practice in Canada. This phenomenon is driven by a great variety of circumstances across the country, including policy and regulatory frameworks, market forces and resource endowment. The economic activities causing forest losses are very diverse; they result in heterogeneous spatial and temporal patterns of forest conversion, which, until recently, were not systematically documented. The challenge has been to develop an approach that would integrate a large variety of information sources to capture the various forest conversion patterns across the Canadian landscape, while maintaining a consistent approach in order to minimize omissions and overlap.

The approach adopted for estimating forest areas converted to other uses is based on three main information sources: systematic or representative sampling of remote sensing imagery, records, and expert judgement. The core method involves mapping of deforestation on samples from remotely sensed Landsat images dated circa 1975, 1990, 2000 and 2008. For implementation purposes, all permanent forest removal wider than 20 m from tree base to tree base and at least 1 ha in area was considered forest conversion. This convention was adopted as a guide to

consistently label linear patterns in the landscape. The other main information sources consist of databases or other documentation on forest roads, power lines, oil and gas infrastructure, and hydroelectric reservoirs. Expert opinion was called upon when the remote sensing sample was insufficient, to resolve differences among records and remote sensing information, and to resolve apparent discrepancies between the 1975–1990, 1990–2000 and 2000–2008 area estimates. A more detailed description of the approach and data sources is provided in Annex 3.4.

All estimates of emissions from biomass and dead organic matter pools due to forest conversion were generated using the CBM-CFS3 (Section 7.3.1.1), except when forests were flooded without prior clearing. Emissions from the soil pool were estimated in different modelling frameworks, except for land conversion to settlements where CBM-CFS3 decay rates were used. Hence, methods are in general consistent with those used in the forest land remaining forest land subcategory. Annex 3.4 summarizes the estimation procedures.

### 7.8.2. Uncertainties and Time-Series Consistency

Based on expert judgement, an overall uncertainty of  $\pm 30\%$  bounds the estimate of the total forest area converted annually in Canada (Leckie 2011), placing with 95% confidence the true value of this area for 2009 between 31 kha and 57 kha. Care should be taken not to apply the 30% range to the cumulative area of forest land converted to another category over the last 20 years (land areas reported in the CRF tables). Annex 3.4 describes the main sources of uncertainty about area estimates derived from remote sensing

### 7.8.3. QA/QC and Verification

Annex 6 describes the general QA/QC procedures being implemented for Canada's GHG inventory; they apply to this category as well. In addition, detailed Tier 2 QA/QC procedures were carried out during estimate development procedures, involving documented QC of imagery interpretation, field validation, cross-calculations and detailed examination of results (Dyk et al. 2011). The calculations, use of records data, and expert judgement are traceable through the compilation system and documented. More information is available in Annex 3.4.



## 7.8.4. Recalculations

The main sources for recalculations in the estimates for forest conversion are due to (i) the mapping of a new time period (2000–2008); (ii) the update of mapping with the use of higher resolution of Earth observation data; (iii) the greater reliance on mapping instead of records; and (iv) the expansion of the sample. The revision based on those changes has resulted in generally small recalculations in terms of overall estimates throughout the time series.

The addition of a third time period to the mapping (2000–2008) added a third anchor point in 2004, resulting in less use of extrapolated information compared to the 2010 submission, where a constant rate was applied from 1995 onwards. For this submission a constant deforestation rate is assumed for the post-2004 period (see annex 3.4 for details).

Overall, the annual rates of forest conversion have generally decreased throughout the time period, by amounts varying between 300 ha/yr (2008 inventory year) and 4.3 kha/yr (1995 inventory year). The largest recalculations have occurred in the area of forest land annually converted to cropland, which went down by up to 7.1 kha/yr (2004–2007 inventory years). Conversely, the annual rates of forest conversion to settlements have increased by up to 5.2 kha/yr (2008 inventory year). In all cases, recalculations were largest in the second half of the time series, due to the replacement of extrapolated estimates by actual mapping data. The effect of the downward recalculations of forest land conversion to cropland was thus partially offset by the upward recalculations of forest land conversion to settlements.

By adding a new period mapping from 2000 to 2008 and refining the mapping, recalculations show that since 1996 settlements are a more important driver of forest conversion than the expansion of agricultural land. In the current submission the annual rates for forest converted to cropland show a decrease from 42 kha in 1990 to 19 kha in 2008 (as opposed to a decline of 45 kha to 25 kha in the previous submission), while forest converted to settlement increased from 22 kha in 1990 to 26 kha in 2008 (as opposed to nearly constant rates of 22–21 kha/yr in the previous submission). Information for an additional time period will confirm this trend in future submissions.

The effects of recalculations on trends are proportionally more important at a spatially disaggregated level. The Boreal Plain reporting zone now shows a significant

reduction in forest converted to cropland, from 28 kha in 1990 to 10 kha in 2008, while the conversion to settlement undergoes a significant increase of 8 kha in 1990 to 12 kha in 2008. This is reflective of enhanced resource development in that region. Refined estimation methods are also being used, for example in estimating co-occurring disturbances such as pipeline construction and access roads. This large increase in the Boreal Plains activity is partially offset at a national scale by revised lower estimates in other regions.

Several changes were also applied on years of reservoir flooding that have an effect on emissions during the subsequent years. For example, correction in the year the Laforge 1 reservoir was flooded from 1995 (in the 2010 submission) to 1993 (this submission), resulted in an increase of 34 kha in forest area losses in 1993 (see Section 7.6.2).

Overall, these changes have a small (between 1% and 7%) impact on emissions. Compared to the 2010 submission, there is a 1.3 Mt reduction in emissions from forest conversion to cropland for the 1990 inventory year, and a 1.6 Mt reduction for the year 2008. For the 1990 inventory year there is a reduction of 0.4 Mt for forest converted in settlement and an increase of 1.5 Mt for year 2008. In 2008, the decrease in emissions from the conversion of forest to cropland has been nearly offset by the increased emissions associated with forest conversion to settlements.

## 7.8.5. Planned Improvements

Planned improvements emphasize QA/QC, increased mapping coverage in areas with high uncertainty, extension of the time period of mapping, field validation, use of additional records, and enhanced efficiency in the data compilation process.

# Chapter 8

## Waste (CRF Sector 6)

### 8.1. Overview

This Sector includes emissions from the treatment and disposal of wastes. Sources include solid waste disposal on land (landfills), wastewater treatment and waste incineration. The categories evaluated are CH<sub>4</sub> emissions from solid waste disposal on land, CH<sub>4</sub> and N<sub>2</sub>O emissions from wastewater treatment, and CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O emissions from waste incineration.

Much of the waste treated or disposed of is biomass or biomass-based. CO<sub>2</sub> emissions attributable to such wastes are not included in inventory totals but are reported in the inventory as a memo item. CO<sub>2</sub> emissions of biogenic origin are not reported if they are reported elsewhere in the inventory or if the corresponding CO<sub>2</sub> uptake is not reported in the inventory (e.g. annual crops). Therefore, under these circumstances, the emissions are not included in the inventory emission totals, since the absorption of CO<sub>2</sub> by the harvested vegetation is not estimated by the Agriculture Sector and, thus, the inclusion of these emissions in the Waste Sector would result in an imbalance. Also, CO<sub>2</sub> emissions from wood and wood products are not included, because these emissions are accounted for in the Land Use, Land-Use Change and Forestry (LULUCF) Sector at the time of tree harvesting. In contrast, CH<sub>4</sub> emissions

from anaerobic decomposition of wastes are included in inventory totals as part of the Waste Sector.

If carbon is lost from forests at an unsustainable rate (i.e. faster than annual re-growth), the carbon budget for forest lands will be negative for net emissions.

In 2009, the greenhouse gas (GHG) emissions from the Waste Sector contributed 22 Mt to the national inventory, compared with 19 Mt for 1990; an increase of 13%. In comparison, the national total emissions increased by 16.9% over the same time interval. The emissions from this sector represented 3.2% and 3.1% of the overall Canadian GHG emissions in 1990 and 2009, respectively.

Emissions from the Solid Waste Disposal on Land sub-sector, which consists of the combined emissions from municipal solid waste (MSW) landfills and wood waste landfills, accounted for 20 Mt or 94% of the emissions from this sector in 2009. The chief contributor to the Waste Sector emissions is the CH<sub>4</sub> released from MSW landfills, which, for 2009, amounted to 18 Mt (0.8 Mt CH<sub>4</sub>). This net emission value is determined by subtracting the amount of CH<sub>4</sub> captured from the total estimated CH<sub>4</sub> generated within the landfill by the Scholl Canyon model, then adding the quantity of the captured CH<sub>4</sub> that was not combusted by the flaring operation, where applicable. From our biennial survey of Canadian landfills, which collected 2008 and 2009 year data, approximately 29% of the CH<sub>4</sub> generated in Canadian MSW landfills in 2009 was captured and combusted (either for energy recovery, or flared).

Overall, the increase in the CH<sub>4</sub> generation rate from MSW landfills is primarily dependent upon the population growth, which in turn governs the waste generation rate, although other factors, such as increased gross domestic product, use of packaging materials and urbanization, could be in play. This upward influence is mitigated by landfill gas capture programs, provincial/municipal waste

Table 8–1 Waste Sector GHG Emission Summary, Selected Years

GHG Source Category		GHG Emissions (kt CO <sub>2</sub> eq)							
		1990	2000	2004	2005	2006	2007	2008	2009
	Waste Sector TOTAL	19 000	20 000	21 000	21 000	22 000	22 000	21 000	22 000
a.	Solid Waste Disposal on Land	18 000	19 000	20 000	20 000	21 000	20 000	20 000	20 000
b.	Wastewater Handling	780	930	970	980	990	1 000	1 000	1 000
c.	Waste Incineration	400	250	230	240	240	250	250	260

Note: Totals may not add up due to rounding.

diversion projects and international exportation of MSW. It is expected that, as larger and more “state-of-the art” landfills are constructed, where gas collection systems will be required, a greater portion of landfill gas will be captured in the future, resulting in a greater reduction of emissions from this sector. Nationally, in 2006, over 35 Mt of non-hazardous waste (residential, institutional, commercial, industrial, construction and demolition) was generated. Waste diversion initiatives began in the early 1990s and, based upon the national figures for 2006, approximately 22% of the waste generated is diverted from disposal (landfill or incineration) (Statistics Canada 2008a).

Table 8–1 summarizes the Waste Sector and subsector GHG contributions for the 1990, 2000, 2004, 2005, 2006, 2007, 2008 and 2009 inventory years.

## 8.2. Solid Waste Disposal on Land (CRF Category 6.A)

### 8.2.1. Source Category Description

Emissions are estimated from two types of landfills in Canada:

- MSW landfills; and
- wood waste landfills.

In Canada, most waste disposal on land occurs in managed municipal or privately owned landfills. Very few, if any, unmanaged waste disposal sites exist. Therefore, it has been assumed that all waste is disposed of in managed facilities. Residential, institutional, commercial and industrial wastes are disposed of in MSW landfills. Over the past 15 years, dedicated construction and demolition landfills were established. Typically, these landfills do not require CH<sub>4</sub> collection systems, as the CH<sub>4</sub> generation rate is very low due to the minimal organic content in the waste stream. Therefore, these landfills are currently excluded from the analysis.

Wood waste landfills are mostly privately owned and operated by forest industries, such as saw mills and pulp and paper mills. These industries use the landfills to dispose of surplus wood residue, such as sawdust, wood shavings, bark and sludges. Some industries have shown increasing interest in waste-to-energy projects that produce steam

and/or electricity by combusting these wastes. In recent years, residual wood previously regarded as a waste is now being processed as a value-added product—e.g., wood pellets for residential and commercial pellet stoves and furnaces, and hardboard, fibreboard and particle board. Wood waste landfills have been identified as a source of CH<sub>4</sub> emissions; however, there is a great deal of uncertainty in the estimates. These landfills are a minor source of CH<sub>4</sub> emissions in comparison with MSW landfills.

The *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997) provides two methodologies for estimating emissions from landfills: a default method and a first-order kinetics method, also known as the Scholl Canyon model. The default method relates emissions to the quantity of waste landfilled in the previous year, whereas the Scholl Canyon model relates emissions to the cumulative biologically available waste that has been landfilled in previous years.

The composition and amount of waste landfilled in Canada have significantly changed over the past several decades, primarily as a result of waste diversion initiatives and population growth, respectively. For this reason, a static model such as the default method is not felt to be appropriate. Therefore, emissions from MSW landfills and wood waste landfills are estimated using the Scholl Canyon model. The Scholl Canyon model, used to estimate Canada's CH<sub>4</sub> emissions from landfills, has been validated independently through a study conducted by the University of Manitoba (Thompson et al. 2006).

Landfill gas, which is composed mainly of CH<sub>4</sub> and CO<sub>2</sub>, is produced by the anaerobic decomposition of organic wastes. The first phase of this process typically begins after waste has been in a landfill for 10 to 50 days. Although the majority of the CH<sub>4</sub> and CO<sub>2</sub> gases are generated within 20 years of landfilling, emissions can continue for 100 years or more (Levelton 1991).

A number of important site-specific factors contribute to the generation of gases within a landfill, including the following:

**Waste composition:** Waste composition is probably the most important factor affecting landfill gas generation rates and quantities. The amount of landfill gas produced is dependent on the amount of organic matter landfilled. The rate at which gas is generated is dependent on the distribution and type of organic matter in the landfill.

**Moisture content:** Water is required for anaerobic degradation of organic matter; therefore, moisture content within a landfill significantly affects gas generation rates.

**Temperature:** Anaerobic digestion is an exothermic process. The growth rates of bacteria tend to increase with temperature until an optimum is reached. Therefore, landfill temperatures may be higher than ambient air temperatures. The extent to which ambient air temperatures influence the temperature of the landfill and gas generation rates depends mainly on the depth of the landfill. Temperature variations can affect microbial activity, subsequently affecting their ability to decompose matter (Maurice and Lagerkvist 2003).

**pH and buffer capacity:** The generation of CH<sub>4</sub> in landfills is greatest when neutral pH conditions exist. The activity of methanogenic bacteria is inhibited in acidic environments.

**Availability of nutrients:** Certain nutrients are required for anaerobic digestion. These include carbon, hydrogen, nitrogen and phosphorus. In general, MSW contains the necessary nutrients to support the required bacterial populations.

**Waste density and particle size:** The particle size and density of the waste also influences gas generation. Decreasing the particle size increases the surface area available for degradation and therefore increases the gas production rate. The waste density, which is largely controlled by compaction of the waste as it is placed in the landfill, affects the transport of moisture and nutrients through the landfill, which also affects the gas generation rate.

## 8.2.2. Methodological Issues

CH<sub>4</sub> produced from the decomposition of waste in landfills is calculated using the Scholl Canyon model, which is a first-order decay model. This reflects the fact that waste degrades in landfills over many years. Data pertaining to landfill gas capture were obtained directly from the owners/operators of specific landfills with landfill gas collection systems.

CH<sub>4</sub> emissions are determined by calculating the amount of CH<sub>4</sub> generated from landfill waste decomposition through the Scholl Canyon model, subtracting the CH<sub>4</sub> captured through landfill gas recovery systems, then adding the quantity of uncombusted CH<sub>4</sub> emitted by the flares for those locations where a portion or all of the recovered landfill gas is burned without energy recovery. The GHG

emissions associated with the combustion of that portion of the landfill gas that is captured and utilized for energy generation purposes are accounted for in the Energy Sector. A more detailed discussion of the methodologies is presented in Annex 3.5.

### 8.2.2.1. CH<sub>4</sub> Generation

The Scholl Canyon model was used to estimate the quantity of CH<sub>4</sub> generated. The model is based upon the following first-order decay equation (IPCC/OECD/IEA 1997):

Equation 8–1:

$$Q_{T,x} = k M_x L_0 e^{-k(T-x)}$$

where:

$Q_{T,x}$	=	amount of CH <sub>4</sub> generated in the current year (T) by the waste $M_x$ , kt CH <sub>4</sub> /year
$x$	=	the year of waste input
$M_x$	=	the amount of waste disposed of in year $x$ , Mt
$k$	=	CH <sub>4</sub> generation rate constant, /year
$L_0$	=	CH <sub>4</sub> generation potential, kg CH <sub>4</sub> /t waste
$T$	=	current year

Equation 8–2:

$$Q_T = \sum Q_{T,x}$$

where:

$Q_T$	=	amount of CH <sub>4</sub> generated in the current year (T), kt CH <sub>4</sub> /year
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In order to estimate CH<sub>4</sub> emissions from landfills, information on several of the factors described above is needed. To calculate the net emissions for each year, the sum of  $Q_{T,x}$  for every section of waste landfilled in past years was obtained (Equation 8–2), from which the captured gas was subtracted for each province. A computerized model has been developed to estimate aggregate emissions on a regional basis (by province and territory) in Canada. The national CH<sub>4</sub> emission value is the summation of emissions from all regions.

## Waste Disposed of Each Year or the Mass of Refuse ( $M_x$ )

### MSW Landfills

Two primary sources were used in obtaining waste generation and landfill data for the GHG inventory. The amounts of MSW landfilled in the years 1941 through to 1990 were estimated by B.H. Levelton (1991). For the years 1998, 2000, 2002, 2004 and 2006, MSW disposal data were obtained from the *Waste Management Industry Survey* that is conducted by Statistics Canada on a biennial basis (Statistics Canada 2000, 2003, 2004, 2007a, 2008a). For the intervening odd years (1999, 2001, 2003, 2005), the MSW disposal values, including both landfilled and incinerated MSW, were obtained by taking an average of the adjacent even years. Quantities of waste landfilled in 2007, 2008 and 2009 were trended from values derived from the Statistics Canada survey. The 2008 waste landfilled quantities from the *Waste Management Industry Survey* would have normally been used; however, these values were not available within the time constraints for the development of the present Inventory. Therefore, trended values were used for 2007, 2008 and 2009. These values will be revised in the 2012 NIR submission based on the *Waste Management Industry Survey* 2010 values. Incinerated waste quantities were subtracted from the Statistics Canada disposal values in order to obtain the amounts of MSW landfilled for 1998–2009. For the years 1991–1997, with the exception of Prince Edward Island, the Northwest Territories, Nunavut and Yukon, the quantities of waste disposed of were estimated from an interpolation using a multiple linear regression approach applied to the B.H. Levelton (1991) and Statistics Canada (2000, 2003, 2004, 2007a, 2008a) MSW landfill values. MSW landfill values for Prince Edward Island, the Northwest Territories, Nunavut and Yukon for the period 1991–2009 are obtained by trending historical landfill data with the provincial populations for 1971–2009 (Statistics Canada 2006, 2010b).

### Wood Waste Landfills

British Columbia, Quebec, Alberta and Ontario together landfill 93% of the wood waste in Canada (NRCan 1997). The amount of wood waste landfilled in the years 1970 through to 1992 has been estimated at a national level based on the National Wood Residue Data Base (NRCan 1997). Data for the years 1998 and 2004 were provided by subsequent publications (NRCan 1999, 2005). A linear regression trend analysis was conducted to interpolate the

amount of wood residue landfilled in the years 1991–1997 and 1999–2009.

## CH<sub>4</sub> Generation Rate Constant (k)

The CH<sub>4</sub> kinetic rate constant (k) represents the first-order rate at which CH<sub>4</sub> is generated after waste has been landfilled. The value of k is affected by four major factors: moisture content, temperature, availability of nutrients and pH. It is assumed that, in a typical MSW landfill, the nutrient and pH conditions are attained and that, therefore, these factors are not limiting. In many parts of Canada, subzero conditions exist for up to seven months of the year, with temperatures dropping below –30°C (Thompson et al. 2006); however, evidence suggests that ambient temperature does not affect landfill decay rates (Maurice and Lagerkvist 2003; Thompson and Tanapat 2005). In addition, seasonal temperature variations in the waste are minimal when compared with atmospheric temperature variations (Maurice and Lagerkvist 2003). At depths exceeding 2 m, the landfill temperature is independent of the ambient temperature. It has been shown in Canadian field experiments that an insignificant amount of variation in landfill CH<sub>4</sub> production occurs between the winter and summer seasons (Bingemer and Crutzen 1987; Thompson and Tanapat 2005). Therefore, of all these factors, moisture content is the most influential parameter for Canadian landfills and is largely determined by the annual precipitation received at the landfills.

### MSW Landfills

The k values used to estimate emissions from MSW landfills were obtained from a study conducted by Environment Canada's Greenhouse Gas Division that employed provincial precipitation data from 1941 to 2007 (Environment Canada 1941–2007). The provincial locations at which the average annual precipitations were calculated were those indicated in the Levelton study where major landfills were located over the 1941–1990 period (Levelton 1991). Since the k values are related to precipitation, and assuming that the moisture content of a landfill is a direct function of the annual precipitation, from these precipitation values, the associated k values were determined using a relationship prepared by the Research Triangle Institute (RTI) for the U.S. EPA (RTI 2004). The RTI assigns default decay values of less than 0.02/year, 0.038/year and 0.057/year to areas with an annual precipitation of less than 20 inches/year (< 500 mm), between 20 and 40 inches/year (500 to 1000 (average 750 mm)) and greater than 40 inches/year



Table 8–2 MSW Landfill k Value Estimates for Each Province/Territory

Time Series	Provinces and Territories											
	N.L.	P.E.I.	N.S.	N.B.	Que.	Ont.	Man.	Sask.	Alta.	B.C.	N.W.T & Nvt	Yk.
1941–1975	0.075	0.056	0.076	0.06	0.053	0.041	0.020	0.01	0.012	0.082	0.001	0.001
1976–1989	0.080	0.062	0.079	0.063	0.057	0.047	0.017	0.009	0.012	0.082	0.002	0.001
1990–2007	0.078	0.061	0.075	0.059	0.059	0.046	0.019	0.012	0.012	0.083	0.003	0.002

(> 1000 mm), respectively. The plot of these decay values and precipitation data showed a linear relationship. Using this relationship and Environment Canada's average provincial precipitation data for 1941–2007, average provincial landfill decay rates were calculated for three time periods that match those used to derive the methane generation potentials ( $L_0$ ), i.e., 1941–1975, 1976–1989 and 1990–2007 (Environment Canada 1941–2007). It is assumed that the provincial k values determined for 1990–2007 are also applicable for 2008 and 2009.

These values are provided in Table 8–2.

### Wood Waste Landfills

Based upon the default value for estimating wood products industry landfill  $\text{CH}_4$  emissions recommended by the National Council for Air and Stream Improvement, Inc., a k value of 0.03/year was assumed to represent the  $\text{CH}_4$  generation rate constant k for all of the wood waste landfills in Canada (NCASI 2003).

### $\text{CH}_4$ Generation Potential ( $L_0$ )

#### MSW Landfills

The values of theoretical and measured  $L_0$  range from 4.4 to 194 kg  $\text{CH}_4$ /t of waste (Pelt et al. 1998). Over the time series used by the MSW portion of the emission estimation model, i.e., 1941 to 2009, three different  $L_0$ s were used to represent discrete time periods where studies showed significant changes in waste composition from one period to the next.  $L_0$  is a function of degradable organic carbon (DOC), which is in turn determined from the composition of the waste, as described below.

The provincial and territorial DOC values were calculated from waste disposal composition values for three distinct time periods: 1941–1975, 1976–1989 and 1990–2009. Using waste composition data obtained from a Natural Resources Canada (NRCAN) study, which was based on the

2002 data year (NRCAN 2006), DOC values were derived and assumed to be constant over the period 1990–2009. Since waste diversion programs were not significant prior to 1990, a second set of DOC values was developed to represent the waste composition at disposal from 1976 to 1989 by adding the NRCAN landfill to the 2004 Statistics Canada recycled waste composition data (Statistics Canada 2007a). A third set of DOC values was developed from a 1967 national study to cover the period from 1941 to 1975 (CRC Press 1973). A summary of the  $L_0$  values for the provinces and territories over the three time periods is given in Table 8–3. The percentages of organic waste diverted in 2002 for all Canadian provinces are also given as a reference for that year. As waste disposal practices in Canada change and as new information is made available, the  $L_0$  values will be adjusted accordingly.

$L_0$  was determined employing the methodology provided by the Revised 1996 IPCC Guidelines (IPCC/OECD/IEA 1997) (Equation 8–3) using the provincial waste composition data as input to the degradable organic carbon (DOC) calculation:

Equation 8–3:

$$L_0 = \text{MCF} \times \text{DOC} \times \text{DOC}_f \times F \times \frac{16}{12} \times \frac{1000 \text{ kg } \text{CH}_4}{\text{t } \text{CH}_4}$$

where:

$L_0$	=	$\text{CH}_4$ generation potential (kg $\text{CH}_4$ /t waste)
MCF	=	$\text{CH}_4$ methane correction factor (fraction)
DOC	=	degradable organic carbon (t C/t waste)
$\text{DOC}_f$	=	fraction DOC dissimilated
F	=	fraction of $\text{CH}_4$ in landfill gas
16/12	=	stoichiometric factor

Table 8–3 CH<sub>4</sub> Generation Potential (L<sub>0</sub>) from 1941 to Present

Province/Territory	2002 Organic Waste Diversion (%)	1941 to 1975		1976 to 1989		1990 to Present	
		DOC	L <sub>0</sub> (kg CH <sub>4</sub> /t waste)	DOC	L <sub>0</sub> (kg CH <sub>4</sub> /t waste)	DOC	L <sub>0</sub> (kg CH <sub>4</sub> /t waste)
Newfoundland	N/A	0.28	112.62	0.18	73.28	0.18	73.35
Prince Edward Island	N/A	0.27	108.74	0.17	67.19	0.16	64.63
Nova Scotia	29.7	0.25	100.89	0.16	62.35	0.16	64.1
New Brunswick	19.8	0.23	93.91	0.16	65.91	0.16	63.22
Quebec	13.7	0.36	144.45	0.21	82.52	0.2	81.23
Ontario	16.4	0.36	143.74	0.21	82.75	0.21	83
Manitoba	4.9	0.33	131.37	0.19	76.82	0.19	76.59
Saskatchewan	4.3	0.36	143.92	0.22	86.39	0.22	86.75
Alberta	16.7	0.39	157.63	0.26	104.46	0.18	71.87
British Columbia	23.3	0.28	111.86	0.17	69.89	0.16	63.71
Territories (Yk., N.W.T. and Nvt.)	N/A	0.22	87.59	0.15	58.54	0.16	65.13

Sources: All values are derived from data obtained from NRCAN (2006), Statistics Canada (2007a) and CRC Press (1973), with the exception of the 2002 Organic Waste Diversion figures, which were obtained from Thompson et al. (2006).

N/A = Unavailable categorical information.

According to the Revised 1996 IPCC Guidelines, the methane correction factor (MCF) for managed landfill sites has a value of 1.0 (IPCC/OECD/IEA 1997). The fraction (F) of CH<sub>4</sub> emitted from a landfill ranges from 0.4 to 0.6 and was assumed to be 0.5. From the IPCC *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (2000), a DOC<sub>F</sub> value of 0.6 was selected from a default range of 0.5 to 0.6. This DOC<sub>F</sub> value best reflects the lower concentration of lignin in the MSW waste, since the majority of wood wastes from pulp and paper industries and saw mills are disposed of in dedicated wood waste landfills.

The DOC calculation is derived from the biodegradable portion of the MSW (Equation 8–4):

Equation 8–4:

$$\text{DOC} = (0.4 \times A) + (0.17 \times B) + (0.15 \times C) + (0.3 \times D)$$

where:

A	=	fraction of MSW that is paper and textiles
B	=	fraction of MSW that is garden or park waste
C	=	fraction of MSW that is food waste
D	=	fraction of MSW that is wood or straw

### Wood Waste Landfills

Equation 8–3 generated an L<sub>0</sub> value of 80 kg CH<sub>4</sub>/t of wood waste, which was used to estimate emissions from wood waste landfills by the Scholl Canyon model. IPCC defaults were used for MCF in unmanaged deep landfills (MCF = 1); the fraction of CH<sub>4</sub> in the landfill gas (F = 0.5); and the fraction of DOC dissimilated (DOC<sub>F</sub> = 0.5), where the lower end of the default range for wastes containing lignin was selected (IPCC/OECD/IEA 1997). A composition of 100% wood waste was assumed in calculating the fraction of DOC in Equation 8–4.

### 8.2.2.2. Captured Landfill Gas

Some of the CH<sub>4</sub> that is generated in MSW landfills is captured as landfill gas and combusted, either by flaring or burning the gas for energy recovery. Combustion of the landfill gas converts CH<sub>4</sub> to CO<sub>2</sub>, thus reducing the CH<sub>4</sub> emissions. To calculate the net CH<sub>4</sub> emissions from landfills, the amount of CH<sub>4</sub> captured, as provided by the landfill facilities, is subtracted from the quantity of CH<sub>4</sub> generated, as estimated by the Scholl Canyon model. Added to this value, to account for the combustion inefficiency of the flares, is the quantity of captured CH<sub>4</sub> that passes through the flare uncombusted. The captured gas is wholly or partially flared or combusted for electricity or heat generation.

GHG emissions affiliated with the use of landfill gas for energy recovery are accounted for in the Energy Sector.

Flaring combustion efficiency for CH<sub>4</sub> in landfill gas of 99.7% was used to determine the quantity of CH<sub>4</sub> that circumvented the flare. This value was obtained from Table 2.4-3 of Chapter 2.4 of the U.S. EPA AP 42 (U.S. EPA 1995). The quantities of landfill gas collected from 1983 to 1996 were obtained from a personal communication.<sup>1</sup> Then, for 1997 to 2003, data on the amount of landfill gas captured were collected directly from individual landfill operators biennially by Environment Canada's National Office of Pollution Prevention (Environment Canada 1997, 1999b, 2001, 2003a). As of 2006, beginning with the 2005 data year, this survey is now being conducted by Environment Canada's Greenhouse Gas Division (Environment Canada 2007). Normally, the landfill gas capture data are collected every odd year; therefore, for the purposes of the national GHG inventory, the landfill gas capture data for the subsequent even years are averaged from adjacent odd years starting from 1997. However, since the 2008 survey, the Division has been collecting two years' data biennially, i.e., 2006–2007 data and 2008–2009 data from the 2008 and 2010 facility surveys, respectively (Environment Canada 2009, 2011).

### 8.2.3. Uncertainties and Time-Series Consistency

The following discussion on uncertainty for the categories within this sector is based upon the results as reported in an uncertainty quantification study by ICF Consulting (2004) of the NIR. This Tier 2 evaluation of uncertainty employed values from the 2001 inventory year (Environment Canada 2003b). However, there have been modifications made to the methodology, emission factors and sources of information as a consequence of the findings of this uncertainty study. Therefore, the results of this study may not be an accurate representation of the current uncertainty around the emissions from this subsector and the model inputs. However, in the absence of a follow-up Tier 2 study, it is expected that the improvements made would result in a reduction of the uncertainty for this subsector.

The CH<sub>4</sub> emissions from this key category include CH<sub>4</sub> emissions from MSW landfills and wood waste landfills.

The level of uncertainty associated with the CH<sub>4</sub> emissions from the combined subsectors was estimated to be in the range of –35% to +40%, which closely resembles the uncertainty range of –40% to +35% estimated in this study for the CH<sub>4</sub> emissions from MSW landfills. The level uncertainty range provided by the ICF Consulting study (2004) is only slightly larger than the  $\pm 30\%$  span estimated with a 90% confidence level by a previous study, which used a Tier 1 approach based upon 1990 data (McCann 1994). However, it should be noted that the uncertainty range of the ICF Consulting study (2004) is quoted for a 95% confidence interval, which would typically be larger than the range quoted for a 90% confidence interval.

The MSW landfills contributed to over 90% of the total CH<sub>4</sub> emissions from this key category in 2001 (Environment Canada 2003b). The uncertainty estimates for CH<sub>4</sub> emissions from MSW landfills seem to have been largely influenced by the uncertainty in the inventory values for L<sub>0</sub> for 1941–1989 and 1990–2001 and the CH<sub>4</sub> generation rate constant k, where the uncertainty for both k and L<sub>0</sub> were based upon an estimate from one expert elicitation. A simplified model of the Scholl Canyon method was used for the Monte Carlo simulation, which may have had a bearing on relevancy of the uncertainty values. An error was introduced in the calculation of the MSW landfill CH<sub>4</sub> emission uncertainty by the use of the year 2000 value (instead of the 2001 value) for the total CH<sub>4</sub> captured in Canada, resulting in an uncertainty range of +20% to +24% for these activity data. The actual uncertainty for this activity data entry should have been  $\pm 2\%$ .

Although the uncertainty range estimated in this study for wood waste landfills was significantly higher (i.e. –60% to +190%) than that for MSW landfills, its contribution to the uncertainty in the key category was much lower, owing to its relatively low contribution of emissions (i.e. less than 10%) (Environment Canada 2003b). The uncertainty estimate for wood waste landfills seems to have been largely influenced by the CH<sub>4</sub> generation rate, carbon content of the waste landfilled, and the biodegradable fraction of the waste, where the uncertainties were assumed by ICF Consulting (2004) based upon the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997) and/or the *IPCC Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000), where available.

The estimates are calculated in a consistent manner over time.

<sup>1</sup> Personal communication with ME Perkin of Environment Canada's National Office of Pollution Prevention in 1998.

### 8.2.4. QA/QC and Verification

A Tier 1 quality control review was conducted for this key category. An error in the distribution of the methane generation rate constants was identified within the application of the Scholl-Canyon model for the MSW landfills; however, this error does not significantly affect the resulting emissions. This error will be corrected for the subsequent Inventory.

### 8.2.5. Recalculations

Splitting the provincial specific  $k$  values from an overall average to distinct averages for the 1941–1975, 1976–1989 and 1990–2009 periods resulted in recalculations for the complete 1990–2009 time series for methane emissions from the MSW Landfill subsector. There was also a revision of the population data used to evaluate the MSW landfill emissions. Both of these changes resulted in differences from the 2010 Inventory for the Managed Waste Disposal on Land subsector ranging from -1.9% to 1.4%. Also added here is the revision of the 2008 recovered landfill gas collected quantity of 5%, without which the range would have been from 0.4% to 1.4%.

### 8.2.6. Planned Improvements

A study that was originally planned to be conducted in 2010 is now being considered for 2011. This study would review the quantity of wood waste being placed in Canadian wood and pulp and paper industry landfills and would verify the methodology, emission factors and historical activity data currently employed.

A review of current data is currently being undertaken for the evaluation of  $\text{CH}_4$  from unmanaged landfills. At present, the sole source of data for quantities of waste landfilled is the Statistics Canada biennial survey: *Waste Management Industry Survey: Business and Government Sectors*. The quantities of waste placed in unmanaged landfills make up a very small fraction of the total waste disposed of in Canada. Reporting of the estimates for this subsector will only be considered if, after the review, it is determined that there is no possibility of significant double counting (i.e. between wastes placed in managed versus unmanaged landfills).

## 8.3. Wastewater Handling (CRF Category 6.B)

### 8.3.1. Source Category Description

Emissions from municipal and industrial wastewater treatment were estimated. Both municipal and industrial wastewater can be aerobically or anaerobically treated. When wastewater is treated anaerobically,  $\text{CH}_4$  is produced; however, it is typical that systems with anaerobic digestion in Canada contain and combust the produced  $\text{CH}_4$ .  $\text{CH}_4$  emissions from aerobic systems are assumed to be negligible. Both types of treatment system generate  $\text{N}_2\text{O}$  through the nitrification and denitrification of sewage nitrogen (IPCC/OECD/IEA 1997).

$\text{CO}_2$  is also a product of aerobic and anaerobic wastewater treatment. However, as detailed in Section 8.1,  $\text{CO}_2$  emissions originating from the decomposition of organic matter are not included with the national total estimates, in accordance with the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997).

The emission estimation methodology for municipal wastewater handling is divided into two areas:  $\text{CH}_4$  from anaerobic wastewater treatment and  $\text{N}_2\text{O}$  from human sewage.

### 8.3.2. Methodological Issues

A more detailed discussion of the methodologies is presented in Annex 3.5.

#### 8.3.2.1. $\text{CH}_4$ Emissions

##### Municipal Wastewater Treatment

The IPCC default method was not used because the required data were not available. A method developed for Environment Canada (AECOM Canada 2010) was used to calculate an emission factor. This country-specific methodology provides for the accurate estimation of provincial methane emissions that best suits the available related activity data. Based on the amount of organic matter generated per person in Canada and the conversion of organic matter to  $\text{CH}_4$ , it was estimated that 1.97 kg  $\text{CH}_4$ /person per year could potentially be emitted from anaerobically treated wastewater. Additional information on the incorporated methodology is provided in Annex 3.5.

CH<sub>4</sub> emissions were calculated by multiplying the emission factor by the population of the respective province (Statistics Canada 2006, 2010b) and by the fraction of wastewater that is treated anaerobically.

### Industrial Wastewater Treatment

A survey was conducted by the Greenhouse Gas Division to obtain methane emissions from facilities that treated their effluent anaerobically on-site over the 1990–2009 time series. Where actual measured facility data were not provided, design specifications particular to that site were used to estimate maximum emissions expected. A complete description of the methodology is provided in Annex 3.5.

#### 8.3.2.2. N<sub>2</sub>O Emissions

##### Municipal Wastewater Treatment

The N<sub>2</sub>O emissions from municipal wastewater treatment facilities were calculated using the IPCC default method

Table 8–4 N<sub>2</sub>O Emission Factors

Year	Annual Per Capita Protein Consumption (kg protein/person per year)	N <sub>2</sub> O Emission Factor (kg N <sub>2</sub> O/person per year)
1990	23.82	0.060
1991	24.16	0.061
1992	24.29	0.061
1993	24.53	0.062
1994	24.77	0.062
1995 <sup>a</sup>	25.01	0.063
1996 <sup>a</sup>	25.04	0.063
1997 <sup>a</sup>	25.50	0.064
1998 <sup>a</sup>	25.75	0.065
1999 <sup>a</sup>	26.01	0.065
2000 <sup>a</sup>	26.26	0.066
2001 <sup>b</sup>	26.63	0.067
2002 <sup>b</sup>	26.57	0.067
2003 <sup>b</sup>	26.19	0.066
2004 <sup>b</sup>	26.35	0.066
2005 <sup>c</sup>	25.96	0.065
2006 <sup>c</sup>	25.93	0.065
2007 <sup>c</sup>	26.20	0.066
2008 <sup>c</sup>	25.64	0.064
2009 <sup>c</sup>	25.50	0.064

Sources: Statistics Canada (2007b), Statistics Canada (2008b) and Statistics Canada (2010a). The data have been adjusted for retail, household, cooking and plate loss

(IPCC/OECD/IEA 1997). This method estimates the N<sub>2</sub>O emission factor as the product of the annual per capita protein consumption, the assumed protein nitrogen content (16%), the quantity of N<sub>2</sub>O-N produced per unit of sewage nitrogen (0.01 kg N<sub>2</sub>O-N/kg sewage nitrogen) and the N<sub>2</sub>O/N<sub>2</sub>O-N conversion factor (1.57). Protein consumption estimates, in kg/person per year, were obtained from an annual food statistics report published by Statistics Canada (2007b, 2008b, 2010a). Data are provided for the years 1991, 1996 and 2001 to 2009. Protein consumption data for missing years are estimated by applying a multiple linear regression application to the Statistics Canada data. Emissions were calculated by multiplying the emission factor by the population of the respective provinces (Statistics Canada 2006, 2010b). A summary of the values for these two parameters over the time series is given in Table 8–4.

##### Industrial Wastewater Treatment

The *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997) do not address the methodology for the estimation of N<sub>2</sub>O emissions from industrial wastewater treatment. Owing to a lack of activity data, the N<sub>2</sub>O emissions from this category have not been evaluated.

#### 8.3.3. Uncertainties and Time-Series Consistency

##### Municipal Wastewater Treatment

The following discussion on uncertainty for the categories within this sector is based upon the results as reported in an uncertainty quantification study of the NIR (ICF Consulting 2004). This Tier 2 evaluation of uncertainty employed values from the 2001 inventory year (Environment Canada 2003b). However, there have been modifications made to the methodology, emission factors and sources of information as a consequence of the findings of this uncertainty study. Therefore, the results of this study may not be an accurate representation of the current uncertainty around the emissions from this subsector and the model inputs. However, in the absence of a follow-up Tier 2 study, it is expected that the improvements made would result in a reduction of the uncertainty for this subsector.

The overall level uncertainty associated with the wastewater treatment subsector was estimated to be in the range of –40% to +55%. The level uncertainty range provided by the ICF Consulting (2004) study is less than the ±60%



span estimated with a 90% confidence level by a previous study, which used a Tier 1 approach based on 1990 data (McCann 1994). This is an improvement to the uncertainty as assessed for this category, since the uncertainty range quoted by ICF Consulting (2004) for a 95% confidence interval should typically show a larger value than that quoted for a 90% confidence interval. Based on 2001 data, the trend uncertainty associated with the total GHG emissions (comprising CH<sub>4</sub> and N<sub>2</sub>O) from the wastewater treatment systems was estimated to be in the range of about +12% to +13%. The extrapolation of trend uncertainty in 2001 to the 2008 inventory should be made with caution, as trend uncertainty is more sensitive than level uncertainty to the changes in the inventory estimate values for the more recent years.

Since the methods and data sources have remained unchanged over the time series, the estimates for this category are consistent over time.

### Industrial Wastewater Treatment

The IPCC *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories* (IPCC 2000) provide for default uncertainties ranging from -25% to +25%. Since these data were for the most part obtained directly from the facility operators, based upon expert opinion, the uncertainty is estimated to be in the range of -15% to +15% or less.

### 8.3.4. QA/QC and Verification

No significant anomalies were identified.

### 8.3.5. Recalculations

Recalculations were made over the 1990–2008 time series due to the use of a new methane emission estimation methodology and emission factor for municipal wastewater treatment. This change resulted in percent differences ranging from 8% to 24%. N<sub>2</sub>O emissions from human sewage had changed slightly over the 2001–2008 time series (from -0.8% to 0.2%) due to the use of revised Statistics Canada protein consumption data from the Food Statistics publication (Statistics Canada 2010a).

### 8.3.6. Planned Improvements

Research is being considered to evaluate the possibility of reliably estimating N<sub>2</sub>O emissions from industrial wastewater treatment.

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## 8.4. Waste Incineration (CRF Category 6.C)

### 8.4.1. Source Category Description

Emissions from both MSW and sewage sludge incineration are included in the inventory. Some municipalities in Canada utilize incinerators to reduce the quantity of MSW sent to landfills and to reduce the amount of sewage sludge requiring land application.

GHG emissions from incinerators vary, depending on factors such as the amount of waste incinerated, the composition of the waste, the carbon content of the non-biomass waste and the facilities' operating conditions.

#### 8.4.1.1. MSW Incineration

A combustion chamber of a typical mass-burn MSW incinerator is composed of a grate system on which waste is burned and is either water-walled (if the energy is recovered) or refractory-lined (if it is not). GHGs that are emitted from MSW incinerators include CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O.

As per the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997), CO<sub>2</sub> emissions from biomass waste combustion are not included in the inventory totals. The only CO<sub>2</sub> emissions detailed in this section are from fossil fuel-based carbon waste, such as plastics and rubber.

CH<sub>4</sub> emissions from MSW incineration are assumed to be negligible and are not calculated owing to a lack of underlying emission research.

#### 8.4.1.2. Sewage Sludge Incineration

Two different types of sewage sludge incinerators are used in Canada: multiple hearth and fluidized bed. In both types of incinerators, the sewage sludge is partially de-watered prior to incineration. The de-watering is typically done in a centrifuge or using a filter press. Currently, municipalities in Ontario and Quebec operate sewage sludge incinerators. GHGs emitted from the incineration of sewage sludge include CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O, as in the case of MSW incinerators; however, since the carbon present in the wastewater sewage sludge is of biological origin, the CO<sub>2</sub> emissions are not accounted for in the inventory totals from this source.

## 8.4.2. Methodological Issues

The emission estimation methodology depends on waste type and gas emitted. A more detailed discussion of the methodologies is presented in Annex 3.5.

### 8.4.2.1. CO<sub>2</sub> Emissions

The *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC/OECD/IEA 1997) do not specify a method to calculate CO<sub>2</sub> emissions from the incineration of fossil fuel-based waste (such as plastics and rubber). Therefore, the following three-step method was developed for MSW incineration:

- *Calculating the amount of waste incinerated:* The amount of waste incinerated each year was estimated based on a regression analysis using data from an Environment Canada (1996) study, which contains detailed provincial incineration data for the year 1992, and from a study performed by A.J. Chandler & Associates Ltd. for Environment Canada, which provided incineration data for 1999, 2000 and 2001 (Environment Canada 2003c).
- *Developing emission factors:* Provincial CO<sub>2</sub> emission factors are founded on the assumption that the carbon contained in waste undergoes complete oxidation to CO<sub>2</sub>. The amount of fossil fuel-based carbon available in the waste incinerated has been determined using typical percent weight carbon content values (Tchobanoglous et al. 1993). The amount of carbon per tonne of waste is estimated and converted to tonnes of CO<sub>2</sub> per tonne of waste by multiplying by the ratio of the molecular mass of CO<sub>2</sub> to that of carbon.
- *Calculating CO<sub>2</sub> emissions:* Emissions were calculated on a provincial level by multiplying the amount of waste incinerated by the appropriate emission factor.

The CO<sub>2</sub> generated from the incineration of sewage sludge is not reported in the inventory emission totals, since the sludge consists solely of biogenic matter.

### 8.4.2.2. N<sub>2</sub>O and CH<sub>4</sub> Emissions

#### MSW Incineration

Emissions of N<sub>2</sub>O from MSW incineration were estimated using the IPCC default method (IPCC/OECD/IEA 1997). An average emission factor was calculated assuming that the IPCC five-stoker facility factors were most representative. To estimate emissions, the calculated emission factor was multiplied by the amount of waste incinerated by each province. CH<sub>4</sub> emissions from MSW incinerators are assumed to be negligible.

## Sewage Sludge Incineration

Emissions generated from the incineration of sewage sludge are dependent on the amount of dried solids incinerated. To calculate the CH<sub>4</sub> emissions, the amount of dried solids incinerated is multiplied by an appropriate emission factor. Estimates of the amount of dried solids in the sewage sludge incinerated in the years 1990–1992 are based on a study completed in 1994, as related in a personal communication with W. Fettes in February of 1994 from an interchange between Senes Consultants and Puitan Bennet. Data for the years 1993–1996 were acquired through telephone surveys of facilities that incinerate sewage sludge. Data for the years 1997 and 1998 were obtained from a Compass Environmental Inc. study prepared for Environment Canada (Environment Canada 1999a). Activity data for 1999, 2000 and 2001 were taken from a study conducted by A.J. Chandler and Associates Ltd. for Environment Canada (Environment Canada 2003c). To estimate the amount of sewage sludge incinerated in the years 2002–2009, a regression analysis was completed using the Chandler and Compass Environmental Inc. incineration values.

CH<sub>4</sub> emissions are estimated based on emission factors obtained from the U.S. EPA publication *Compilation of Air Pollutant Emission Factors* (U.S. EPA 1995). It is assumed that sewage sludge incineration is conducted with fluidized bed incinerators. Therefore, the emission factor is 1.6 t CH<sub>4</sub>/kt of total dried solids for fluidized bed sewage incinerators equipped with venture scrubbers. To estimate emissions, the emission factor was multiplied by the amount of waste incinerated by each province. The national emissions were then determined as the summation of these emissions for all provinces.

Emissions of N<sub>2</sub>O from sewage sludge incineration were estimated using the IPCC default emission factor for fluidized beds, 0.8 kg N<sub>2</sub>O/t of dried sewage sludge incinerated (IPCC 2000). To estimate emissions, the emission factor was multiplied by the amount of waste incinerated by each province. The national emissions were then determined as the summation of these emissions for all provinces.

## 8.4.3. Uncertainties and Time-Series Consistency

The following discussion on uncertainty for the categories within this subsector is based upon the results as reported in an uncertainty quantification study of the Canadian NIR

(ICF Consulting 2004). This Tier 2 evaluation of uncertainty employed values from the 2001 inventory year (Environment Canada 2003b). However, there have been modifications made to the methodology, emission factors, and sources of information as a consequence of the findings of this uncertainty study. Therefore, the results of this study may not be an accurate representation of the current uncertainty around the emissions from this subsector and the model inputs. However, in the absence of a follow-up Tier 2 study, it is expected that the improvements made would result in a reduction of the uncertainty for this subsector.

The overall level uncertainty associated with the waste incineration source category was estimated to be in the range of -12% to +65%. For 2001 inventory estimates, the overall trend uncertainty associated with the total GHG emissions (comprising CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O) from incineration of wastes (comprising MSW and sewage sludge) was estimated to be in the range of about +10% to +11%. The inventory trend uncertainty was estimated at +10%. The extrapolation of trend uncertainty in 2001 to the 2009 inventory should be made with caution, as the trend uncertainty is more sensitive than level uncertainty to the changes in the inventory estimate values for the more recent years.

#### 8.4.4. QA/QC and Verification

A Tier 1 quality assessment review was conducted for the key category: CO<sub>2</sub> emissions from MSW incineration as well as for the non-key category N<sub>2</sub>O and CH<sub>4</sub> emissions. No significant anomalies were identified.

#### 8.4.5. Recalculations

This year's recalculations resulted in no change for the 1990–2006 time series. For 2007 and 2008, there was a reduction of the emission values ranging from -0.07% to -0.005% for biogenic origin CO<sub>2</sub> emissions. For the same period the non-biogenic CO<sub>2</sub> and N<sub>2</sub>O emissions from municipal waste incineration decreased in ranges of -0.07% to -0.003% and -0.07 to -0.004%, respectively, from the previous submission due to a revision of the national population statistics.

#### 8.4.6. Planned Improvements

Facility-level incineration surveys have been conducted in 2008 and in 2010. The data from these surveys will be reviewed for completeness and accuracy before being

considered for incorporation into the Waste Sector model and into the Energy Sector methodologies (where energy recovery systems are in place).

A review of current information will be conducted to quantify the CH<sub>4</sub> emissions from the incineration of municipal solid wastes and report them in future submissions.

An uncertainty quantification study is being considered to revise uncertainty values for activity data, model parameters and emission factors used in the development of emission estimates for the Waste Sector. It is expected that these values will provide inputs to the Tier 1 and Tier 2 Sector and national emission uncertainty estimates.

# Chapter 9

## Recalculations and Improvements

This chapter summarizes the recalculations implemented in Canada's national greenhouse gas (GHG) inventory since its 2010 submission in order to facilitate an integrated view of changes in, and impacts on, emission levels and trends. Improvements to the 2011 submission due to methodological changes or refinements are presented in Section 9.1, while a description of planned improvements for future inventories can be found in Section 9.2.

### 9.1. Explanations and Justifications for Recalculations

The United Nations Framework Convention on Climate Change (UNFCCC) requires all Annex I Parties to continually improve their national greenhouse gas inventories. As new information and data become available and more accurate methods are developed, previous estimates are updated to provide a consistent and comparable trend in emissions and removals. On a continuous basis, Environment Canada consults and works jointly with key federal and provincial partners along with industry stakeholders, research centres and consultants to improve the quality of the under-

lying variables and scientific information for use in the compilation of the national inventory. Where necessary, Environment Canada revises and recalculates the emission and removal estimates for all years in the inventory.

Table 9-1 provides a summary of the recalculations that occurred due to methodological changes or refinements since the previous submission, with a brief description, justification and summary of individual impacts on emissions and trends. In addition to the changes listed in Table 9-1, further recalculations may have occurred due to updates in activity data, reallocations of emissions, the correction of errors discovered since the previous submission, or minor incremental enhancements. Further details on sectoral recalculations may also be found within the individual chapters for each sector. Estimated impacts on levels and trends at a national level are presented in Sections 9.1.1 and 9.1.2.

It should be noted that in response to the findings of the UNFCCC Expert Review Team's (ERT) review of Canada's 2010 submission, the following revised estimates of GHG emissions were required: CH<sub>4</sub> emissions from Common Reporting Format (CRF) category 1.B.2.b.iv Natural Gas Distribution from the Energy Sector; CH<sub>4</sub> and N<sub>2</sub>O emissions from CRF category 4.F Field Burning of Agricultural Residues from the Agriculture Sector; and CH<sub>4</sub> and N<sub>2</sub>O emissions from CRF category 6.B.1 Industrial Wastewater Handling from the Waste Sector. These recalculations were reflected in a complete re-submission of Canada's 2010 CRF tables on October 18, 2010. These changes were not significant and related to only a few source categories. The recalculations discussed in Table 9-1 reflect the comparisons made between Canada's original 2010 inventory submission and the 2011 submission.

Table 9-1 Summary of Recalculations Due to Methodological Change or Refinement

CRF#	Category	Description	Justification	Impact on Emissions
1.A. (including 1.A.1 and also impacts 1.A.2, and 1.A.4)	Stationary Combustion	Revision of coal emission factors	The current coal CO <sub>2</sub> emission factors were reviewed and updated based on detailed data analysis and support from the Geological Survey of Canada and some electric utilities. The new emission factors will replace the previous "time series" adjusted emission factors with a single value, with appropriate uncertainty bounds.	New emission factors have resulted in a change to the overall emissions. The impact amounts to an increase of about 0.8 Mt in 2008 and a drop of about 0.3 Mt in 1990

Table 9–1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
1.A	Energy – Electric Power and Industrial Processes – Mining, Pulp & Paper, Iron & Steel, Non-Ferrous, Cement, RPP Manufacturing, Chemicals, Other Manufacturing, Forestry, Construction	Re-allocation of Electricity by Industry and Steam Generation lines of the Report on Energy Supply-Demand in Canada (RES-D) to the specific industry generating the electricity or steam	The 1996 IPCC Guidelines for National Greenhouse Gas Inventories require that emissions from facilities that generate electricity/heat wholly or partly for their own use, as an activity that supports their primary activity, should be assigned to the sector where they were generated. Previous NIRs did not follow this procedure; however, a new re-allocation brings the NIR in line with the 2006 guidelines.	Overall, the national and provincial GHG emissions will remain the same. Specific industry lines will increase based on the Electricity and Steam Generation attributed to the industry. The emissions reported under Utility Electricity and Steam Generation will be reduced (as some of the Steam Generation comes from Utilities, it is not fully eliminated).
1.A.2	1.A.2. - Pulp & Paper, and Memo Items - CO <sub>2</sub> from biomass combustion	Revised wood waste and spent pulping liquor emission factors for industrial facilities	"All emission factors reviewed were converted to a mass basis at 50% m.c. (moisture content). 50% m.c. was considered appropriate based on a study done for Environment Canada by AMEC in 2007, and supported by the US EPA AP-42, Chapter 1.6 (2003). The relationship between higher heating value and moisture content was developed based on data collected from the Phyllis Database (Energy Research Centre of the Netherlands) on Canadian tree species.	The effect of applying the new emission factor is primarily a decrease in biomass CO <sub>2</sub> (a memo item).
1.A.2.a	Energy - Iron and Steel - Gaseous Fuels	Revision of coke oven gas emission factor to a more representative value	The emission factor (EF) for coke oven gas (COG) has been changed from a country-specific factor of 1600 g/m <sup>3</sup> to the <i>Revised 1996 IPCC Guidelines'</i> COG EF of 878.9 g/m <sup>3</sup> to more closely align with industry.	Small overall decrease in CO <sub>2</sub> emissions from 1990 onwards (from approximately 55 kt to 66 kt per year).
1.A.3.a, 1.C.1a, 1.A.5.b	Civil Aviation, International Aviation Bunkers and Other Mobile (Domestic Military)	Development of a model for the aviation sector based on origin-destination data and aircraft-specific fuel consumption parameters. Updated CH <sub>4</sub> and N <sub>2</sub> O emission factors were also incorporated.	The Tier 3 method is a bottom-up, flight-by-flight estimation tool that improves the Aviation Turbo Fuel split between civil (domestic) aviation and international bunker aviation. The new method also allows for domestic military emissions to be broken out from civil aviation for reporting purposes. The correct domestic/international split (as defined by the IPCC) is important since Canada is held accountable for its domestic aviation emissions. Additionally, more detailed analysis can be performed because of the flight-by-flight nature of the model.	The result is increased emissions allocated to the domestic component in the earlier years of the time series and a shift to decreased allocation in the later years.
1.A.3.b and 1.A.3.e	Energy - Road Transportation & Other Transportation	Revised kilometre accumulation rates (KARs), averaged by vehicle age and vehicle class, was integrated into the Mobile Greenhouse Gas Emission Model (MGEM).	MGEM requires, as one of its inputs, the annualized distances travelled by each type of vehicle. These KARs have been updated following the acquisition of enhanced data that more accurately represent Canadian on-road vehicle activity.	Minor adjustments resulted for all inventory years.



Table 9–1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
1.A.3.b and 1.A.3.e	Energy - Road Transportation & Other Transportation	Revised light-duty vehicle fuel consumption ratios (FCRs) were incorporated into MGEM.	Natural Resources Canada (NRCan) issued revised FCRs for model years 1966–2009.	Minor adjustments resulted for all inventory years.
1.A.3	Transport	Revised methodology for ethanol fuel	In previous inventories, Statistics Canada has been ambiguous regarding the inclusion of biofuels in their annually reported National Energy Balance (RESO – 57-003). In order to ensure NIR completeness, known quantities of biofuel consumption were modelled for emissions over and above the amounts of base-fuels reported in the RESO. However, in the current inventory, Statistics Canada has stated that ethanol has in fact always been included in the RESO. As a result of this new information, known quantities of ethanol are first removed from the quantity of base-fuel (gasoline) reported in the RESO prior to emissions modelling.	All historic estimates of Transport emissions from gasoline consumption reflect this reduction (1990–2008).
1.A.3.b	Energy - Road Transportation	Revised 12 on-road emission factors for CH <sub>4</sub> and N <sub>2</sub> O	Emission factors have been updated using Canadian vehicle emissions testing data (i.e. tested with high-sulfur gasoline) as opposed to previously applied emission factors sourced from U.S. testing/modelling data (and low-sulphur gasoline).	There was no change in 1990 emissions; however, there was a slight increase in emissions in the later years.
1.B.2	Fugitive Emissions from Petroleum Refining	Implementation of consistent methodology to generate emission estimates across all petroleum refining emissions subcategories: Fugitive, Flaring and Venting	The existing model used inconsistent methodologies to estimate emissions for the reported subcategories of Flaring, Fugitive and Venting. Modifying the model to use production numbers results in a higher quality estimate of refinery fugitive GHG emissions. There was an error found in the original model which was corrected. In addition, emission factors used in the Fugitive category were not consistently applied across all years.	Impact on fugitive emissions varies from an increase of approximately 125 kt in 2008 to a decrease of approximately 600 kt in the period 1991 to 1994.
1.B.2.b.iv.	Fugitive Emissions - Natural Gas - Distribution	Revision of the calculated natural gas distribution emission factor based on a new study performed by Radian International for the Canadian Gas Association	A study was conducted to update the natural gas distribution emissions originating from station vents. The emission factor has been revised and now more closely aligns with industry.	Emissions from natural gas distribution will decrease by 1482 kt CO <sub>2</sub> eq (1990) to 1855 kt CO <sub>2</sub> eq (2008), which is a difference of over 50%.
1.B.2.c.i.2	Fugitive Emissions - Oil and Natural Gas - Venting - Natural Gas	Update to historical CO <sub>2</sub> emissions from natural gas process venting based on natural gas "process shrinkage" instead of "gross new	Refinement of the fugitive extrapolation model for natural gas wells was implemented in the 2009 submission for the years	The analysis shows that formation CO <sub>2</sub> emissions will change slightly. Only the provinces of Alberta, Saskatchewan and British

Table 9-1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
1.B.2.c.i.2 (cont'd)	Fugitive Emissions - Oil and Natural Gas - Venting - Natural Gas	production" for the years 1990–1999	2001 onwards. However, the backcast of emissions from 2000 backwards was not. This change maintains a consistent methodology for the entire time series. The current method for 1990–1999 uses the same approach as that for 2001 onwards. The formation CO <sub>2</sub> emissions for 2000 were backcasted to 1990 using the ratio of gross new production. The new method will use natural gas shrinkage to backcast the emissions.	Columbia, as well as the Yukon are affected by this change, as they are the only provinces/territories that have formation CO <sub>2</sub> emissions in the Upstream Oil and Gas (UOG) fugitive model. Change in emis- sions ranges from a 903-kt CO <sub>2</sub> increase to a 13-kt CO <sub>2</sub> decrease.
(2.A.3) and (2.A.4)	Limestone / Dolomite and Soda Ash Use	Provincial reallocation of emissions based on gross output	This is a change to the way in which emissions get allocated to the provinces and territories (where the production activities take place). For the 1990–2009 NIR, to allocate emis- sions from limestone and dolomite use by province/ territory, it is proposed to apply the gross output (GO) ratio (i.e. GO of a province over GO of Canada) instead of the gross domestic product (GDP) ratio (i.e. GDP of a province over GDP of Canada).	No change in total emis- sions
2.B.5, 2.B.8.a, 2.B.8.b, 2.B.8.f, 2.B.10	Chemical Industry - Other CH <sub>4</sub>	Inclusion of previously not included CH <sub>4</sub> estimates for styrene, EDC, ethylene, etc., for the chemical industry	The ERT recommended that Canada estimate CH <sub>4</sub> emis- sions for specific sources reported as "not estimated" (NE) in the 2009 submission such as carbide production, methanol, ethylene, carbon black, dichloroethylene (EDC) and styrene.	A new category called "Pet- rochemical Production" has been added to the inven- tory. Total impact (increase) around 33 kt CO <sub>2</sub> eq.
2.C.1	Industrial Processes - Iron and Steel Production	Changes for this sector include inclusion of new Canada-specific emission factors for metallurgical coke used as reductant, the use of Canada-specific carbon content values for pig iron and crude steel (from the Canadian Steel Producers Association) and inclusion of previously omitted emissions from the scrap steel used in the steelmaking process and the consumption of electrodes in secondary ladle metal- lurgy equipment.	The proposed changes will increase estimate accuracy by using country-specific emission factors and carbon content values instead of IPCC default values, will ac- count for previously missing estimates and will address the discrepancies seen be- tween the data reported by industry to the Greenhouse Gas Reporting Program and the estimates from in the GHG inventory.	Significant increase in pro- cess emissions (upwards of 43–47%) for each year in the time series
2.F	Industrial Processes - Consumption of SF <sub>6</sub> in electrical equip- ment	Update to the emission methodology for SF <sub>6</sub> from electricity systems	To address the issue of overestimation and provide estimates at the provincial level, the current Tier 1 SF <sub>6</sub> estimates have been replaced with the Tier 3 estimates provided by the Canadian Electricity Assoc-	This change will result in a reduction of approximately 86% for each year in the time series.

Table 9–1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
2.F (cont'd)	Industrial Processes - Consumption of SF <sub>6</sub> in electrical equipment		iation and Hydro-Quebec for the years where data are available (i.e. 2006 and after).	
3.D.1	Solvent, Use of N <sub>2</sub> O as anesthetic	Revision to emission factor for N <sub>2</sub> O used as anesthetic	The original reference (U.S. GHG inventory) had updated its emission factor to use the value found in the 2006 IPCC Guidelines.	Slight increase in emissions (~3%)
4.A /4.B	Agriculture - Enteric Fermentation / Manure Management	Inclusion of emission factor correction factor for slaughter heifers and steers	A factor was missing from the calculations; this factor corrects the emission factor for length of time that animals designated for Canadian feedlots live and emit GHGs. The proposed change reintegrated this missing factor into the calculations in order to make the inventory calculation model consistent with the Boadi et al. (2004) documentation.	This issue has a direct impact on emissions of CH <sub>4</sub> from enteric fermentation, CH <sub>4</sub> emissions from manure management and N <sub>2</sub> O emissions from manure management. The changes reduced emission estimates from non-dairy cattle by -1.2 Mt CO <sub>2</sub> eq for 1990 and -1.9 Mt CO <sub>2</sub> eq for 2008. This change also led to recalculations and reductions in N <sub>2</sub> O emissions for several source categories: 4.B.2 (-0.43 Mt CO <sub>2</sub> eq for 1990 and -0.62 Mt CO <sub>2</sub> eq for 2008); 4.D.1 (-0.09 Mt CO <sub>2</sub> eq for 1990 and -0.16 Mt CO <sub>2</sub> eq for 2008); 4.D.2 (-0.42 Mt CO <sub>2</sub> eq for 1990 and -0.62 Mt CO <sub>2</sub> eq for 2008); and 4.D.3 (-0.33 Mt CO <sub>2</sub> eq for 1990 and -0.43 Mt CO <sub>2</sub> eq for 2008). This issue was the dominant change that resulted in the reduction of agricultural emissions estimates by -1.6 Mt CO <sub>2</sub> eq for 1990 and -4.1 Mt CO <sub>2</sub> eq for 2008; however, other revisions to the animal livestock database (See Chapter 6), contributed to this change and the agricultural emissions trend.
4.F -	Agriculture - Field Burning of Agriculture Residues	Inclusion of previously omitted CH <sub>4</sub> and N <sub>2</sub> O emissions	This change addresses the omission of emissions of methane and nitrous oxide from field burning of agriculture residues. (Note: since the amount of crop residue nitrogen subject to field burning no longer goes through decomposition, this nitrogen should be deducted from crop residue decomposition. As a result, emissions/removals from summerfallow, conservation tillage, irrigation as well as leaching are affected.)	In the CRF Reporter, this is not shown as recalculations because emissions of CH <sub>4</sub> and N <sub>2</sub> O from the field burning of crop residue were reported in the re-submission of the 2010 NIR.

Table 9-1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
5.A	LULUCF - Forest Land	Modification of algorithm used to simulate the growth of young forest stands	The algorithm used to simulate the growth of very young stands was further refined to reduce discontinuities as these stands grew to merchantable volume and were assigned standard growth curves.	The increase in C uptake by biomass (by on average 0.0001 kt CO <sub>2</sub> /ha/year or 1%, for the 1990–2007 time period) was largely due to this modification.
5.A	LULUCF - Forest Land	Recalibration of C transfer parameters for fire	The process of recalibrating the C transfer parameters in fire disturbance matrices incorporates new fire and weather into the averaging process used to derive these parameters.	Downward recalculations of immediate CO <sub>2</sub> emissions (which ranged from less than 0.002 to 0.025 kt CO <sub>2</sub> /ha/year burned, for the years 1990–2007) were largely the result of the recalibration of C transfer parameters in fire disturbance matrices. Lower rates of biomass consumption during a fire leave more dead biomass remaining on site after the fire and result in higher subsequent emissions from decay (by on average 0.0001 kt CO <sub>2</sub> /ha/year or 2%).
5.B.2	LULUCF - Forest Land converted to Cropland (FLCL)	Introduction of a "holding period" for Forest Land converted to Cropland (FLCL) for Western Canada	The current method assumes that the FLCL goes into annual or perennial cropping systems and violates the assumption of no change in soil organic carbon over the long term for FLCL on pasture and forage crops for Western Canada. Therefore it is proposed that no change management practices be allowed on FLCL for 20 years.	This correction resulted in a net increase of FLCL emissions by 16 kt in 1990 and 46 kt in 2008.
5.B.2, 5.D.2, 5.E.2	LULUCF - Forest Land Conversion LCL, LWL, LSL	Revised deforestation activity data for Forest Land Conversion	Estimates for the post-2000 period were revised to reflect new sampling and mapping. Additionally, previous estimates for 1970–2000 time periods were revised based on additional sampling and quality control activity. These estimates are developed by the Canadian Forest Service (CFS) for the Monitoring, Accounting and Reporting System (MARS).	Due to their cross-cutting nature, the recalculations of forest area converted to other land use affected estimates for several LULUCF sub-categories (FLCL, FLWL and FLSL). The impacts of these changes are not readily discernable unless spatially disaggregated or considered in each end-use class. For example, there is a reduction of FLCL emissions by 1300 kt in 1990 and 1600 kt in 2008.
5.D.2.	Wetlands -Reservoirs - Emissions from Reservoirs	A single temporal emission curve replaced the two regional curves used to determine CO <sub>2</sub> emissions from the surface of reservoirs.	This change is based on a statistical assessment revealing that the curve parameters for the Taiga and Boreal and Montane Cordillera regions were not significantly different from	Changes in EFs resulted in small increases in emissions from the surface of reservoirs, where increases in emissions never exceed 23 kt CO <sub>2</sub> /ha/year or a 1% change.

Table 9–1 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Justification	Impact on Emissions
5.D.2. (cont'd)	Wetlands -Reservoirs - Emissions from Reservoirs		each other. Combining the datasets of the two regions improves EFs for the overall time series on which the model is based by filling gaps and increasing sample size.	
6.A.1	Waste - Solid Waste Disposal on Land	A change was made to the methane generation rate constant, where three average provincial k values were derived to now coincide with the three periods used for the methane generation potential: 1941–1975, 1976–1989 and 1990–2009.	The change in this Scholl-Canyon input is expected to provide more refined methane emission estimates over each of these three time periods.	Splitting the provincial specific k values from an overall average to distinct averages for the 1941–1975, 1976–1989 and 1990–2009 resulted in recalculations for the complete 1990–2009 time series for methane emissions from the Municipal Solid Waste (MSW) Land-fill subsector. This resulted in a -1.9% to 1.4% change from the 2010 Inventory for the Managed Waste Disposal on Land subsector.
6.B.1	Waste - Industrial Wastewater	Inclusion of previously not estimated emissions of CH <sub>4</sub> from industrial wastewater	In past submissions the emissions from this subcategory, although identified, were not estimated. In order to correct this incompleteness, data were collected through an in-house survey of industrial facilities that were suspected or known to employ anaerobic units to treat their effluent on-site.	The magnitude of the CH <sub>4</sub> emission recalculations from this subsource in the CRF Reporter are shown to be relatively smaller than that which would be expected from a newly reported source. This is due to these emissions being previously reported in the re-submission of the 2010 NIR.
6.B.2.a	Waste - Domestic and Commercial Wastewater	Revisions to the Commercial and Domestic Wastewater (Bo) and the methane conversion factor (MCF).	The change involved a modification of the ORTECH method to more closely match the IPCC GPG method, a new maximum methane producing capacity (Bo), and revision of the methane conversion factor.	Recalculations were made over the 1990–2008 time series due to the use of a new methane emission estimation methodology and emission factor for municipal wastewater treatment. This change resulted in percent differences ranging from 8% to 24%.

### 9.1.1. Implications for Emission Levels

Overall GHG emissions (excluding the LULUCF Sector) were revised by relatively small amounts for all years. Recalculations had the largest implications for 1990, 2007 and 2008 (re-submission), which showed decreases of 0.23%, 0.28% and 0.39%, respectively. See Table 9–2 for additional details.

Large recalculations occurred in the LULUCF sector (Table 9–2) as a result of the re-calibration of model parameters.

Chapter 7 provides more information on these and other improvements.

### 9.1.2. Implications for Emission Trends

Overall, the recalculations of the total GHG estimates (excluding the LULUCF Sector) also had a downward effect on the trend between 1990 and 2008 (re-submission), now reported as a 23.9% increase in total GHG emissions since 1990 instead of the previously reported 24.1% increase.



Table 9–2 Summary of Recalculations Due to Methodological Change or Refinement

	GHG Emissions per Year							
	1990	2000	2004	2005	2006	2007	2008	2009
<b>National Total</b>								
Current (Mt CO <sub>2</sub> eq)	590	716	742	731	719	748	732	690
2010 submission (Mt CO <sub>2</sub> eq)	592	717	741	731	718	750	734	-
2010 resubmission (Mt CO <sub>2</sub> eq)	592	717	741	731	718	750	735	-
Change (%)	-0.25%	-0.15%	0.22%	0.06%	0.13%	-0.29%	-0.39%	-
<b>LULUCF</b>								
Current (Mt CO <sub>2</sub> eq)	-67	-62	97	54	65	51	-17	-12
2010 submission (Mt CO <sub>2</sub> eq)	-52	-80	117	41	41	45	-13	-
Change (%)	30.9%	-22.7%	-17.4%	29.3%	56.2%	13.1%	32.1%	-

## 9.2. Planned Improvements

Canada has identified planned improvements in that, when implemented, will impact the inventory time series from 1990 onwards. Further details on sectoral improvement plans may also be found within the individual chapters for each sector. Annual improvements to the national inventory must be applied to all estimation years in order to maintain time series consistency.

The planned improvement activities are based on recommendations from both internal sources and external review processes such as the UNFCCC expert reviews, and

collaborative work between inventory sector experts and industry, other government departments, and academia. As part of Canada's National System, improvement activities and work plans are developed on a continuous basis to further refine and increase the transparency, completeness, accuracy, consistency and comparability of the Canadian GHG inventory. Improvement activities are developed by sector experts and prioritized by a prioritization and planning committee (P&PC) using key category contributions, quality assurance/quality control (QA/QC) activities, uncertainty assessments, resource availability and potential impacts, as primary considerations.

Table 9–3 Principal Planned Improvements

CRF#	Category	Description	Basis of Planned Improvement
1.A.3.B	Energy - Road Transportation	The development of better on-road activity data	Based on a UNFCCC Expert Review Team recommendation
1.A.3.d	Marine Transport	Development of an improved model to estimate marine emissions	Based on a UNFCCC Expert Review Team recommendation
2.B.1	Industrial Processes – Ammonia Production	Determination of the amounts of natural gas used as feedstock and as fuel, and update of the EF	Based on QA/QC activities
2.F.	Industrial Processes - Consumption of HFCs	Development and use of country-specific EFs	Continuous improvement
2.G	Industrial Processes – Other & Undifferentiated Production	Allocation of CO <sub>2</sub> emissions from petrochemical industry to the category Petrochemical Production	Continuous improvement
4.A/4B	Enteric Fermentation/ Manure Management	Integration of time series of nutrition data for certain animal categories	Improve capacity to capture change in farm practices in emission estimates
4.A/4B	Enteric Fermentation/ Manure Management	Review of activity data used in the calculation of emissions from slaughter heifers and steers	Good practice: Tier 2 review of activity data sources

Table 9-3 Principal Planned Improvements (cont'd)

CRF#	Category	Description	Basis of Planned Improvement
4.B	Manure Management	Improvement to the description of regional manure management systems	Based on a UNFCCC Expert Review Team recommendation
4A/4B/4D	Enteric Fermentation, Manure Management and Agricultural Soils	Tier 2 uncertainty analyses will be updated, based on improved methods and integrated data management.	Based on a UNFCCC Expert Review Team recommendation
5. D	Wetlands	Improve estimates from peatland conversion, drainage and extraction, and restoration.	Improve estimates of the impacts of peatland extraction
5. D	Wetlands	Analyze geospatial information related to peatlands under extraction to improve the preconversion above-ground biomass estimate (biomass loss).	Decrease uncertainty of land-use change impact
5. D	Wetlands	Scoping of horticultural peatland emission methodology (2006 GLs)	To comply with future reporting obligations
5. D/E	Wetlands & Settlements	Completion of uncertainty assessments	Based on a UNFCCC Expert Review Team recommendation
6.A.2	Waste - Solid Waste Disposal on Land - Unmanaged Waste Disposal Sites	Currently conducting a study to provide estimates for waste quantities landfilled in unmanaged landfills in Canada over the 1941–2009 time series. These data will be incorporated in the subsequent inventory if the data collected are reliable, the methodology is transparent and can be applied consistently over the time series, and the methodology does not reduce the present accuracy of the emissions from this subsector.	Highlighted in the Annual Review Report by the UNFCCC Expert Review Team on the 2009 NIR submission as a subsector emission not currently estimated
6.A.3	Waste - Solid Waste Disposal on Land - Wood Waste Landfills	This study would review the quantity of wood waste being placed in Canadian wood and pulp and paper industry landfills and would verify the methodology, emission factors and historical activity data currently employed.	Continuous improvement initiative

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